

## CONSERVATION

# Needed: Better metrics, rigorously tested

Predictive models of biodiversity change are required to inform conservation policy decisions

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Biodiversity is measured in many different ways, because no single measure adequately captures nature's many forms and functions. Over the past decade, numerous metrics for biodiversity—including species abundance, extinction risk, distribution, genetic variability, species turnover, and trait diversity—have been used to create indicators to track how biodiversity has changed (1–3). These indicators have made it clear that biodiversity loss, however it is measured, is showing little sign of abatement (1,4) and that humans must respond to safeguard the provision of natural services on which we all rely (5,6). But which metrics provide the most informative indicators under which circumstances? And how can the growing list of indicators of biodiversity change best serve conservation policy decisions?

## Alignment of target and indicator

If we are interested in the outcome of a prospering economy, we measure its performance over time using metrics such as cost of goods, income, and employment numbers. Those metrics are then used to produce indicators such as GDP and RPI, which indicate how the economy is performing. Similarly, metrics like species abundance are used to create indicators of the health of biodiversity.

For an indicator to help achieve a particular conservation target, target and indicator need to be closely aligned (4). There is little point in measuring progress toward a target with an indicator based on a metric that is only loosely related to the desired outcome. For example, ensuring that protected areas maintain their biodiversity is a fundamental goal of conservation. Targets are frequently centered on the extent of area under protection in a

given country or region (e.g. 17% of land should be under protection by 2020; 6). Here, the implicit assumption is that the greater the area protected, the more biodiversity will prosper. However, this ignores factors that influence the effectiveness of the protected area: governance, funding,

on Biological Diversity (CBD) (6) tend to be less specific. As a result, alignment between metric, indicator, and target can be poor. For example, one CBD target calls for pollution to be reduced to levels that are not detrimental to ecosystem function or biodiversity. This laudable target does not detail important features: Which pollutants, ecosystem functions, or particular aspects of biodiversity should be addressed? The distinction is important because many functions will trade off with one another, and prioritizing some aspects of biodiversity will be at the cost of others. Efforts to measure progress toward this target, hold polluting countries and industries to account, and diagnose which interventions work best are made more difficult.

The outcomes of global biodiversity targets are perhaps inevitably less focused than those in specified circumstances such as fisheries management. However, with greater demand and scrutiny placed on biodiversity indicators (4) through targets such as CBD, how can they better support conservation efforts? One way forward is improved prediction.

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## Projecting forward

Effective conservation policy decisions require an explicit understanding of the links between desired outcomes of conservation, how those outcomes can be measured, and the proposed actions needed to achieve them (10). One way to accomplish this is to project forward the impacts of a prospective policy. In doing so, both the impact of the policy and the ability of indicators to detect change in biodiversity can be measured (see the first figure). By assessing alternative policies against a suite of metrics, the best combination of metrics and indicators for evaluating policy impacts can be identified.

In a recent study, Kelly *et al.* showed

the type of species within them. Protected areas differ greatly in the protection they afford species, but management effectiveness indicators are currently only available for a fraction (<5%) of protected areas (4). Using just one metric as an indicator may not achieve the desired outcome.

Ideally, the chain between metric, indicator, and policy should start with specific targets. In fisheries management, targets are often very explicit, typically relating to the sustainability of fish stocks; this helps to guide fisheries policy, management interventions, and detecting fishing impacts on marine biodiversity (7). Targets can vary widely in scale; metrics such as change in total biomass, catch, and mean trophic level are used to evaluate sustainability targets for whole ecosystem management (8), whereas changes in metrics such as recruitment and abundance are used to construct indicators under alternative scenarios for management of specific fish stocks (9).

In contrast, global biodiversity targets such as those agreed to in the Convention

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1 how indicators can provide a link between  
2 broad conservation targets and local scale  
3 implementation (11). The authors applied  
4 abundance metrics to decisions on wild-  
5 fire management in Australia. The results  
6 showed that optimizing fire management  
7 using an indicator based on geometric  
8 mean abundance of the community re-  
9 sulted in improved biodiversity outcomes  
10 compared with the conventional wisdom  
11 of maximizing the prevalence of different  
12 fire management regimes. This approach  
13 demonstrates that clearly defined man-  
14 agement goals are necessary to maximize  
15 biodiversity in fire-prone ecosystems.

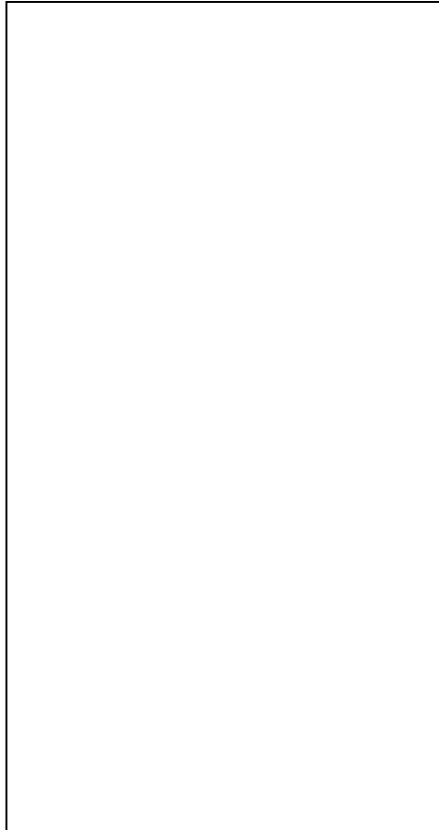
16 In fisheries science, substantial empha-  
17 sis has been placed on ensuring that indi-  
18 cators respond in predictable ways to par-  
19 ticular interventions or pressures,  
20 enabling decision makers to tease apart  
21 the impacts of different drivers of change  
22 (12). In-depth knowledge of such indicator  
23 behavior is currently lacking from anal-  
24 yses of most global biodiversity indicators.  
25 Put simply, any useful indicator must be  
26 able to pick up the impact of a manage-  
27 ment intervention.

### 28 **Rigorous testing**

29 Two aspects that affect indicator perfor-  
30 mance are the design of the indicator and  
31 the quality of the data that underpin it. An  
32 indicator may perform badly because data  
33 available to calculate it are difficult to ob-  
34 tain or are biased geographically or taxo-  
35 nomically, rendering it unrepresentative  
36 of the components of biodiversity that the  
37 indicator is designed to reflect (10). In oth-  
38 er cases, excellent data may be available  
39 but the metric may be a poor proxy for the  
40 aspects of biodiversity of interest. For ex-  
41 ample, the rate of forest loss is often cited  
42 as a chief driver of decline of wildlife  
43 populations (5,6), but the impact of bush-  
44 meat hunting, also recognized to be a key  
45 driver, is rarely measured. Thus, biodiver-  
46 sity declines in forested areas can continue  
47 undetected even when the rate of forest  
48 loss is slowed or halted.

49 Different metrics can give varying im-  
50 pressions of conservation success. For ex-  
51 ample, if a wildlife population collapses  
52 leaving a much smaller population sur-  
53 viving in only one region, tracking abun-  
54 dance yields a picture of decline whereas  
55 tracking extinction risk suggests recovery  
56 (see the second figure). One problem lies  
57 in adapting metrics designed for another  
58 purpose: Extinction risk assessments pro-  
59 vide an instantaneous snapshot, and were  
not designed to evaluate change through

time (10). Another problem is expectation:  
The risk to the species has diminished to-  
wards the end of the example because the  
population is now stable, albeit at a much  
lower level than before.



Several fisheries indicators have been  
subjected to rigorous evaluation of wheth-  
er or not they reliably predict changes in  
marine ecosystems (7, 8), but few biodi-  
versity indicators have been tested in this  
way. There are exceptions; tests of indica-  
tor performance have shown that data bi-  
ases can give a misleading impression of  
policy impacts (7,10). Other recent studies  
favorably related the Living Planet Index's  
underlying metric (geometric mean abun-  
dance) to models of species viability from  
ecological theory (13), and explored its  
mathematical properties (14),  
demonstrating it is fit for purpose to  
measure trends in species extinction risk.

More extensive stress testing of biodi-  
versity indicators would enhance  
knowledge of how biodiversity is chang-  
ing, show whether the existing indicators  
can measure that change, and help identi-  
fy the most appropriate policies to coun-  
teract biodiversity declines. Predictive  
modelling will help ensure that biodiver-  
sity indicators are capable of supporting

conservation policy decisions. For exam-  
ple, sampling model systems, mimicking  
the way data are collected in biodiversity  
monitoring programmes, can be used to  
calculate indicators, and to provide a  
completeness that is not available in real-  
world monitoring data. This framework—  
referred to as management strategy evalu-  
ation—has been used to test indicators for  
fisheries management (7, 8) and has also  
been applied to the evaluation of other bi-  
odiversity indicators (10), showing that  
while some indicators perform well,  
others need rethinking.

### **Toward a meaningful set of metrics**

If the right information to guide conserva-  
tion policy cannot be gleaned from exist-  
ing metrics, then gathering global-level  
data for an array of new metrics would be  
a costly endeavor. But the inconsistent de-  
livery measures of biodiversity change  
means that a set of agreed metrics of bio-  
diversity is urgently required (2). Striking  
the right balance between expanding exist-  
ing datasets and developing new, more  
appropriately designed monitoring pro-  
grams and metrics will be vital if measures  
of biodiversity change are to robustly  
support conservation decisions.

In doing so, conservation science must  
be rigorous. Testing the modeled perfor-  
mance of alternative management actions  
prior to implementation should be the  
gold standard for conservation decision  
making. Indicators of change must also be  
subjected to rigorous performance tests.  
Such evaluation was mentioned in the se-  
lection of indicators of the CBD 2010 tar-  
get, with all indicators “identified for im-  
mediate testing.” Yet, with few exceptions,  
the indicators remain largely unevaluated  
in their capacity to report meaningfully on  
conservation targets and the means of  
achieving them; this remains a critical task  
for predictive conservation science if it is  
to influence conservation progress.

### **REFERENCES AND NOTES**

1. S. H. M. Butchart *et al.*, *Science* **328**, 1164 (2010).
2. H. M. Pereira *et al.*, *Science* **339**, 277 (2013).
3. B. Collen *et al.*, *Conserv. Biol.* **23**, 317 (2009).
4. D. P. Tittensor *et al.*, *Science* **xxx**, xxx (2014).
5. G. M. Mace *et al.*, *Biodiversity: Millennium Ecosystem Assessment: current state and trends: findings of the condition and trends working group. Ecosystems and Human well-being* (Island Press, 2005).
6. Secretariat of the Convention on Biological Diversity, *Global Biodiversity Outlook 3* (2010).
7. T. A. Branch *et al.*, *Nature* **468**, 431 (2010).

- 1 8.E. Fulton, A. Smith, A. Punt, *ICES J. Mar.*  
2 *Sci.* **62**, 540 (2005).  
3 9.F. J. Mueter, N. A. Bond, J. N. Ianelli, A. B.  
4 Hollowed, *ICES J. Mar. Sci.* **68**, 1284 (2011).  
5 10.E. Nicholson *et al.*, *PLoS One* **7**, e41128  
6 (2012).  
7 11.L. T. Kelly, A. F. Bennett, M. F. Clarke, M. A.  
8 McCarthy, *Conserv. Biol.* **10**, 1111/cobi.12384 (2014).  
9 12.J. S. Link *et al.*, *ICES J. Mar. Sci.* **67**, 787  
10 (2009).  
11 13.M. McCarthy *et al.*, *Conserv. Biol.* **00**, 1  
12 (2014).  
13 14.S. T. Buckland A. E. Magurran, J. B. Illian, S.  
14 E. Newson, *Ecosphere* **2**, 1 (2011).

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17 **The power of prediction.** Predictive model-  
18 ing of the impacts of alternative policies on  
19 biodiversity (A, B and C vs. business-as-  
20 usual, BAU). Assessment of alternative poli-  
21 cies demonstrates their potential contribution  
22 to meeting biodiversity targets.

23 **Recovery or decline?** Two metrics lead to  
24 different conclusions following the regional  
25 extirpation of a hypothetical species to a low  
26 but stable population size.