

# Fuel Processing Technology

## Effect of CO<sub>2</sub> on HCl removal from syngas using normal and modified Ca-based hydrotalcites: a comparative study

--Manuscript Draft--

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<b>Abstract:</b>	<p>MSW pyrolysis and gasification technologies have been recognized as effective means to enhance the resource utilization of MSW and promote a circular economy. However, the presence of HCl gas can significantly impact the quality and application of syngas. To maximize syngas resource utilization, develop highly efficient HCl adsorbent, this study investigates the performance and mechanism of HCl removal from syngas using a conventional hydrotalcite (Mg-Al-CO<sub>3</sub>) and modified Ca-based hydrotalcite (Ca-Mg-Al-CO<sub>3</sub>). The impact of CO<sub>2</sub>, a component naturally presents in syngas, on the performance of both materials, were also investigated. Characterization techniques, including XRD, TGA, SEM, and analysis of pore properties and specific surface area, were employed to understand the underlying reaction mechanism. The results demonstrated that the performance of Ca-Mg-Al-CO<sub>3</sub> was significantly superior to that of conventional Mg-Al-CO<sub>3</sub> sorbents, particularly in the presence of CO<sub>2</sub>. However, the presence of CO<sub>2</sub> had a detrimental impact on the performance of Ca-Mg-Al-CO<sub>3</sub> in HCl removal, and this effect became increasingly pronounced with higher concentrations of CO<sub>2</sub>. TGA results revealed a competitive relationship between HCl and CO<sub>2</sub> during the adsorption process. Additionally, the fitting results of adsorption kinetics suggested that the adsorption reaction of HCl and CO<sub>2</sub> by Ca-Mg-Al-CO<sub>3</sub> followed multiple rate-controlling mechanisms.</p>
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<b>Response to Reviewers:</b>	

**Dear Editor:**

I am pleased to submit an original research manuscript entitled 'Effect of CO<sub>2</sub> on HCl removal from syngas using normal and modified Ca-based hydrotalcites: a comparative study' on behalf of all authors. We are submitting it for consideration as a research article in *FUEL PROCESSING TECHNOLOGY*.

This study investigates the performance and mechanism of HCl removal from waste derived syngas using a conventional hydrotalcite (Mg-Al-CO<sub>3</sub>) and modified Ca-based hydrotalcite (Ca-Mg-Al-CO<sub>3</sub>), analyzing the impact of CO<sub>2</sub>, naturally present in syngas, on both materials. The results demonstrated that the performance of Ca-Mg-Al-CO<sub>3</sub> was significantly superior to that of conventional Mg-Al-CO<sub>3</sub> sorbents when CO<sub>2</sub> is present. According to the in-situ TGA results, a competitive relationship between HCl and CO<sub>2</sub> was observed during the adsorption process. Initially, the adsorption rate of HCl surpassed that of CO<sub>2</sub>, but the adsorption capacity of CO<sub>2</sub> is significantly higher, likely due to its significantly higher concentration, which was hundreds of times greater than that of HCl. This concentration disparity serves as an additional factor contributing to the observed decline in the performance of HCl removal. Additionally, the fitting results of adsorption kinetics revealed that the adsorption reaction of HCl and CO<sub>2</sub> by Ca-Mg-Al-CO<sub>3</sub> followed multiple rate-controlling mechanisms.

We believe that this work will increase the understanding of HCl removal and therefore prompt researchers to test their catalysts under more realistic conditions. We believe this is a good match with the *FUEL PROCESSING TECHNOLOGY*, where the most relevant literature on this topic is published.

The authors guarantee that none of the content in the paper has been published or is under consideration for publication elsewhere. All authors have approved the manuscript and have agreed to this submission. This paper is the original work of the all authors.

Thank you very much for your consideration.

Yours faithfully,

Songshan Cao

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### **The references of potential reviewers:**

- [1] **D. Li**, J. Yang, Y. Zhao, H. Yuan, Y. Chen, Ultra-highly porous carbon from Wasted soybean residue with tailored porosity and doped structure as renewable multi-purpose absorbent for efficient CO<sub>2</sub>, toluene and water vapor capture, *Journal of Cleaner Production*, 337 (2022) 130283.
- [2] E. Garcia, I.F. Ejim, **H. Liu**, Thermogravimetric analysis of co-combustion of a bituminous coal and coffee industry by-products, *Thermochimica acta*, 715 (2022) 179296.
- [3] Y. Ren, C. Cao, H. Hu, S. Lei, X. Yuan, X. Li, **H. Yao**, Transformation behavior and fate of chlorine in polychloroprene (PCP) during its pyrolysis, *Fuel*, 317 (2022) 123573.
- [4] Y. Feng, B. Xiao, K. Bo, H. Chen, **W. Yang**, Controllable preparation of porous Ca-Mg-Al hydroxides based adsorbents and their CO<sub>2</sub> adsorption performances, *Ferroelectrics*, 594 (2022) 44-56.
- [5] T. Jiang, F. Xiao, Y. Zhao, **S. Wang**, X. Ma, High-temperature CO<sub>2</sub> sorbents with citrate and stearate intercalated CaAl hydrotalcite-like as precursor, *Chinese Journal of Chemical Engineering*, 50 (2022) 177-184.

Dear editor and reviewers,

The authors gratefully acknowledge the valuable comments made by the reviewers. We have studied their comments carefully and have made corrections, which we hope meet with their approval.

Editor and Reviewer comments:

**Editor:**

The manuscript needs to be re-written and re-reviewed.

We appreciate your feedback on our manuscript. We take your comments seriously and are committed to improving the quality of our work. We understand the importance of delivering a well-written and well-structured manuscript.

We thank you for your guidance and look forward to the opportunity to provide a significantly improved manuscript for your consideration.

**Reviewer #1:**

We sincerely appreciate your time and efforts in reviewing our manuscript titled "*Effect of CO<sub>2</sub> on HCl removal from syngas using normal and modified Ca-based hydrotalcites: a comparative Study*". We sincerely appreciate your provision for allowing us to undertake some necessary revision of the manuscript. Your insightful suggestions have not only been fully considered but have been the catalyst for a sweeping and

transformative overhaul of the manuscript.

We have carefully crafted detailed responses to each of your insightful suggestions and comments, showcasing the depth and breadth of the revisions undertaken. We look forward to presenting you with a significantly enhanced version of our manuscript, refined to embody a higher standard of scholarly work. We greatly appreciate your feedback, and we would like to address your comments in a comprehensive, point-by-point manner as follows:

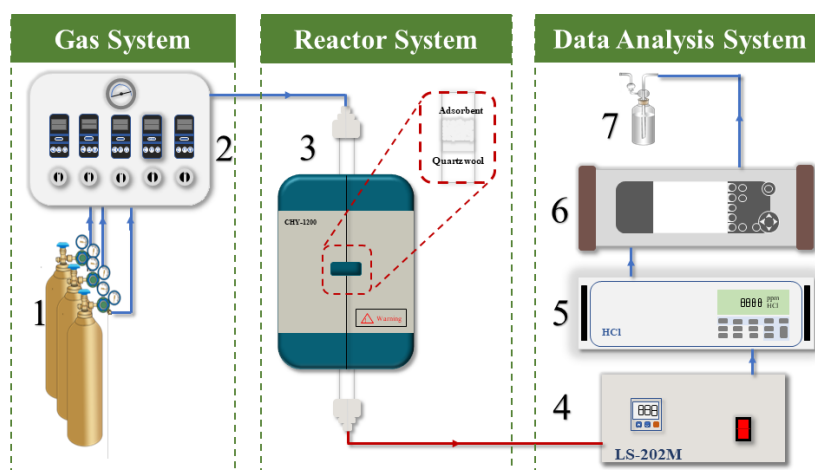
**Issue 1:**

As shown in the schematic diagram of the setup, there is a condenser before the HCl analyzer. There is a concern that part of HCl will condense in this step can cause the deviation of the measurement of the analyzer.

**Response 1:**

Thank you for your careful comments. Due to the operating condition limitations of the HCl analyzer from Signal Group Ltd., UK, in order for the HCl analyzer to operate normally and ensure measurement accuracy, the temperature of the introduced HCl gas must be maintained below 50°C. Therefore, we discussed this issue with engineers from Signal Group Ltd., and as shown in Figure 3A, a condenser has been added. This condenser is used to process gaseous HCl, preventing it from condensing when it enters the analyzer. The purpose of this condenser is to cool and convert HCl gas back into a gaseous state, ensuring accurate measurements. While some condensation of HCl may occur, the design and operation of the condenser aim to minimize this phenomenon and

maintain measurement accuracy. Furthermore, we will conduct regular inspections and maintenance on the condenser to ensure its proper functioning and to avoid significant deviations in measurement results.



*1. Gas cylinder; 2. Mass flowmeter; 3. Reactor; 4. LJ-202M condenser; 5. HCl analyzer; 6. Flue gas analyzer; 7. Gas-washing bottle*

*Figure3A. The schematic diagram of experiments*

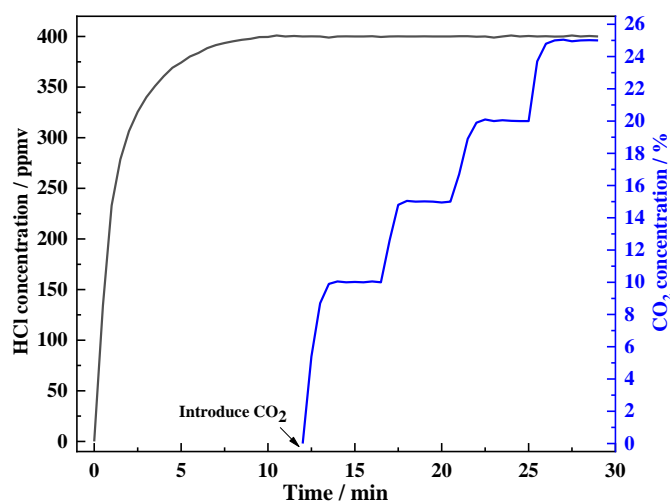
## **Issue 2:**

The presence of CO<sub>2</sub> could possibly affect the accurate measurement of HCl (cross interface between CO<sub>2</sub> and HCl). Blank tests should be conducted without the adsorbents to check the stability of the HCl measurement under different CO<sub>2</sub> concentrations.

## **Response 2:**

Thank you for your comment. Regarding the HCl analyzer, we communicated our requirements to the engineer during the equipment procurement process, emphasizing that HCl should not be affected by other gases such as CO<sub>2</sub>, CO, O<sub>2</sub>, and H<sub>2</sub> during

testing. To address this, the manufacturer added an internal data processing module to the device. Before conducting our experiments, we also performed a blank test involving both CO<sub>2</sub> and HCl to ensure that there was no mutual interference. The relevant test data are as follows: Initially, we measured HCl using the HCl analyzer. Once the HCl gas had stabilized, we introduced CO<sub>2</sub> gas, gradually increasing the concentration from 0% to 25%. The experimental results confirmed that different concentrations of CO<sub>2</sub> had no impact on the HCl analyzer.



The blank test of the impact of CO<sub>2</sub> concentration on HCl concentration

### Issue 3:

In Figure 7D, the HCl capacity was 217.9 for Ca-Mg-Al-CO<sub>3</sub> at 300C and 20% CO<sub>2</sub>. Yet in Figure 8D under the same operation condition, the HCl capacity was 271.9. The repeatability seems questionable. The realizability and repeatability of the data should be double checked for the whole manuscript.

### Response 3:

We apologize for this mistake. After a thorough review, we have recalculated the HCl capacity, and it has been confirmed to be 217.9 mg·g<sup>-1</sup> for Ca-Mg-Al-CO<sub>3</sub> at 300°C and

20% CO<sub>2</sub>. The manuscript has been updated with this corrected data."

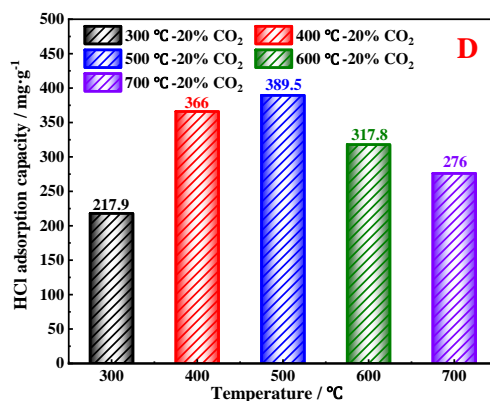


Fig. 8 the maximum breakthrough chlorine capacity

*This behavior is further supported by Fig.8 D, which indicates that the maximum breakthrough chlorine capacity of Ca-Mg-Al-CO<sub>3</sub> increased from 217.9 mg/g to 389.5 mg/g as the reaction temperature increased from 300 to 500 °C then decreased from 389.5 mg/g to 276 mg/g as the reaction temperature further increased from 500 to 700 °C*

#### **Issue 4:**

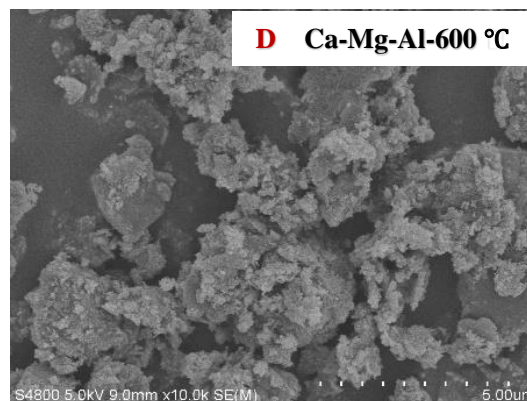
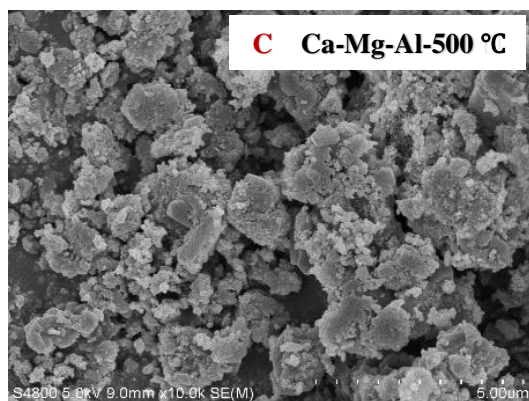
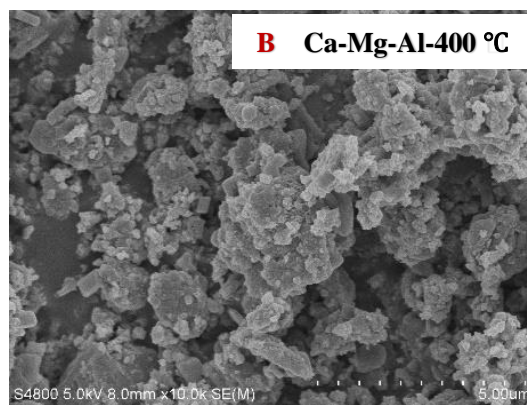
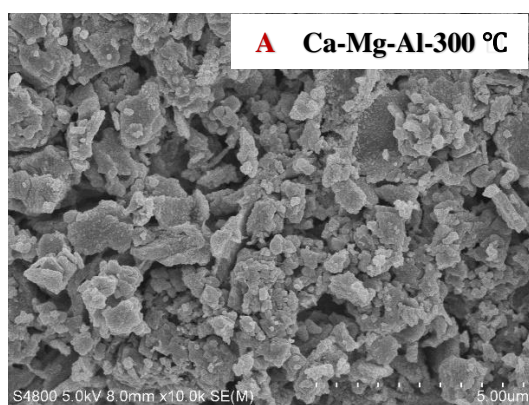
The authors indicated that the pore structure variation at different temperatures would cause the optimum removal capacity of HCl at 500C. Yet no evidence was provided. The surface area, pore size distribution, and surface morphology at different temperatures should be provided to support their assumptions.

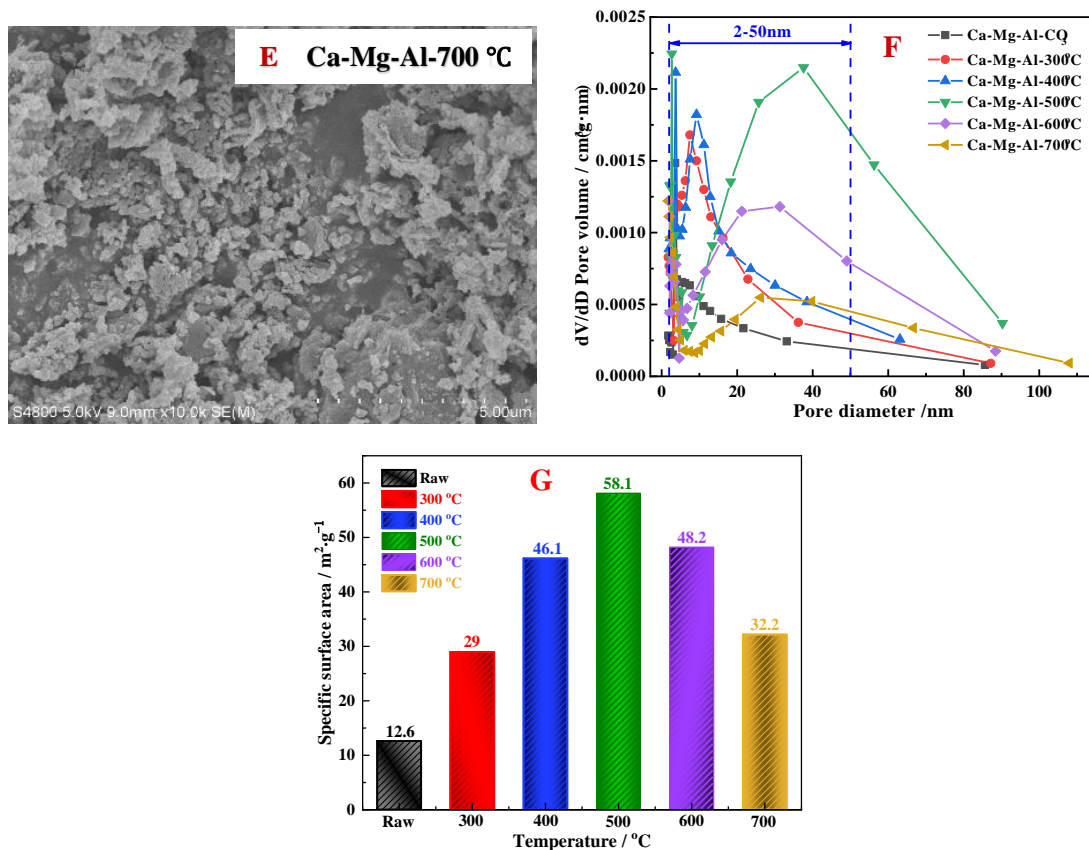
#### **Response 4:**

According to the comment of the reviewer, we would like to provide a detailed explanation regarding the adsorption mechanism as follows, and we have added the explanation to the manuscript.



The surface morphology, pore size distribution, and specific surface area of Ca-Mg-Al- $\text{CO}_3$  at different temperatures were shown in Fig.10. As shown in the SEM images of Fig 10 A-E, with increase of temperature, surface morphology of Ca-Mg-Al- $\text{CO}_3$  becomes fluffy, however, there is a sintering phenomenon when the temperature exceeds 500 °C In Fig.10 F, the pores of Ca-Mg-Al- $\text{CO}_3$  are mainly mesopores. With the increase of temperature, the number of pores first increases and then decreases, and pores develop towards larger pores, the pore structure is optimal at 500 °C. Therefore, as shown in Fig. 10 G, the specific surface area also first increases and then decreases, the specific surface area was larger at 500 °C 58.1  $\text{m}^2/\text{g}$ . The large specific surface area provides more active sites, facilitates the availability of more active sites. Above results further illustrated that why the Ca-Mg-Al- $\text{CO}_3$  exhibits excellent performance in removing HCl at 500 °C.





*Fig. 10 The surface morphology, pore size distribution, and specific surface area of Ca-Mg-Al-CO<sub>3</sub> at different temperatures (A-E: Surface morphology; F: Pore size distribution; G: Specific surface area)*

### Issue 5:

Why the intra-particle diffusion model, Elovich model and Bangham model were selected for fitting study? All of them are faired to present the q variation at the later adsorption stage, even though they showed R2 higher than 0.9. The conclusion made based on this kinetic study may not be reasonable.

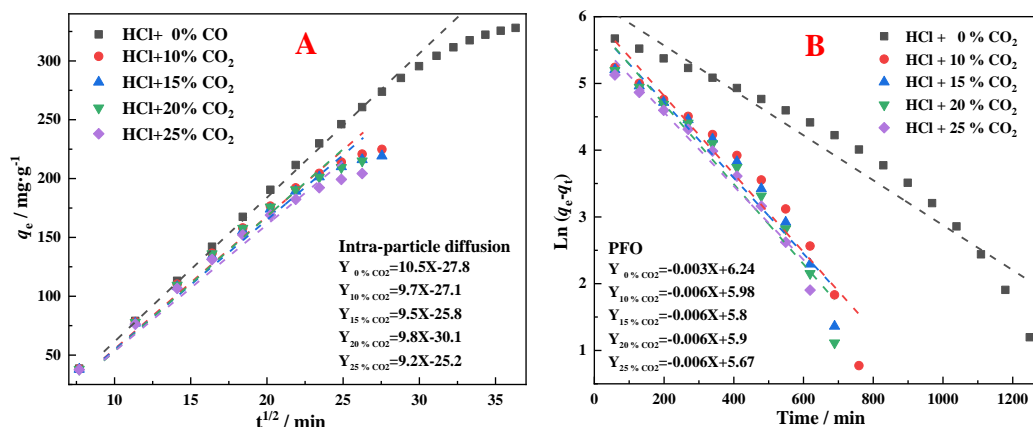
### Response 5:

Thank you for your comment. We reference the literatures and then selected five kinetic fitting models for gas-solid reaction, therefore, we reanalyzed the data and drew new

figures. Based on the results of fitting, the fitting results of the PSO is the best, indicating that the chemical adsorption is main reaction mechanism throughout the entire reaction process, and chemical adsorption involves electron sharing or electron transfer between Ca-Mg-Al-CO<sub>3</sub> and CO<sub>2</sub>, HCl. Other fitting results also are good in the early stage of the reaction, showing that the reaction was controlled by multiple rate-controlling mechanisms. The new results are as follows,

### 3.6 Apparent adsorption kinetic

The apparent adsorption kinetic models mainly include kinetic control type and diffusion control type. The diffusion control type includes Intra-particle diffusion model, which describes gas film or intra-particle diffusion as a rate-controlling step. The commonly used kinetic control models are Pseudo-first-order model (PFO), Pseudo-second-order model (PSO), Elovich model and Bangham model, all of which describe the kinetic adsorption process of surface adsorption or surface chemical reactions as rate-controlling steps. The experimental results of HCl removal by Ca-Mg-Al-CO<sub>3</sub> at different CO<sub>2</sub> concentrations were analyzed using the above models, and the fitting results of adsorption kinetics are presented in Fig. 12.



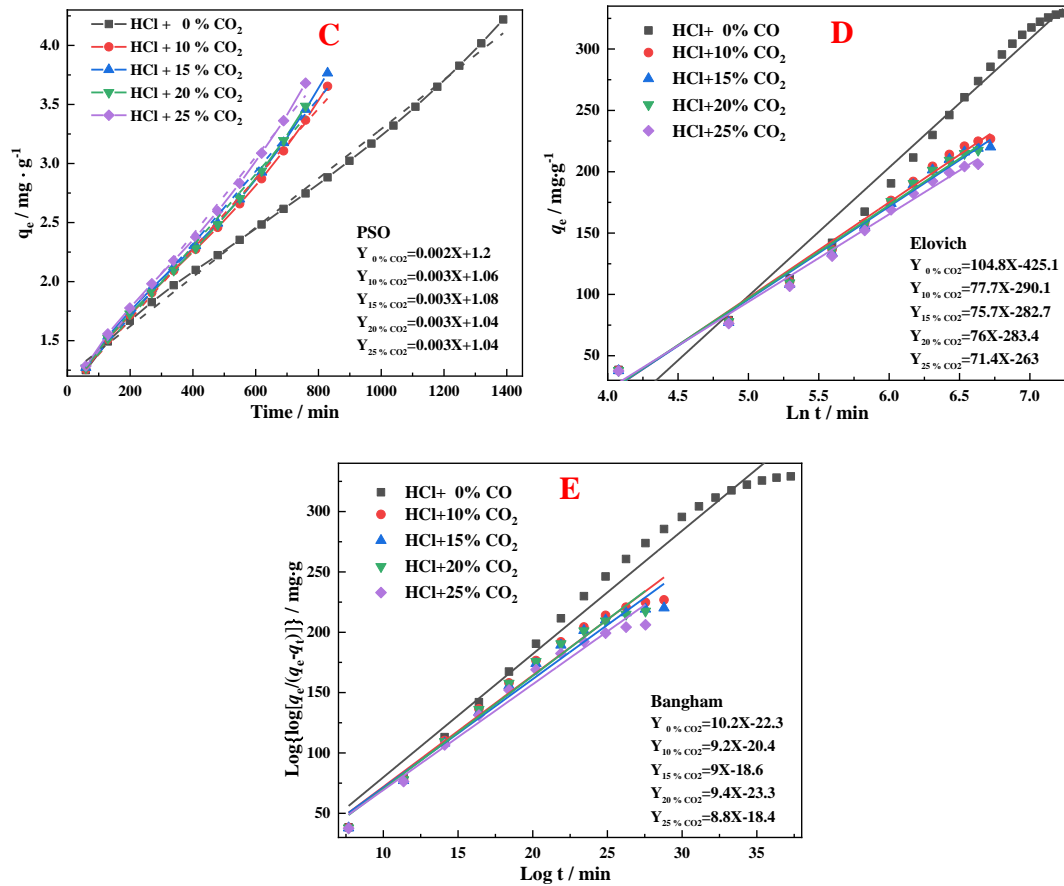


Fig. 12 The fitting results of different adsorption kinetic model (A: Intra-particle diffusion model; B: Pseudo-first-order model (PFO), C: Pseudo-second-order model; D: Elovich model; E: Bangham model)

As depicted in Fig. 12 and the correlation coefficients ( $R^2$ ) summarized in Table 2, based on the fitting result of the intra-particle diffusion model, the fitted line does not pass through the origin and did not exhibit a linear relationship, the  $R^2$  values are more than 0.98, indicating that diffusion control is not the solely governed by single rate-determining steps. The fitting results of 4 kinetic control types showed a relatively best agreement with the PSO model, the  $R^2$  values exceed 0.99. The result illustrates that the chemical adsorption is main reaction mechanism in whole reaction process, and chemical adsorption involves electron sharing or electron transfer between Ca-Mg-Al-

$CO_3$  and  $HCl$ ,  $CO_2$ . Compared to  $PSO$  model. Although the  $R^2$  values of other 3 adsorption kinetic models were slightly lower than that of the  $PSO$  model, all models exhibited correlation coefficients above 0.97, as shown in Fig. 12B, D and E, there a good fitting result in the early stage of the reaction, indicating it is controlled by multiple control steps in the early stage of the reaction.

*Table. 2 The correlation coefficients ( $R^2$ ) of different adsorption kinetic models*

<i>Model</i>	<i>0% CO<sub>2</sub></i>	<i>10% CO<sub>2</sub></i>	<i>15% CO<sub>2</sub></i>	<i>20% CO<sub>2</sub></i>	<i>25% CO<sub>2</sub></i>
	<i>+HCl</i>	<i>+ HCl</i>	<i>+ HCl</i>	<i>+HCl</i>	<i>+HCl</i>
<i>Intra-particle</i>					
<i>diffusion</i>	0.983	0.985	0.983	0.987	0.985
<i>PFO</i>	0.94	0.933	0.946	0.943	0.968
<i>PSO</i>	0.996	0.996	0.994	0.996	0.996
<i>Elovich</i>	0.98	0.988	0.987	0.989	0.99
<i>Bangham</i>	0.977	0.976	0.972	0.978	0.974

**Issue 6:**

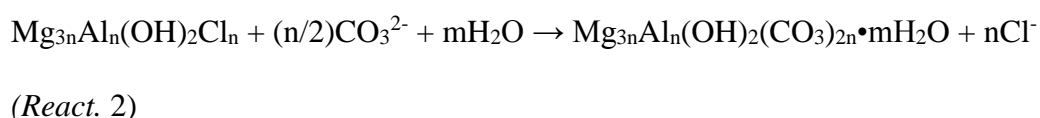
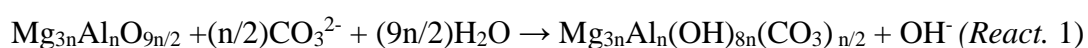
Can the absorbent be regenerated for cyclic utilization?

**Response 6:**

Thank you for your comment. Regeneration of the absorbent is indeed a critical aspect, and we appreciate your interest in this matter.

According to the properties of hydrotalcite, which include its layered structure, exchangeability of interlayer anions, thermal stability, and memory effect, it is known

that under certain conditions, hydrotalcite can be restored to its original form. However, the effectiveness of the restoration process is dependent on the reaction temperature. If the temperature exceeds a specific threshold, the structure of hydrotalcite may be irreversibly altered, making restoration challenging. Based on our experiments, we have determined that the Mg-Al-CO<sub>3</sub> sample retained a significant portion of its hydrotalcite structure with only minimal decomposition under 400°C. To regenerate the sorbent, an anion exchange process was employed as depicted in *React. 1* and *2*.



To replenish the missing water and carbonate ions in the product, in the test, the product was immersed in a sodium carbonate solution of a specific concentration. Then the repaired Mg-Al-400 was obtained by calcining the product at 400 °C. This repaired sorbent was then subjected to breakthrough tests to remove HCl, under the following reaction conditions: an HCl concentration of 400 ppmv, a reaction temperature of 300 °C, and a gas flow of 500 ml/min. The experimental results indicate that even after cycling the sorbent three times, the adsorption capacity of HCl remained at approximately 50%. However, a subsequent decline in performance was observed. There are two main reasons for the decrease in performance. Firstly, according to *React. 1* and *2*, the restoration of Mg-Al-CO<sub>3</sub> and the subsequent desorption of Cl<sup>-</sup> could not reach 100%. Secondly, a portion of the Mg<sup>2+</sup> may have already reacted with Cl<sup>-</sup> to generate MgCl<sub>2</sub>. Therefore, the recovery is not 100% and the crystallinity is reduced

during the recovery process.

Regarding calcium based modified hydrotalcite, the structure of Ca-Mg-Al-CO<sub>3</sub> is similar to Mg-Al-CO<sub>3</sub>, suggesting the potential for regeneration. We apologize for not conducting a detailed study on the regeneration ability of Ca-Mg-Al-CO<sub>3</sub>. In our future work, we plan to explore and provide comprehensive insights into the regeneration and recyclability of hydrotalcites for HCl removal. Consequently, we are actively exploring improved methods to enhance the recyclability of the sorbent, aiming to achieve environmental sustainability and cost reduction.

We appreciate your feedback, and we will ensure that the regeneration experiments and their outcomes are included in our future work to provide a comprehensive understanding of the regeneration and cycle usage of hydrotalcites for HCl removal.

**Reviewer #2:**

We would like to express our sincere gratitude for your time and effort invested in the review of our manuscript titled "*Effect of CO<sub>2</sub> on HCl removal from syngas using normal and modified Ca-based hydrotalcites: a comparative Study*". We are thankful for your consideration in allowing us to make essential revisions to our manuscript.

Your valuable suggestions have not only been thoroughly taken into account but have also played a pivotal role in instigating a comprehensive transformation of our paper.

We have diligently prepared detailed responses to each of your insightful recommendations and comments, illustrating the extensive scope of the revisions that have been carried out. We are eagerly looking forward to presenting you with an

upgraded version of our manuscript, refined to meet the highest standards of academic work. Your feedback is highly appreciated, and we are committed to addressing your comments meticulously, item by item, as outlined below:

**Issue 1:**

I do not like this abstract. It did not tell the reader why you did this work, and which methodology you take to do the research. I mean, this abstract is more like a conclusion, rather than an abstract.

**Response 1:**

Thank you for your comment, we have modified abstract.

*MSW pyrolysis and gasification technologies have been recognized as effective means to enhance the resource utilization of MSW and promote a circular economy. However, the presence of HCl gas can significantly impact the quality and application of syngas. To maximize syngas resource utilization, develop highly efficient HCl adsorbent, this study investigates the performance and mechanism of HCl removal from syngas using a conventional hydrotalcite (Mg-Al-CO<sub>3</sub>) and modified Ca-based hydrotalcite (Ca-Mg-Al-CO<sub>3</sub>). The impact of CO<sub>2</sub>, a component naturally presents in syngas, on the performance of both materials, were also investigated. Characterization techniques, including XRD, TGA, SEM, and analysis of pore properties and specific surface area, were employed to understand the underlying reaction mechanism. The results demonstrated that the performance of Ca-Mg-Al-CO<sub>3</sub> was significantly superior to that of conventional Mg-Al-CO<sub>3</sub> sorbents, particularly in the presence of CO<sub>2</sub> However, the*



*presence of CO<sub>2</sub> had a detrimental impact on the performance of Ca-Mg-Al-CO<sub>3</sub> in HCl removal, and this effect became increasingly pronounced with higher concentrations of CO<sub>2</sub>. TGA results revealed a competitive relationship between HCl and CO<sub>2</sub> during the adsorption process. Additionally, the fitting results of adsorption kinetics suggested that the adsorption reaction of HCl and CO<sub>2</sub> by Ca-Mg-Al-CO<sub>3</sub> followed multiple rate-controlling mechanisms.*

**Issue 2:**

Pages missing.

**Response 2:**

Thank you for your comment, we have added page numbers and verified the manuscript to ensure that the page numbers are accurate.

**Issue 3:**

"According to the in-situ TGA results,..." But TGA are always in-situ, right?

**Response 3:**

Thank you for your comment, we have changed in-situ TGA to TGA, and deleted 'in-situ'.

**Issue 4:**

Please reorganize the introduction to highlight the novelty and contribution of this work.

(1) Please consider the significance of studying the CO<sub>2</sub> concentration on the removal of HCl. Can you control the CO<sub>2</sub> concentration in the pyrolysis gas??

#### Response 4:

Thank you for your comment, based on your suggestion, we have reorganized the introduction to highlight the novelty and contribution. The modification content is as follows,

*It is important to note that CO<sub>2</sub>, a non-combustible gas, is present in syngas at significantly higher concentrations than HCl. CO<sub>2</sub> is a key component in various processes, including the water-gas shift reaction, producing CO and so on, It has been established that hydrotalcite can adsorb CO<sub>2</sub> effectively.[42, 43] CO<sub>2</sub> is also confirmed to be adsorbed by hydrotalcite [44, 45]. Therefore, both HCl and CO<sub>2</sub> can be adsorbed by hydrotalcite. Especially in the case of Ca-Mg-Al-CO<sub>3</sub>, it exhibits remarkable capabilities in adsorbing CO<sub>2</sub> due to the generation of calcium-containing oxide at mid-high temperatures[46, 47]. Consequently, the presence of CO<sub>2</sub> may influence the performance of HCl removal using hydrotalcite and Ca-Mg-Al-CO<sub>3</sub>, although the specific effect tendency is not yet well understood. Although some studies have reported on the effect of CO<sub>2</sub> on HCl removal using other calcium-based compounds such as CaO, Ca(OH)<sub>2</sub> and CaCO<sub>3</sub>[48, 49], it should be noted that the research objectives and operating conditions varied, resulting in inconsistent results. Therefore, when studying the performance of HCl removal in syngas, the effect of CO<sub>2</sub> on HCl removal using hydrotalcite and hydrotalcite-like compounds should not be overlooked. Surprisingly, there is a lack of research focusing on the characteristics of HCl removal using hydrotalcite and hydrotalcite-like compound in the presence of CO<sub>2</sub>. Therefore, it is necessary to analyze the effect and mechanism of CO<sub>2</sub> on HCl removal by adsorbent.*

*Obtaining the impact tendency of CO<sub>2</sub> on HCl removal by hydrotalcite is crucial for ensuring the production of high-quality syngas and expanding the applications of syngas in various fields.*

**Issue 5:**

Line 49. as small footprint and higher energy efficiencies, . but footprint of whom?? CO2 footprint??> If it weren't for the pressure of carbon footprint, these solid wastes could all be landfilled and composted, so there's no need for such a complex disposal process. You can start by rewriting the introduction from the perspective of energy consumption and CO2 emission reduction. At present, the background and significance of this MS version is not enough persuasive. In addition, here the CO2 footprint should cite some related references such as "[1] Integrated CO2 capture and utilization with CaO-alone for high purity syngas production. Carbon Capture Science & Technology 1 (2021) 100001 [2]Use of copper carbonate as corrosion inhibitor for carbon steel in post combustion carbon capture. Carbon Capture Science & Technology.6(2023) 100095"

**Response 5:**

Thank you for your comment, based on your suggestion, we have made modifications to introduction and added relevant references, including those recommended by you.

*The disposal of municipal solid waste (MSW) has become a pressing issue due to rapid growth of human population and associated industrial activities[1]. The proper utilization of MSW can significantly contribute to the goal of carbon emission reduction[2]. The modern waste management strategies prioritize waste minimization,*

*recycling, and reuse, with landfill disposal considered as the least desirable option to prevent pollution of surface water, groundwater, soil and air, as well as to reduce GHG emissions from the landfill sites. In the background of global carbon neutrality, finding rational ways to utilize MSW is of paramount importance for reducing energy waste and carbon dioxide emissions [3, 4].*

...

*These technologies not only enable waste disposal while reducing emissions of gaseous pollutants such as NO<sub>x</sub> and SO<sub>x</sub>, but also exhibit characteristics such as a small footprint and higher energy efficiencies[8]. Especially in the context of achieving carbon neutrality, the use of pyrolysis and gasification technology for waste processing becomes particularly vital in reducing the pressure of carbon emissions[9, 10].*

**Issue 6:**

Line 72. title case of therefore?

**Response 6:**

We apologize for this mistake. We have changed 'therefore' to 'Therefore'.

**Issue 7:**

2.1 Synthetic steps,. Here, there are so many paragraphs that are only one sentence.

**Response 7:**

Thank you for your comment, we have reorganized section 2.1, which is as follows,

*In this study, the chemical reagents of Mg (NO<sub>3</sub>)<sub>2</sub>·6H<sub>2</sub>O, Al (NO<sub>3</sub>)<sub>3</sub>·9H<sub>2</sub>O, Ca (NO<sub>3</sub>)<sub>2</sub>·4H<sub>2</sub>O, NaOH and Na<sub>2</sub>CO<sub>3</sub> with analytically pure grade were produced from*

Sinopharm Chemical Reagent Co., Ltd. The Mg-Al hydrotalcite containing  $\text{CO}_3^{2-}$  intercalated (Mg-Al- $\text{CO}_3$ ), and Ca-based modified hydrotalcite-like (Ca-Mg-Al- $\text{CO}_3$ ) were synthesized using the coprecipitation method for gaseous HCl removal. The synthetic process is shown in Fig. 2 and the synthetic steps are as follows.

*Preparation of Mg-Al- $\text{CO}_3$ :* (1) According to the molar ratio of  $\text{Mg}/\text{Al}=3$ ,  $[\text{OH}^-]=2([\text{Mg}^{2+}] + [\text{Al}^{3+}])$  and  $[\text{CO}_3^{2-}] = 1/2[\text{Al}^{3+}]$ , the corresponding chemical reagents were weighed. (2) The nitrates and two bases were dissolved in ultra-pure water at 60 °C, respectively. (3) The two solutions were dropped into a four-port flask at a certain speed using a peristaltic pump at 60 °C. During this process, the pH value was maintained at 11, and the stirring speed was set at 300 r/min. The resulting mixture was continuously stirred for 2 hours at 60 °C. (4) The obtained substance was subjected to crystal growth for 20 hours at 80 °C. (5) The solid part was washed using ultra-pure water until neutral, dried and grounded. The Mg-Al- $\text{CO}_3$  was obtained, and the particles in size range of 40 to 60 mesh were chosen for subsequent experiments.

*Preparation of Ca-Mg-Al- $\text{CO}_3$ :* The Ca-Mg-Al- $\text{CO}_3$  was prepared with a molar ratio of  $\text{Ca}^{2+}/\text{Mg}^{2+} = (\text{Ca}+\text{Mg})^{2+}/\text{Al}^{3+} = 3$ ,  $[\text{OH}^-]=2([\text{Ca}+\text{Mg}]^{2+}] + [\text{Al}^{3+}]$  and  $2[\text{CO}_3^{2-}] = [\text{Al}^{3+}]$ . The step (2) to (5) mentioned above were repeated to obtain Ca-Mg-Al- $\text{CO}_3$ , and the particle sizes range was the same as that of Mg-Al- $\text{CO}_3$ .

**Issue 8:**

The Figure 3 should better be moved to the SI.

**Response 8:**

Thank you for your suggestion. Figure 3 illustrates the schematic diagram of both the HCl removal experiment and the TGA experiment. We believe that Figure 3 is

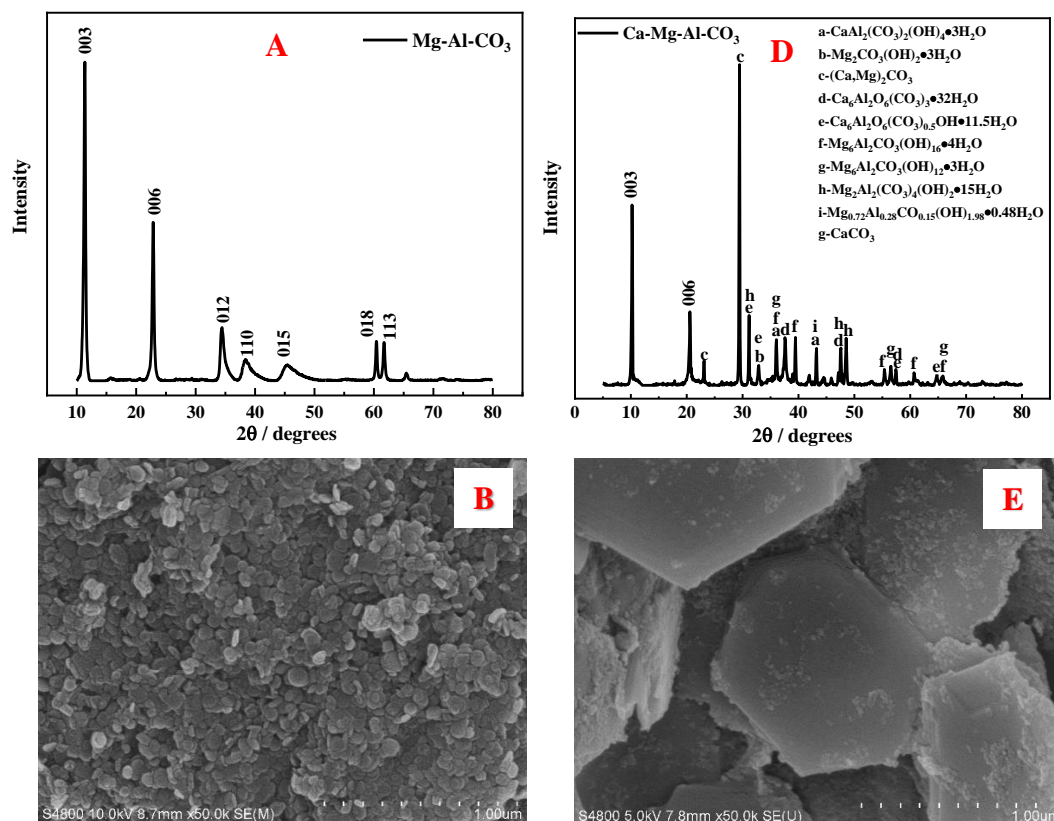
beneficial for readers to have a clearer understanding of the experimental steps and methods, facilitating a better grasp of the author's experimental intentions.

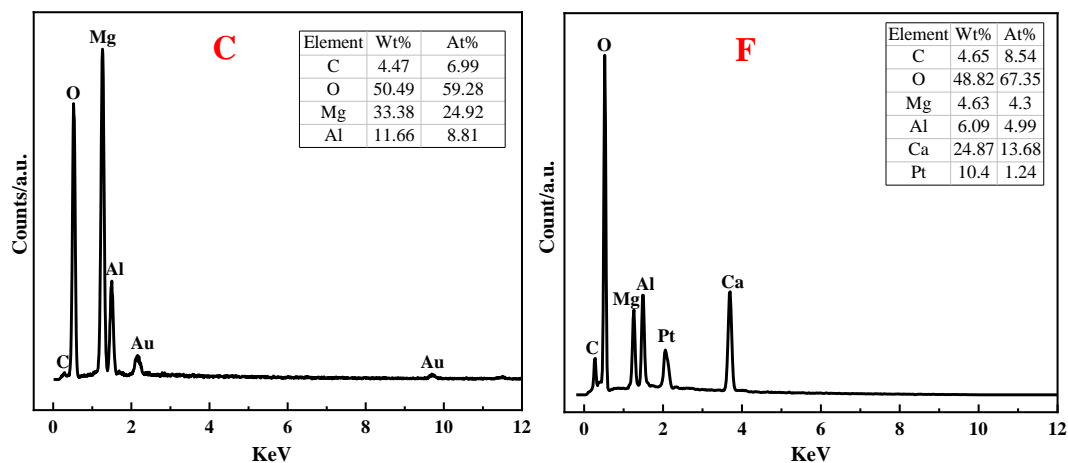
**Issue 9:**

Figure 4. The scare bar is Unclear.

**Response 9:**

Thank you for your comment. Based on the reviewer's feedback, we have made the necessary modifications to Figure 4 and replaced the SEM image to ensure that Figure 4 is clear and visually accurate. We appreciate the reviewer's attention to detail and their valuable feedback.





*Fig. 4 Characterization of adsorbents (Mg-Al-CO<sub>3</sub>: A: XRD pattern, B: SEM image, C: EDS pattern; Ca-Mg-Al-CO<sub>3</sub>: D: XRD pattern, E: SEM image, F: EDS pattern)*

**Issue 10:**

The Figures all should better be reorganized. The explanation of the content can be placed below (caption), and it is important to ensure the clarity and aesthetics of the images.

**Response 10:**

Thank you for your comment. We have addressed your suggestion to enhance the clarity and aesthetics of the figures. Based on the reviewer's feedback, we have made modifications to all figures in the manuscript. An example of the modified figure is shown below (using Fig. 7 as an illustration).

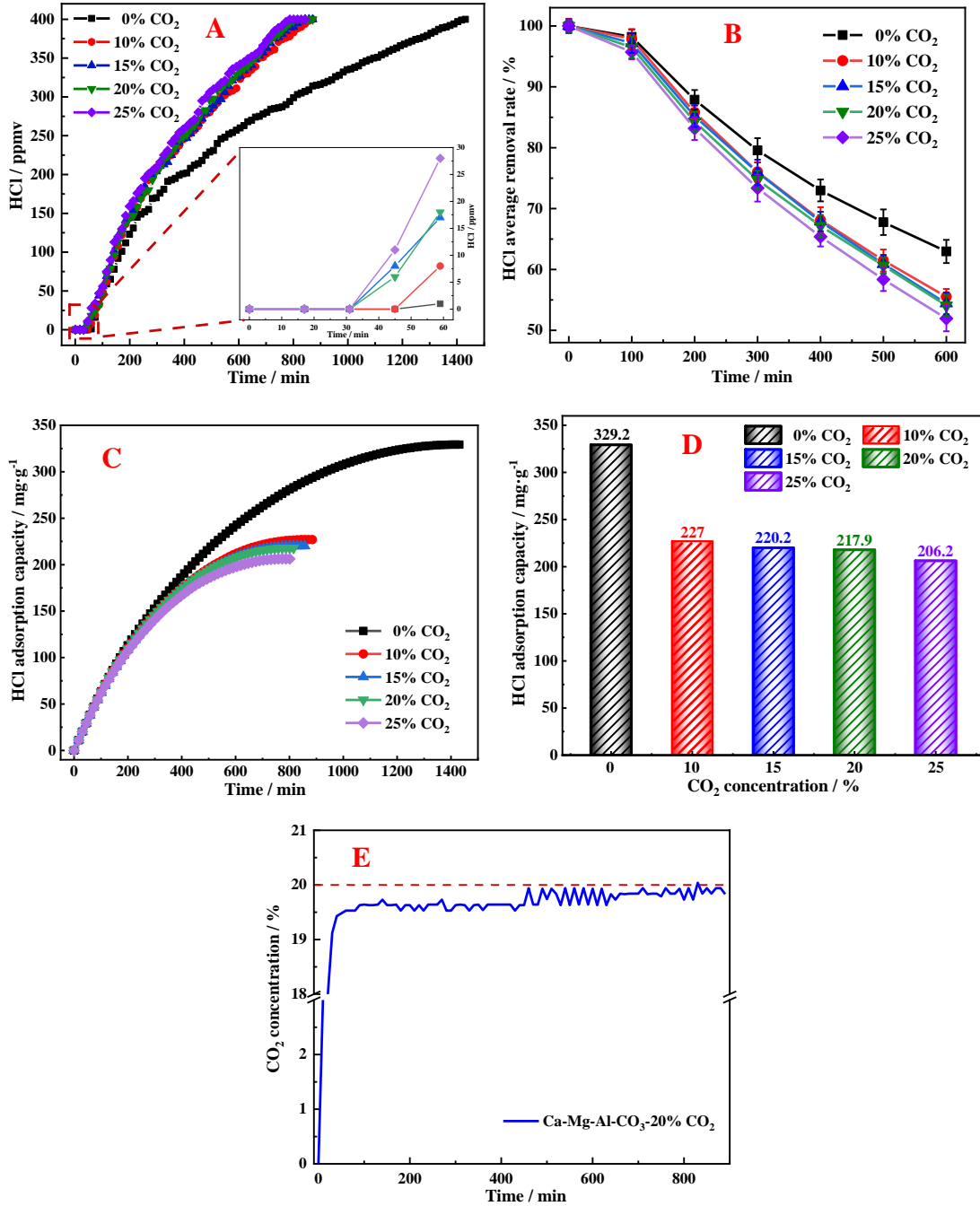


Fig.7 The performance of HCl removal using  $\text{Ca-Mg-Al-CO}_3$  under different  $\text{CO}_2$  concentrations and the change curve of  $\text{CO}_2$  (A: the breakthrough curves; B: the average removal of HCl rate; C: the adsorption chlorine capacity curves; D: the maximum breakthrough chlorine capacity; E: the curve of  $\text{CO}_2$  concentration during the HCl removal process)



### Issue 11:

Fig9. The XRD results. Please consider to cut the angle below 10 (Leave too much white space).

### Response 11:

Thank you for your comment. Based on reviewer's feedback, we have made modifications to Figure 7, and the revised image is shown below,

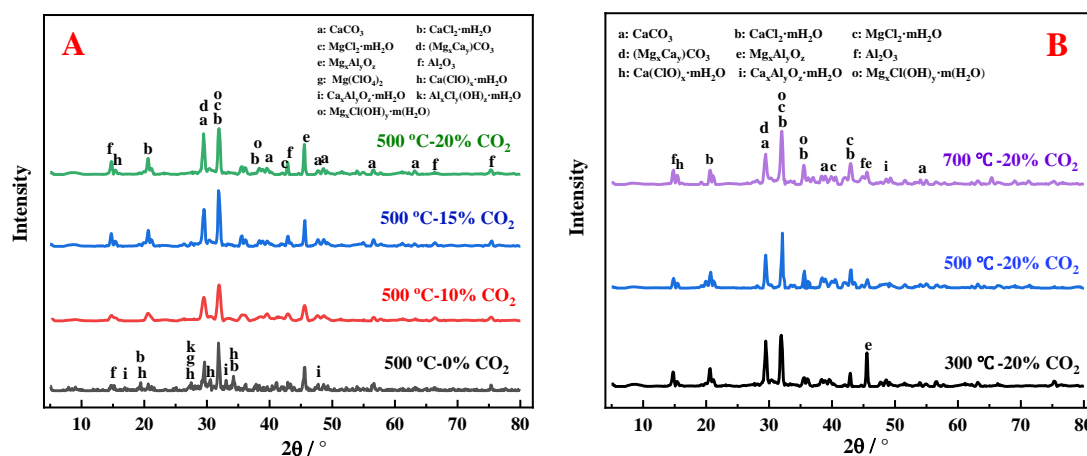


Fig. 9 The XRD patterns of reaction products at different  $\text{CO}_2$  concentrations and temperatures (A: the different  $\text{CO}_2$  concentrations; B: the different temperatures)

### Issue 12:

4 conclusions. Why are the fonts in this section different? The last sentence "These findings align with the proposed reaction mechanism." Is this the key point of the work? It would be better to reach the academic and industrial significance.

### Response 12:

Thank you for your comment. Firstly, we apologize for the formatting issue, and we have modified. Secondly, based on your suggestion, we have revised the conclusion, which is as follows,

#### **4. Conclusions**

*In this study, the performance of HCl removal using Mg-Al-CO<sub>3</sub> and Ca-Mg-Al-CO<sub>3</sub>, as modified sorbent, in the presence of CO<sub>2</sub> were investigated.*

*The Ca-Mg-Al-CO<sub>3</sub> was successfully synthesized with a larger layer spacing  $d_{003}$  and particle size compared to Mg-Al-CO<sub>3</sub>. CO<sub>2</sub> had a positive impact on the removal of HCl using Mg-Al-CO<sub>3</sub>, but a negative impact on the removal of HCl using Ca-Mg-Al-CO<sub>3</sub>. However, the performance of HCl removal using Ca-Mg-Al-CO<sub>3</sub> was superior to that of Mg-Al-CO<sub>3</sub>. The main reason for this difference is that CO<sub>2</sub> can react with Mg-Al-CO<sub>3</sub> to generate new hydrotalcite intercalated anions through a regeneration process, thereby enhancing the performance of HCl removal. In contrast, Ca-Mg-Al-CO<sub>3</sub> reacted with CO<sub>2</sub> to form large particles of CaCO<sub>3</sub>, which not only occupied the active sites but also hindered the reaction between HCl and the adsorbent, leading to a decrease in the performance of HCl removal. The main products of reaction between Ca-Mg-Al-CO<sub>3</sub> and HCl were CaCl<sub>2</sub> and MgCl<sub>2</sub>. The presence of CO<sub>2</sub> plays a competitive role in the removal of HCl. The adsorption rate of HCl was initially faster than that of CO<sub>2</sub>, but the adsorption capacity of CO<sub>2</sub> was stronger. This can be attributed to the significantly higher concentration of CO<sub>2</sub> compared to HCl. The adsorption reactions of HCl and CO<sub>2</sub> by Ca-Mg-Al-CO<sub>3</sub> was controlled by multiple rate-controlling mechanisms in the early stage of the reaction, and follow a PSO model in the whole process. The study demonstrates that Ca-Mg-Al-CO<sub>3</sub> exhibits excellent performance in removing HCl, which is of significant importance for achieving high-quality synthesis gas and further promoting pyrolysis gasification technology. The findings have academic and industrial significance in advancing waste-to-energy conversion technologies and contributing to the reduction of carbon emissions.*

**Issue 13:**

Superscript and subscript.

**Response 13:**

Thank you for your comment. Based on the reviewer's feedback, we have thoroughly reviewed the superscripts and subscripts, and made the necessary corrections to ensure their accuracy and proper formatting in the revised manuscript.

## Highlights

- The performance of Ca-Mg-Al-CO<sub>3</sub> was superior to that of Mg-Al-CO<sub>3</sub> when CO<sub>2</sub> existed
- A competitive relationship between HCl and CO<sub>2</sub> was observed during the adsorption
- The adsorption of HCl and CO<sub>2</sub> followed multiple rate-controlling mechanisms

# Effect of CO<sub>2</sub> on HCl removal from syngas using normal and modified Ca-based hydrotalcites: a comparative study

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**Abstract:** MSW pyrolysis and gasification technologies have been recognized as effective means to enhance the resource utilization of MSW and promote a circular economy. However, the presence of HCl gas can significantly impact the quality and application of syngas. To maximize syngas resource utilization, develop highly efficient HCl adsorbent, this study investigates the performance and mechanism of HCl removal from syngas using a conventional hydrotalcite (Mg-Al-CO<sub>3</sub>) and modified Ca-based hydrotalcite (Ca-Mg-Al-CO<sub>3</sub>). The impact of CO<sub>2</sub>, a component naturally presents in syngas, on the performance of both materials, were also investigated. Characterization techniques, including XRD, TGA, SEM, and analysis of pore properties and specific surface area, were employed to understand the underlying reaction mechanism. The results demonstrated that the performance of Ca-Mg-Al-CO<sub>3</sub> was significantly superior to that of conventional Mg-Al-CO<sub>3</sub> sorbents, particularly in the presence of CO<sub>2</sub>. However, the presence of CO<sub>2</sub> had a detrimental impact on the performance of Ca-Mg-Al-CO<sub>3</sub> in HCl removal, and this effect became increasingly pronounced with higher concentrations of CO<sub>2</sub>. TGA results revealed a

26 competitive relationship between HCl and CO<sub>2</sub> during the adsorption process. Additionally, the  
27 fitting results of adsorption kinetics suggested that the adsorption reaction of HCl and CO<sub>2</sub> by Ca-  
28 Mg-Al-CO<sub>3</sub> followed multiple rate-controlling mechanisms.

29 Keywords: Ca-Mg-Al-CO<sub>3</sub>; CO<sub>2</sub>; HCl removal; TGA; Adsorption kinetics

## 30 **1. Introduction**

31 The disposal of municipal solid waste (MSW) has become a pressing issue due to rapid growth  
32 of human population and associated industrial activities[1]. The proper utilization of MSW can  
33 significantly contribute to the goal of carbon emission reduction[2]. The modern waste  
34 management strategies prioritize waste minimization, recycling, and reuse, with landfill disposal  
35 considered as the least desirable option to prevent pollution of surface water, groundwater, soil and  
36 air, as well as to reduce GHG emissions from the landfill sites. In the background of global carbon  
37 neutrality, finding rational ways to utilize MSW is of paramount importance for reducing energy  
38 waste and carbon dioxide emissions [3, 4].

39 The technologies for recovering energy from “residual waste” (i.e., remaining waste that cannot  
40 be economically or practically reused or recycled) can play a critical role in mitigating the  
41 environmental issues associated to waste disposal. Aside from the valuable product, these  
42 technologies can result in a large decrease in the overall amounts of material requiring final  
43 disposal. This allows for simpler management in a controlled way while still adhering to pollution  
44 control regulations [5]. A host of technologies are available for realizing the potential of residual  
45 waste as an energy source (as power or fuel), but the availability and general composition of waste  
46 affects the technologies that are suitable to deliver environmental benefits[6]. Thermochemical  
47 technologies have historically been used to produce heat and electricity (Waste-to-Energy, or WtE)

48 via incineration of the waste feedstock, alone or together with other fuels. Most recently, pyrolysis  
49 and gasification technologies have been recognized as better methods for achieving higher  
50 resource utilization of MSW and circular economy[7]. Compared to traditional treatment methods  
51 such as incineration and landfill, pyrolysis and gasification technologies offer several advantages.  
52 These technologies not only enable waste disposal while reducing emissions of gaseous pollutants  
53 such as NO<sub>x</sub> and SO<sub>x</sub>, but also exhibit characteristics such as a small footprint and higher energy  
54 efficiencies[8]. Especially in the context of achieving carbon neutrality, the use of pyrolysis and  
55 gasification technology for waste processing becomes particularly vital in reducing the pressure of  
56 carbon emissions[9, 10]. In the application of the pyrolysis and gasification technology, the  
57 gasification technology has a better advantage in resource utilization[11-14]. In the gasification  
58 process, as shown in the [Fig.1](#), the combustible solid waste of MSW is introduced into the gasifier.  
59 Under a semi-reductive atmosphere, the MSW undergoes thermal decomposition, resulting in the  
60 conversion of a substantial portion of the waste into high-value syngas primarily composed of CO,  
61 H<sub>2</sub>, CO<sub>2</sub> and CH<sub>4</sub> with conversion rates as high as 70% [15]. Additionally, a variable fraction of  
62 char or ash is generated. The ash and char residues are transported to specialized factories for  
63 comprehensive utilization, including applications such as soil remediation[16] and building  
64 materials[17]. The high temperature syngas can directly be utilized in gas turbines or boilers for  
65 power generation or heating purposes[18]. Alternatively, it can be directed to industries for the  
66 production of high-value by-products[19], showcasing the significant utilization potential of  
67 syngas[20, 21].

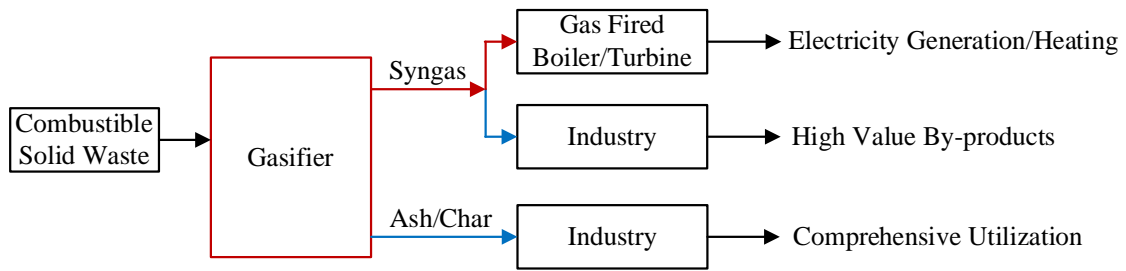


Fig.1 Comprehensive utilization diagram of MSC gasification technology

68

69

70 However, due to the large amount of chlorine-containing substances in MSW[22], those

71 substance are primarily released in the form of HCl gas during gasification process, resulting in

72 the inclusion of significant amounts of HCl gas in the syngas[23]. Consequently, the presence of

73 HCl in syngas can impose limitations and increase application costs. Firstly, the HCl can react

74 with alkali metal, Na and K even at high temperatures, to generate chlorides; these chlorides can

75 cause high temperature corrosion, fouling and slagging of the gasifier and related devices[24, 25].

76 Secondly, HCl can accelerate migration of heavy metals such as mercury and the formation of

77 dioxins, thereby causing environmental damage and endangering human health[26]. Finally, the

78 high HCl content in syngas results in by-products with high acidity, thereby restricting their

79 quality and application. Those issues pose some serious challenge towards downstream syngas

80 utilization. Therefore, it is necessary to remove HCl prior to application. Currently, conventional

81 methods for HCl removal involve spraying Ca/Na-containing substances into the flue gas or using

82 active carbon to adsorb HCl at low temperatures[27], which have shown favorable results[28-30].

83 However, these methods have certain disadvantages, particularly when it comes to HCl removal at

84 low temperatures. When applied to remove HCl from syngas, these technologies can result in heat

85 loss in syngas and inefficient heat utilization. Furthermore, the low-temperature HCl removal

86 process can lead to secondary generation of pollutants and partially contaminated waste. Therefore,

87 it is crucial to address these issues by reducing HCl gas using a suitable adsorbent at medium to



88 high temperatures (300 – 700 °C). This step is essential for enhancing the quality of syngas and  
89 enabling further applications.

90 Hydrotalcite is a kind of metal oxide compound with a layered porous structure and double  
91 hydroxyl groups, which exhibits acid-base bifunctionality and thermal stability. Since its discovery,  
92 it has been extensively used as adsorbent, catalyst and catalyst carrier[31-33]. Researchers have  
93 also explored the modification of hydrotalcite by incorporating metals with different properties,  
94 resulting in hydrotalcite-like compound. These modified hydrotalcite materials demonstrate  
95 enhanced performances to satisfy the specific application requirements[34, 35]. Hydrotalcite and  
96 hydrotalcite-like compounds have demonstrated effective adsorption performance for Cl<sup>-</sup> in an  
97 aqueous environment [36, 37], as well as HCl from flue gases[38, 39]. Previous studies have  
98 specifically investigated the effects of hydrotalcite preparation conditions and operating  
99 parameters on the performance of HCl removal using simplified HCl gas simulations on the  
100 performance of HCl removal using simulated HCl gas atmospheres[40]. In these studies, it was  
101 observed that the hydrotalcite-like compound modified with Ca<sup>2+</sup> (Ca-Mg-Al-CO<sub>3</sub>) exhibited  
102 superior HCl removal performance. Additionally, Ca-Mg-Al-CO<sub>3</sub> can be utilized at higher  
103 temperatures[41]. Therefore, in this current study, we chosen Ca-based nitrate to modify Mg-Al-  
104 CO<sub>3</sub>, and both Ca-Mg-Al-CO<sub>3</sub> and Mg-Al-CO<sub>3</sub> were selected to remove HCl as a comparison. It is  
105 important to note that CO<sub>2</sub>, a non-combustible gas, is present in syngas at significantly higher  
106 concentrations than HCl. CO<sub>2</sub> is a key component in various processes, including the water-gas  
107 shift reaction, producing CO and so on , It has been established that hydrotalcite can adsorb CO<sub>2</sub>  
108 effectively.[42, 43] CO<sub>2</sub> is also confirmed to be adsorbed by hydrotalcite [44, 45]. Therefore, both  
109 HCl and CO<sub>2</sub> can be adsorbed by hydrotalcite. Especially in the case of Ca-Mg-Al-CO<sub>3</sub>, it exhibits

110 remarkable capabilities in adsorbing CO<sub>2</sub> due to the generation of calcium-containing oxide at  
111 mid-high temperatures[46, 47]. Consequently, the presence of CO<sub>2</sub> may influence the performance  
112 of HCl removal using hydrotalcite and Ca-Mg-Al-CO<sub>3</sub>, although the specific effect tendency is not  
113 yet well understood. Although some studies have reported on the effect of CO<sub>2</sub> on HCl removal  
114 using other calcium-based compounds such as CaO, Ca(OH)<sub>2</sub> and CaCO<sub>3</sub>[48, 49], it should be  
115 noted that the research objectives and operating conditions varied, resulting in inconsistent results.  
116 Therefore, when studying the performance of HCl removal in syngas, the effect of CO<sub>2</sub> on HCl  
117 removal using hydrotalcite and hydrotalcite-like compounds should not be overlooked.  
118 Surprisingly, there is a lack of research focusing on the characteristics of HCl removal using  
119 hydrotalcite and hydrotalcite-like compound in the presence of CO<sub>2</sub>. Therefore, it is necessary to  
120 analyze the effect and mechanism of CO<sub>2</sub> on HCl removal by adsorbent. Obtaining the impact  
121 tendency of CO<sub>2</sub> on HCl removal by hydrotalcite is crucial for ensuring the production of high-  
122 quality syngas and expanding the applications of syngas in various fields.

123 To obtain high-performance adsorbents and maximize syngas resource utilization, this study  
124 focuses on the utilization of Mg-Al-CO<sub>3</sub> and modified Ca-based hydrotalcite-like (Ca-Mg-Al-CO<sub>3</sub>)  
125 for HCl removal at mid-high temperatures, to analyze of the effect of Ca<sup>2+</sup> modification on the  
126 HCl removal performance of hydrotalcite Firstly, the Mg-Al-CO<sub>3</sub> and Ca-Mg-Al-CO<sub>3</sub> were  
127 synthesized using the coprecipitation method. Subsequently, the impact of CO<sub>2</sub> on HCl removal by  
128 Mg-Al-CO<sub>3</sub> and Ca-Mg-Al-CO<sub>3</sub> was investigated under different conditions, with a comparative  
129 analysis. To further understand the adsorption relationship of CO<sub>2</sub> and HCl by Ca-Mg-Al-CO<sub>3</sub>,  
130 thermogravimetric analysis (TGA) was employed. The characterization instruments of XRD and  
131 the specific surface area and pore size were used to analysis the reaction mechanism. Finally, the

132 adsorption kinetics were chosen to investigate the adsorption mechanism of HCl removal using  
133 Ca-Mg-Al-CO<sub>3</sub> in the presence of CO<sub>2</sub>.

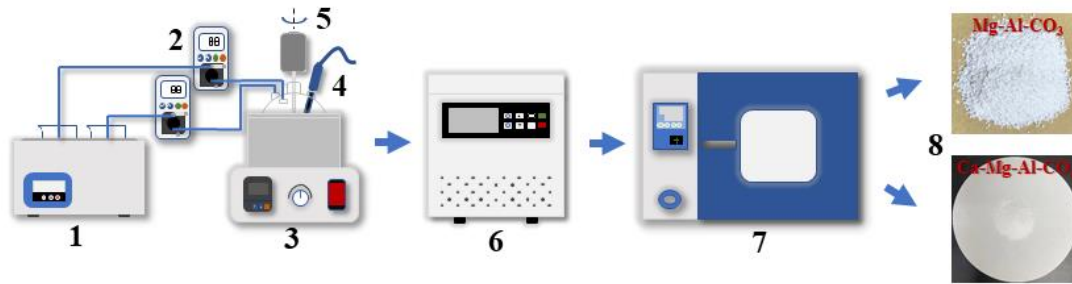
## 134 2. Material and Methods

### 135 2.1 Synthetic steps

136 In this study, the chemical reagents of Mg (NO<sub>3</sub>)<sub>2</sub>·6H<sub>2</sub>O, Al (NO<sub>3</sub>)<sub>3</sub>·9H<sub>2</sub>O, Ca (NO<sub>3</sub>)<sub>2</sub>·4H<sub>2</sub>O,  
137 NaOH and Na<sub>2</sub>CO<sub>3</sub> with analytically pure grade were produced from Sinopharm Chemical  
138 Reagent Co., Ltd. The Mg-Al hydrotalcite containing CO<sub>3</sub><sup>2-</sup> intercalated (Mg-Al-CO<sub>3</sub>), and Ca-  
139 based modified hydrotalcite-like (Ca-Mg-Al-CO<sub>3</sub>) were synthesized using the coprecipitation  
140 method for gaseous HCl removal. The synthetic process is shown in Fig. 2 and the synthetic steps  
141 are as follows.

142 Preparation of Mg-Al-CO<sub>3</sub>: (1) According to the molar ratio of Mg/Al=3, [OH<sup>-</sup>]=2([Mg<sup>2+</sup>]  
143 +[Al<sup>3+</sup>]) and [CO<sub>3</sub><sup>2-</sup>]=1/2[Al<sup>3+</sup>], the corresponding chemical reagents were weighed. (2) The  
144 nitrates and two bases were dissolved in ultra-pure water at 60 °C, respectively. (3) The two  
145 solutions were dropped into a four-port flask at a certain speed using a peristaltic pump at 60 °C.  
146 During this process, the pH value was maintained at 11, and the stirring speed was set at 300 r/min.  
147 The resulting mixture was continuously stirred for 2 hours at 60 °C. (4) The obtained substance  
148 was subjected to crystal growth for 20 hours at 80 °C. (5) The solid part was washed using ultra-  
149 pure water until neutral, dried and grounded. The Mg-Al-CO<sub>3</sub> was obtained, and the particles in  
150 size range of 40 to 60 mesh were chosen for subsequent experiments.

151 Preparation of Ca-Mg-Al-CO<sub>3</sub>: The Ca-Mg-Al-CO<sub>3</sub> was prepared with a molar ratio of  
152 Ca<sup>2+</sup>/Mg<sup>2+</sup>=(Ca+Mg)<sup>2+</sup>/Al<sup>3+</sup>=3, [OH<sup>-</sup>]=2([(Ca+Mg)<sup>2+</sup>]+[Al<sup>3+</sup>]) and 2[CO<sub>3</sub><sup>2-</sup>]=[Al<sup>3+</sup>]. The step (2)  
153 to (5) mentioned above were repeated to obtain Ca-Mg-Al-CO<sub>3</sub>, and the particle sizes range was  
154 the same as that of Mg-Al-CO<sub>3</sub>.



155

156 1. Thermostatic water pot; 2. Peristaltic pump; 3 Thermostatic oil pot; 4. pH meter; 5. Mixer; 6.

157

Centrifuge; 7. Drying oven; 8. Adsorbents

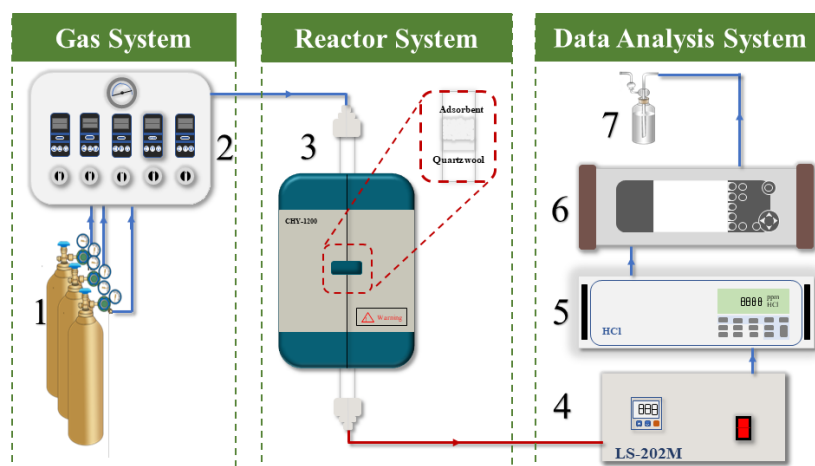
158

Fig. 2 Synthetic process of adsorbents

## 159 2.2 Characterization

160 X-ray diffraction (XRD) analysis was made using the Bruker D8 ADVANCE equipment by  
 161 Bruker Daltonics Inc., Germany. The diffraction angle range was set from 5 to 80 ° with a step  
 162 size of 0.08 °. The equipment model of scanning electron microscope (SEM) was S4800 from  
 163 Hitachi Limited. Energy Dispersive Spectroscopy (EDS) equipment from EDAX. Inc was  
 164 employed for elemental analysis. The specific surface area and pore size analyzer (ASAP 2460,  
 165 Micromeritics Instruments Corporation, America) was used to analyze the specific surface area  
 166 and pore structure of adsorbents. Thermogravimetric analysis (TGA) for thermal stability analysis  
 167 was carried out using the (PerkinElmer TA 8000) produced by platinum Elmer Co., Ltd.

## 168 2.3 Experimental equipment and steps



169

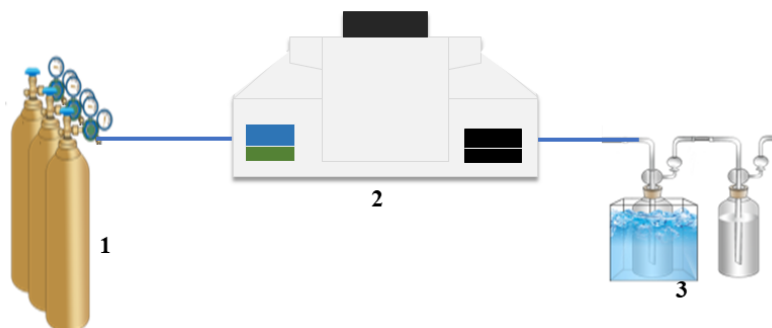
170 1. Gas cylinder; 2. Mass flowmeter; 3. Reactor; 4. LJ-202M condenser; 5. HCl analyzer; 6. Flue

171

gas analyzer; 7. Gas-washing bottle

172

A. The schematic diagram of experiments



173

1. Gas cylinder; 2. Thermogravimetric analyzer; 3. Gas-washing bottle

174

175

B. Experimental schematic diagram of thermogravimetry analysis (TGA)

176

Fig.3 Experimental schematic diagram

177

The HCl (Signal Group Ltd., UK) and the CO<sub>2</sub> (MRU Instruments, Inc., Germany) analyzers warmed up and calibrated. Before conducting experiments, a blank test was performed involving both CO<sub>2</sub> and HCl to ensure that there was no mutual interference. Then the absorbent was placed in the middle of the reactor, the reactor temperature was gradually increased to desired reaction temperature at a heating rate of 10 °C/min in a N<sub>2</sub> atmosphere. Once the reaction temperature reached a steady state, a mixture of N<sub>2</sub>, HCl and CO<sub>2</sub> in specific proportions was introduced into the reactor using mass flowmeters. Simultaneously, the analyzers started monitoring the gas composition. To prevent the release of HCl and CO<sub>2</sub> into the environment, an aqueous solution of 0.3 mol/L NaOH was used to adsorb the excess HCl and CO<sub>2</sub>. The specific operating conditions of the experiments are listed in Table.1. High-purity N<sub>2</sub> and CO<sub>2</sub> with a grade of 99.999 % grade and HCl concentration of 2000 ppmv from Nanjing Special Gas Plant Co., Ltd. were employed in the tests.

189

Table. 1 The operating conditions of tests

Parameters	Unit	HCl Removal	TGA
Material	/	Mg-Al-CO <sub>3</sub> , Ca-Mg-Al-CO <sub>3</sub>	Ca-Mg-Al-CO <sub>3</sub>
Reaction temperature	°C	300, 400, 500, 600, 700	500
HCl concentration	ppmv	400	400
CO <sub>2</sub> concentration	vol. %	0, 10, 15, 20, 25	20

Quality	mg	500	20
Gas flow	L/min	0.5	0.1
Reaction time	h	/	5
Particle diameter	nm		0.25-0.38
Gas	/		HCl, N <sub>2</sub> , CO <sub>2</sub>

190

## 191 2.4 Experimental data processing

192 The instantaneous HCl removal rate is calculated by Eq. 1.

$$193 \quad \eta = \frac{C_{in} - C_{out}}{C_{in}} \times 100\% \quad \text{Eq. 1}$$

194 The HCl average removal rate is calculated by Eq. 2.

$$195 \quad \bar{\eta} = \frac{1}{N} \sum_{i=1}^N \eta_i \quad \text{Eq. 2}$$

196 The breakthrough chlorine capacity refers to the total mass of HCl adsorbed per unit mass of  
 197 adsorbent from the start of the reaction until complete breakthrough occurs. The breakthrough  
 198 chlorine capacity is calculated by Eq. 3[50].

$$199 \quad Q_t = 10^{-3} \int_0^t \frac{36.5(C_0 - C_{out})V_s}{22.4m} dt \quad \text{Eq. 3}$$

200 Where  $\eta$  and  $\eta_i$ ,  $\bar{\eta}$  are instantaneous HCl removal rate, HCl removal rate at  $t$  time and HCl  
 201 average removal rate, %,  $i=1, 2, 3, \dots, N$ .  $C_0$ ,  $C_{in}$  and  $C_{out}$  are HCl concentration of initial, inlet,  
 202 and outlet, ppmv.  $Q_t$  is the saturation chlorine content, mg/g.  $V_s$  is volume flow of reaction gas,  
 203 L/min.  $m$  is the quality of adsorbent, g.

204 The Intra-particle diffusion model is commonly used to assess whether diffusion is the sole rate-  
 205 controlling step in a reaction. The equation is given by Eq. 4[51].

$$206 \quad q_t = k_1 t^{1/2} + C \quad \text{Eq. 4}$$

207 Where  $q_t$  is the adsorption capacity of adsorbent at time  $t$ , mg·g<sup>-1</sup>.  $k_1$  is the rate constant of  
 208 intragranular diffusion model, mg·(g·min<sup>-1/2</sup>)<sup>-1</sup>.  $C$  is constant, mg·g<sup>-1</sup>.

209 The pseudo-first-order model (PFO) is employed to determine whether the adsorption process is  
 210 controlled by diffusion or surface reaction, when adsorption is controlled by chemical adsorption  
 211 of chemical factors, the equation is described by Eq. 5[52].

212 
$$\log(q_e - q_t) = \log q_e - \log\left(\frac{k_1}{2.303}\right) \quad \text{Eq. 5}$$

213 Where  $q_e$  represents HCl equilibrium adsorption amount,  $\text{mg}\cdot\text{g}^{-1}$ .  $q_t$  is the HCl adsorption  
214 capacity at time  $t$ .  $k_1$  is the rate constant,  $\text{min}^{-1}$ .

215 The pseudo-second-order (PSO) model is predicated on the assumption that the adsorption rate  
216 is controlled by a chemical adsorption mechanism, involving electron sharing or electron transfer  
217 between the adsorbent and the adsorbate. The equation for this model is represented as Eq. 6[53].

218 
$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{1}{q_e} \quad \text{Eq. 6}$$

219 Where  $q_e$  represents HCl equilibrium adsorption capacity,  $\text{mg}\cdot\text{g}^{-1}$ .  $q_t$  is the HCl adsorption  
220 capacity at time  $t$ .  $k_2$  is the rate constant,  $\text{min}^{-1}$ .

221 The Elovich model is indeed a commonly used model for describing the chemical adsorption of  
222 gases on solid surfaces. The Elovich model is defined as Eq.7[54].

223 
$$q_t = \left(\frac{2.3}{k_0}\right) \lg(t + t_0) - \left(\frac{2.3}{k_0}\right) \lg t_0 \quad \text{Eq. 7}$$

224 Where  $t_0 = 1/k_0 k_f$ .  $k_0$  is initial adsorption rate at  $q_t = 0$ ,  $\text{mg}/\text{g}\cdot\text{min}$ .  $k_f$  is desorption constant,  $\text{g}\cdot\text{mg}$ .  
225 Eq. 7 is simplified with  $t k_0 k_f \ll 1$ ,  $t=0$ ,  $q_t=0$ .

226 
$$q_t = k_0 \ln(k_f k_0) + k_0 \ln(t) \quad \text{Eq. 8}$$

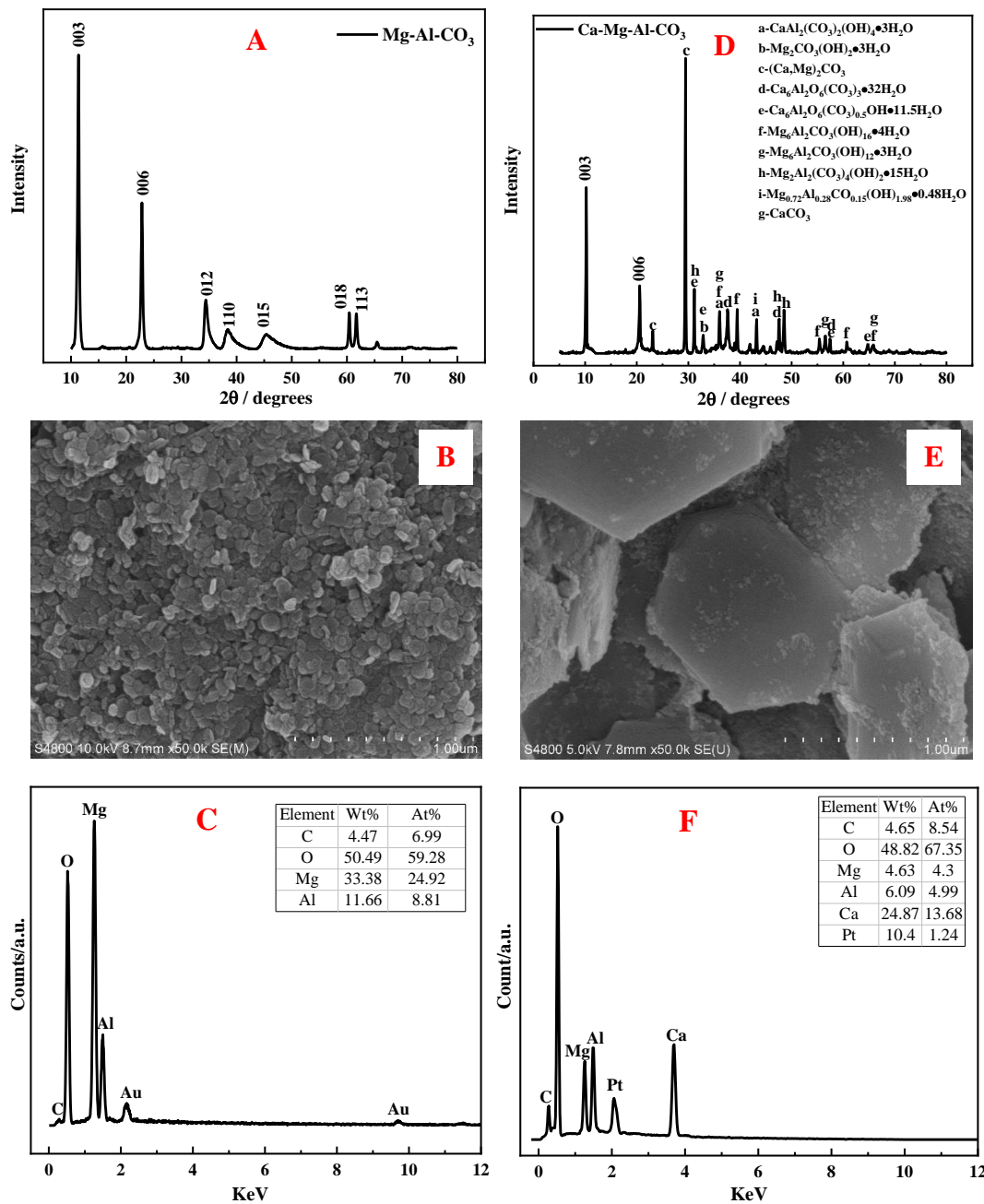
227 The Bangham model is used to describe pore diffusion, which optimizes Lagergren model  
228 through time compensation, so that it has been well applied in predicting the adsorption process of  
229 various adsorbents. The model function is shown in Eq. 9[55].

230 
$$\lg\{\lg[q_e / (q_e - q_t)]\} = \lg\left(\frac{k_2}{2.303}\right) + n \lg t \quad \text{Eq. 9}$$

231 Where  $q_t$  is adsorption capacity of adsorbent at time  $t$ ,  $\text{mg}\cdot\text{g}^{-1}$ .  $q_e$  is HCl equilibrium adsorption  
232 capacity,  $\text{g}\cdot\text{min}$ .  $k_1$  is rate constant of intragranular diffusion model,  $\text{mg}\cdot(\text{g}\cdot\text{min}^{-1/2})^{-1}$ .  $C$  is constant,  
233  $\text{mg}\cdot\text{g}^{-1}$ .  $k_2$  is the Bangham rate constant.

### 234 3. Results and Discussion

#### 235 3.1 Characterization of adsorbents



236

237

238

239 Fig. 4 Characterization of adsorbents (Mg-Al-CO<sub>3</sub>: A: XRD pattern, B: SEM image, C: EDS

240 pattern; Ca-Mg-Al-CO<sub>3</sub>: D: XRD pattern, E: SEM image, F: EDS pattern)

241 The characterization results of Mg-Al-CO<sub>3</sub> and Ca-Mg-Al-CO<sub>3</sub> are presented in Fig. 4. From

242 Fig.4 A, B and C, it can be observed that the molar ratio of Mg/Al was 2.8, the characteristic

243 diffraction peaks of hydrotalcite (PDF- # 35-0965) were evident in Mg-Al-CO<sub>3</sub>, including crystal

244 faces of hydrotalcite such as (003), (006), (015), (018), (110), and (113). The layer spacing



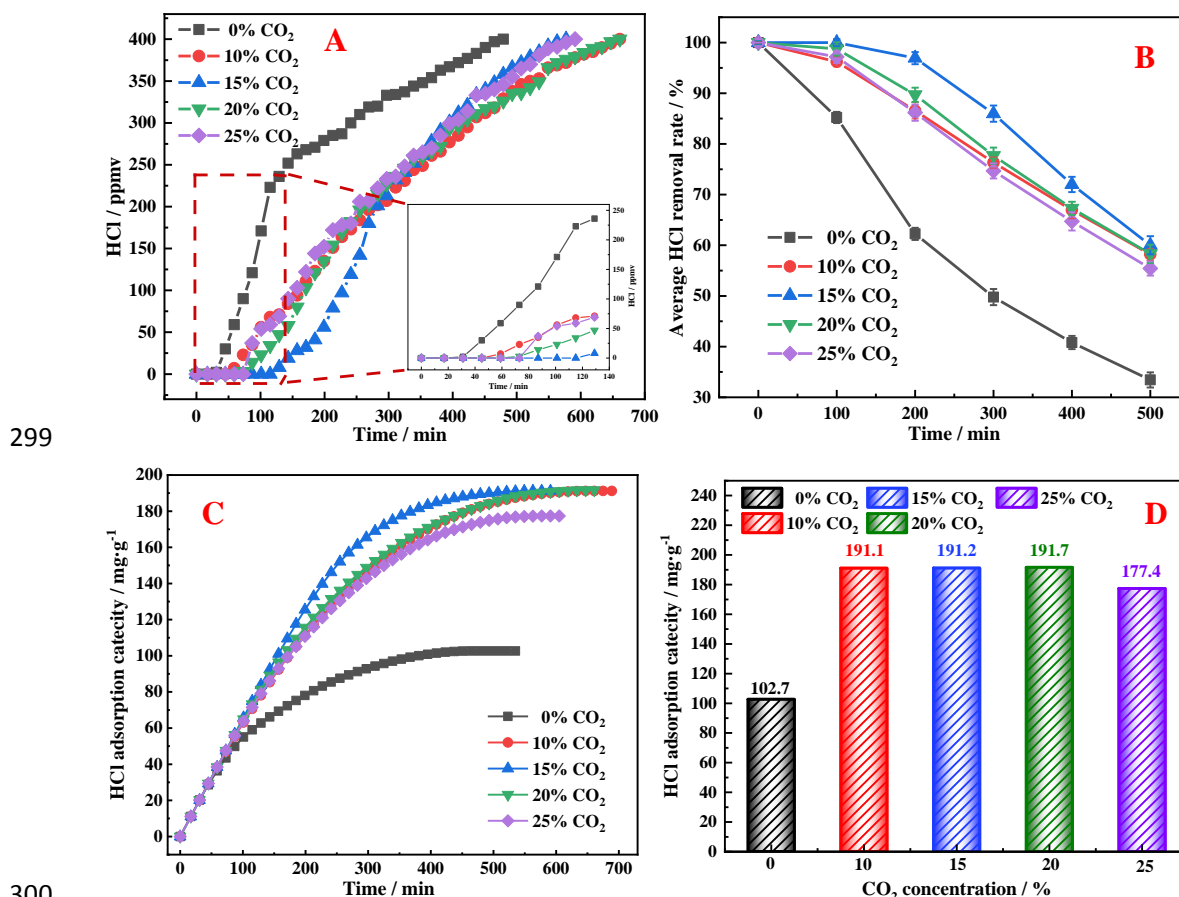
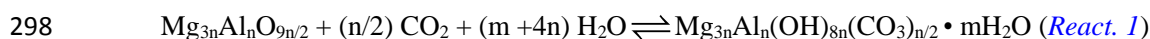
245 of  $d_{003}$  of Mg-Al-CO<sub>3</sub> was determined to be 0.78 nm. The SEM image revealed that Mg-Al-CO<sub>3</sub>  
246 exhibited a regular hexagonal lamellar structure, the width and thickness of layered Mg-Al-CO<sub>3</sub>  
247 were approximately around 400 nm and 50nm, respectively. Fig. 4 D, E, and F present the XRD  
248 pattern, SEM image and EDS pattern of Ca-Mg-Al-CO<sub>3</sub>, respectively. Ca-Mg-Al-CO<sub>3</sub> showed  
249 obvious characteristic diffraction peaks of (003) and (006) with a narrow peak pattern and high  
250 intensity, the layer spacing of  $d_{003}$  in Ca-Mg-Al-CO<sub>3</sub> was 1.02 nm. The SEM image in Fig. 4E  
251 confirmed the presence of a hexagonal lamellar structure in Ca-Mg-Al-CO<sub>3</sub>. The molar rate of  
252 Ca/Mg was determined to be 3.18, and the ratio of  $(Ca+Mg)^{2+}/Al^{3+}$  was 3.6. These results indicate  
253 the successful synthesis of the layered structure of Mg-Al-CO<sub>3</sub> and Ca-Mg-Al-CO<sub>3</sub>. A comparison  
254 between the characterization results of Mg-Al-CO<sub>3</sub> and Ca-Mg-Al-CO<sub>3</sub> suggests that the prepared  
255 Mg-Al-CO<sub>3</sub> exhibits a single crystal phase with good crystallinity. The structure size of Ca-Mg-  
256 Al-CO<sub>3</sub> was larger than that of Mg-Al-CO<sub>3</sub> from Fig. B and E. However, Ca-Mg-Al-CO<sub>3</sub> had  
257 some impurities, as evidenced by the presence of characteristic peaks corresponding to CaCO<sub>3</sub>  
258 found in the XRD pattern, this suggests the occurrence of a small amount of CaCO<sub>3</sub> impurities in  
259 Ca-Mg-Al-CO<sub>3</sub>. The main reason for this phenomenon can be attributed to the difference in ion  
260 radius that the ion radius of Ca<sup>2+</sup>, Mg<sup>2+</sup> and Al<sup>3+</sup> are 0.100nm, 0.072nm and 0.0535nm, respectively,  
261 the ion radius of Ca<sup>2+</sup> is larger than that of Mg<sup>2+</sup> and Al<sup>3+</sup>. The presence of Ca<sup>2+</sup> as a substitute for  
262 some Mg<sup>2+</sup> in the formation of Ca-Mg-Al-CO<sub>3</sub> resulted in incompatibility, leading to the  
263 formation of CaCO<sub>3</sub>. Additionally, the low solubility product constant of CaCO<sub>3</sub> contributed to its  
264 formation during the preparation of Ca-Mg-Al-CO<sub>3</sub>. These results showed that the structure of  
265 hydrotalcite can be modified by the addition of Ca<sup>2+</sup>, resulting in changes in its characteristics and  
266 the formation of CaCO<sub>3</sub> impurities.

## 267 3.2 The performance of HCl removal by Mg-Al-CO<sub>3</sub>

### 268 3.2.1 The effect of CO<sub>2</sub> concentration on HCl removal

269 When there is HCl gas in the outlet gas, the corresponding point on the breakthrough curve is  
270 referred to as the breakthrough point, and the time at this point is known as the breakthrough time.  
271 In Fig.5, the breakthrough curve, average HCl removal rate, and adsorption chlorine capacity of  
272 Mg-Al-CO<sub>3</sub> are displayed for an HCl concentration of 400 ppmv, a flow rate of 500 mL/min, a  
273 reaction temperature of 300 °C, and CO<sub>2</sub> concentrations ranging from 0 to 25%. As shown in Fig.  
274 5A, the breakthrough curves initially shift towards left and then towards the right with increasing  
275 CO<sub>2</sub> concentration. The Mg-Al-CO<sub>3</sub> adsorbent exhibits the highest adsorption capacity when the  
276 CO<sub>2</sub> concentration is at 15 % during a period of 300 min. The breakthrough time increased from  
277 approximately 17 min to 115 min, and then decreased from approximately 115 min to 73 min.  
278 Compared to the adsorption effect without CO<sub>2</sub>, the adsorption capacity has increased by at least  
279 329.4% with the presence of CO<sub>2</sub>, indicating that CO<sub>2</sub> has a positive effect on improving the  
280 adsorption capacity of Mg-Al-CO<sub>3</sub>. The average HCl removal rates are presented in Fig 5B. It can  
281 be observed that the average HCl removal rates decreased rapidly with the increase of reaction  
282 time. When the CO<sub>2</sub> concentration rises from 0 to 15%, the average HCl removal rate increased  
283 from 85.2 % to 100 % when the CO<sub>2</sub> concentration increased from 0 to 15 %, showing an increase  
284 of 17.4%. Subsequently, it decreased from 100 % to 97.2 % over a period of 100 mins. As the  
285 reaction time extended from 100 to 500 mins, the HCl removal rates decreased by 67.6 %, 51.8 %,  
286 40 %, 51.7 %, and 54.6 %, respectively, as the CO<sub>2</sub> concentration increased from 0 to 25 %.  
287 Comparing the HCl removal effect with and without CO<sub>2</sub>, it can be concluded that the  
288 performance of HCl removal was enhanced when the simulation gas contained CO<sub>2</sub>. This main

289 reason for this improvement is that hydrotalcite exhibits a memory function, enabling it to  
 290 regenerate under appropriate conditions when the structure is not completely destroyed[50], this  
 291 process was employed as depicted in *React. 1*. Therefore, the structure of Mg-Al-CO<sub>3</sub> undergoes  
 292 regeneration when CO<sub>2</sub> participated in the reaction at 300 °C. Mg<sub>3n</sub>Al<sub>n</sub>O<sub>9n/2</sub>, which was difficult to  
 293 react with HCl, participated in the reaction again, resulting in the addition of the active sites on the  
 294 surface of the adsorbent. The structure regeneration of Mg-Al-CO<sub>3</sub> leads to the replacement of  
 295 more CO<sub>3</sub><sup>2-</sup> for Cl<sup>-</sup>, and further enhancing HCl removal performance of the adsorbents.  
 296 Consequently, the effect of HCl removal increases with the increase of CO<sub>2</sub> concentration. And the  
 297 regeneration process was optimal with 15% CO<sub>2</sub>.



301 Fig. 5 The performance of HCl removal using Mg-Al-CO<sub>3</sub> at different CO<sub>2</sub> concentrations (A: the

302 breakthrough curves; B: the average HCl removal rate; C: the adsorption chlorine capacity curves;  
303 D: the maximum breakthrough chlorine capacity)

304 The chlorine adsorption capacity curves and the breakthrough chlorine capacity of Mg-Al-CO<sub>3</sub>  
305 are shown in Fig. 5C and D. In the presence of CO<sub>2</sub> (Fig. 5C), Mg-Al-CO<sub>3</sub> exhibited significantly  
306 higher chlorine adsorption capacity compared to the absence of CO<sub>2</sub>. Additionally, Fig. 5 D  
307 indicates an increase in the breakthrough chlorine capacity of approximately 86 % - 87 % with  
308 CO<sub>2</sub> concentrations ranging from 10 %-20 %, stabilizing at around 191 mg·g<sup>-1</sup>. However, beyond  
309 a CO<sub>2</sub> concentration of 20%, the breakthrough chlorine capacity started to decrease, indicating a  
310 decline in the HCl removal performance of Mg-Al-CO<sub>3</sub>. This decline can be attributed to the high  
311 CO<sub>2</sub> content, which increased the diffusion resistance of HCl, impeding its reaction with the  
312 adsorbent.

### 313 3.2.2 The effect of reaction temperature on HCl removal

314 The breakthrough curves, average HCl removal rates and the adsorption chlorine capacity of  
315 Mg-Al-CO<sub>3</sub> are depicted in Fig. 6. The experiments were conducted at an HCl concentration of  
316 400 ppmv, a flow rate of 500 mL/min, a CO<sub>2</sub> concentration of 20 %, and a reaction temperature  
317 ranging from 300 to 700 °C with intervals of 100 °C. In Fig. 6A, as the temperature increased, the  
318 breakthrough curves shifted to left, indicating a decrease in the effective HCl adsorption time from  
319 650 min to 130min. The breakthrough time initially increased from 59 min to 73 min as the  
320 reaction temperature increased from 300 to 500 °C, but then decreased from 73 min to 31 min.  
321 The adsorption capacities at 400 °C and 500 °C were better than that at 300 °C, occurring at around  
322 170 min and 100 min, respectively. The optimal performance of HCl removal using Mg-Al-CO<sub>3</sub>  
323 was observed at 400 °C within the first 200 min. However, although the adsorption capacity

324 increased within the first 200 min at 400 °C, the rate of HCl adsorption showed a declining trend.

325 The average HCl removal rates of Mg-Al-CO<sub>3</sub> are presented in Fig. 6B, the average HCl removal

326 rates of Mg-Al-CO<sub>3</sub> at 400 and 500 °C were higher than that at 300 °C in the first 100 min, and the

327 average HCl removal rate of Mg-Al-CO<sub>3</sub> at 400 °C surpassed that at 300 °C within the first 200 min,

328 reaching 90.7%. This improvement can be attributed to the creation of the new porous structure

329 resulting from the release of moisture and CO<sub>2</sub> during the thermal decomposition of Mg-Al-

330 CO<sub>3</sub>[31]. The abundant pore structure can provide more active sites, facilitating the reaction

331 between the absorbent with HCl. With the increase in reaction temperature, the pore structure

332 becomes more pronounced. However, as the reaction temperature exceeded the pore structure of

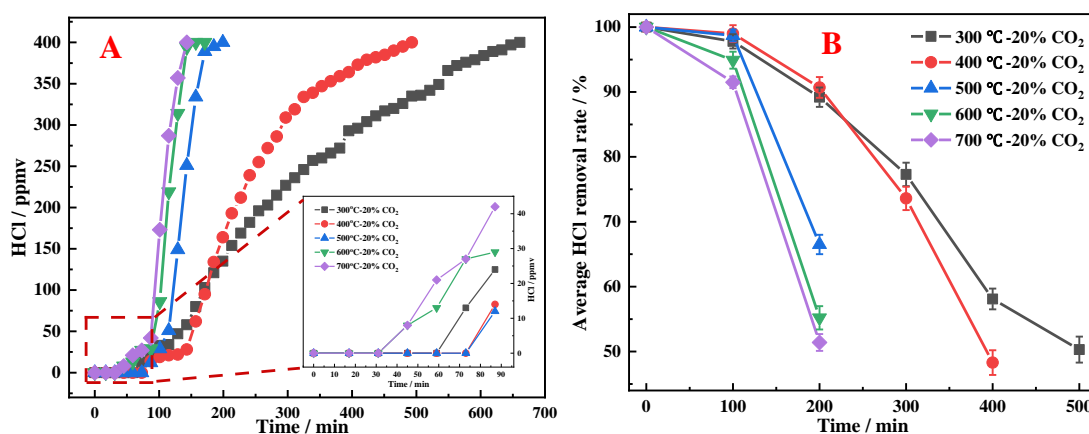
333 Mg-Al-CO<sub>3</sub> was disrupted, leading to the formation of spinel and a loss of regenerative capacity.

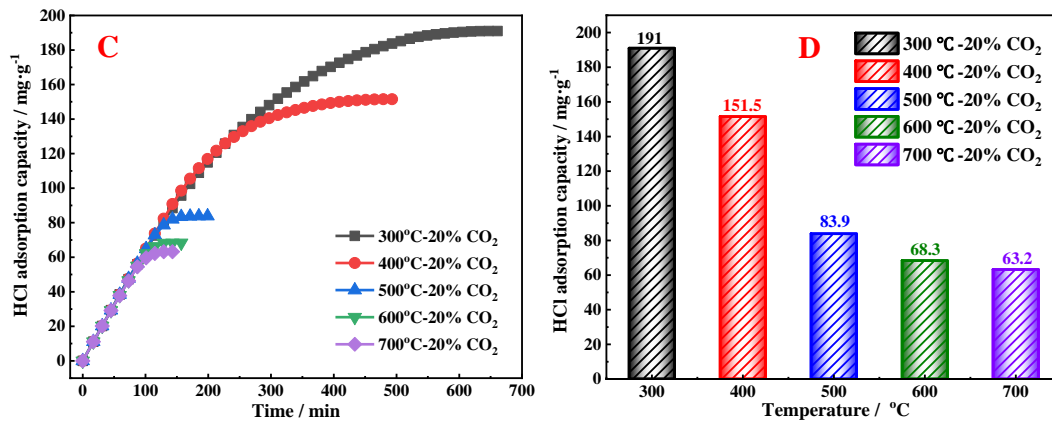
334 The magnesium aluminum oxide generated in high-temperature was not conducive to the reaction

335 of HCl removal. Consequently, the HCl adsorption capacity decreased and the HCl removal rates

336 diminished when the reaction temperature surpassed 400 °C.

337





338

339 Fig. 6 The performance of HCl removal using Mg-Al-CO<sub>3</sub> at different temperatures (A: the  
 340 breakthrough curves; B: the average HCl removal rate; C: the adsorption chlorine capacity curves;  
 341 D: the maximum breakthrough chlorine capacity)

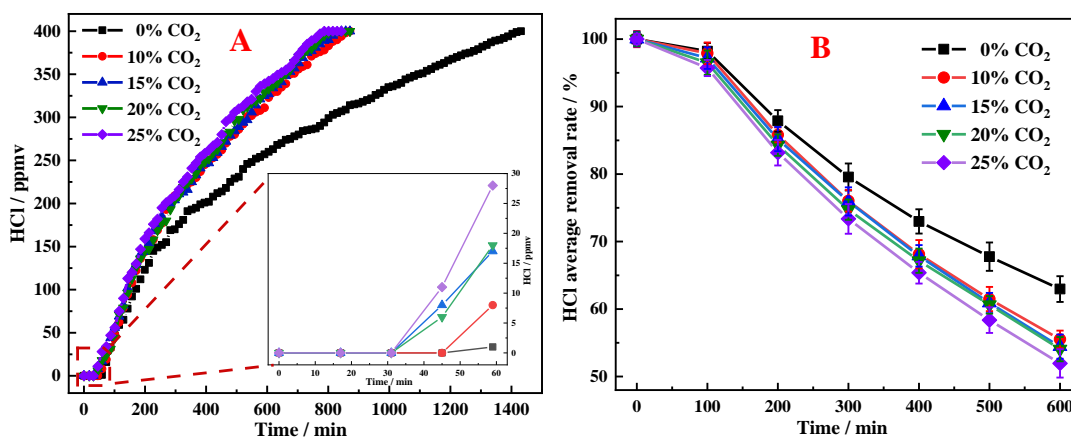
342 In Fig. 6C, the adsorption chlorine capacity of Mg-Al-CO<sub>3</sub> exhibits trends consistent with the  
 343 breakthrough curves. Fig. 6B shows the HCl removal rates, which demonstrated good  
 344 performance at 400 °C in first 200 min. However, the chlorine adsorption capacity decreased from  
 345 191 to 63.2 mg·g<sup>-1</sup> as the reaction temperature increased from 300 to 700°C, resulting in a decrease  
 346 of 66.9 %. This reduction can be attributed to the decrease in effective reactive substances as the  
 347 reaction temperature increased. These results indicate that while CO<sub>2</sub> enhanced the HCl removal  
 348 performance of Mg-Al-CO<sub>3</sub>, temperature played a more significant role in determining the HCl  
 349 removal performance of Mg-Al-CO<sub>3</sub>. Additionally, and the complete structure of hydrotalcite also  
 350 contributed to the favorable HCl removal capacity.

### 351 3.3 The performance of HCl removal by Ca-Mg-Al-CO<sub>3</sub>

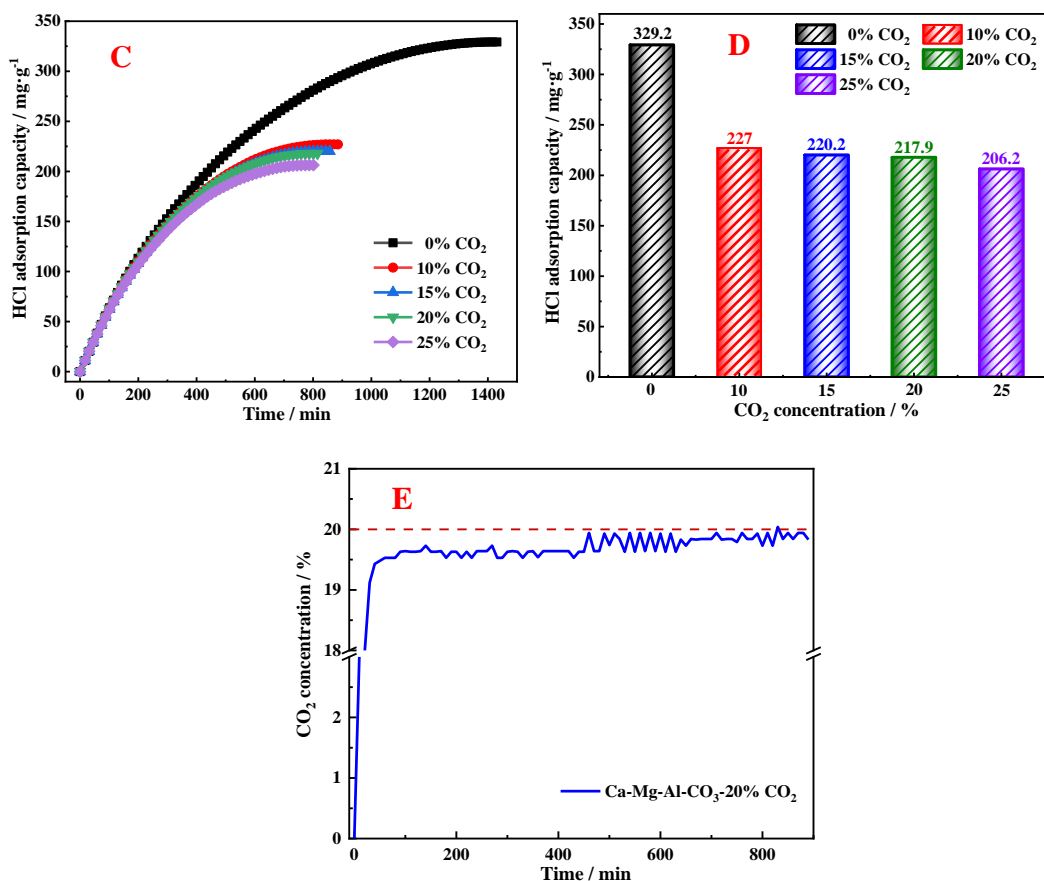
#### 352 3.3.1 The effect of CO<sub>2</sub> concentration on HCl removal performance

353 The effects of different CO<sub>2</sub> concentrations on HCl removal by Ca-Mg-Al-CO<sub>3</sub> were  
 354 investigated, and the results are presented in Fig. 7 A-D. The experiments were conducted at a  
 355 temperature of 300 °C, HCl concentration of 400 ppmv and CO<sub>2</sub> concentrations ranging from 0 to

356 25 %. The breakthrough curves of HCl adsorption by Ca-Mg-Al-CO<sub>3</sub> are shown in Fig.7A. The  
 357 reaction time of HCl removal by Ca-Mg-Al-CO<sub>3</sub> was 3 times that of Mg-Al-CO<sub>3</sub> without CO<sub>2</sub>. It  
 358 can be observed that the reaction time decreased from 1430 min to 783 min with the increase of  
 359 CO<sub>2</sub> concentrations from 0 to 25 %, decreased by 45.24 %, indicating that the adsorption  
 360 capacities of Ca-Mg-Al-CO<sub>3</sub> decreased with the increase of CO<sub>2</sub> concentration. Compared to the  
 361 optimal HCl removal performance of Mg-Al-CO<sub>3</sub>, the reaction time increased by more than 2  
 362 times. When CO<sub>2</sub> concentrations rose from 10 to 25 %, the adsorption time only decreased by  
 363 9.64 %, suggesting a minor impact on HCl removal by Ca-Mg-Al-CO<sub>3</sub> with a continuous increase  
 364 of CO<sub>2</sub> concentration. The breakthrough time also decreased from approximately 60 min to 45 min,  
 365 representing a decrease of 25 %. In Fig. 7B, the average HCl removal rates exhibited a decreasing  
 366 trend with the presence of CO<sub>2</sub>. Compared to the reaction without CO<sub>2</sub>, the reduction in average  
 367 HCl removal rates was more pronounced in the presence of CO<sub>2</sub> after 100 min. The average HCl  
 368 rates of adsorbent at 0 % and 25 % CO<sub>2</sub> were 98.2 % and 95.7 %, respectively, at 100 min,  
 369 indicating a decrease of only 2.5 %. However, a decrease of 11 % was observed at 500 min. The  
 370 phenomenon demonstrates that the presence of CO<sub>2</sub> has a negative effect on HCl removal, and this  
 371 effect becomes increasingly evident with higher CO<sub>2</sub> concentration.



372



373

374

375 Fig.7 The performance of HCl removal using Ca-Mg-Al-CO<sub>3</sub> under different CO<sub>2</sub> concentrations  
 376 and the change curve of CO<sub>2</sub> (A: the breakthrough curves; B: the average removal of HCl rate; C:  
 377 the adsorption chlorine capacity curves; D: the maximum breakthrough chlorine capacity; E: the  
 378 curve of CO<sub>2</sub> concentration during the HCl removal process)

379 Fig. 7C illustrates the breakthrough chlorine capacity of Ca-Mg-Al-CO<sub>3</sub> at different CO<sub>2</sub>  
 380 concentrations. It is evident from the figure that as the CO<sub>2</sub> concentration, the breakthrough  
 381 chlorine capacity decreased rapidly. This trend is further supported by Fig. 7D, which  
 382 demonstrates that the maximum breakthrough chlorine capacity of Ca-Mg-Al-CO<sub>3</sub> decreased from  
 383 329.2 mg/g to 227 mg/g when the CO<sub>2</sub> concentrations increased from 0 to 10 %, representing a  
 384 decrease of 37.4 %. However, the decrease was only 6.4 % when the CO<sub>2</sub> concentration increased  
 385 from 10 to 25 %. Fig. 7E shows the change in CO<sub>2</sub> concentration during the removal of HCl,

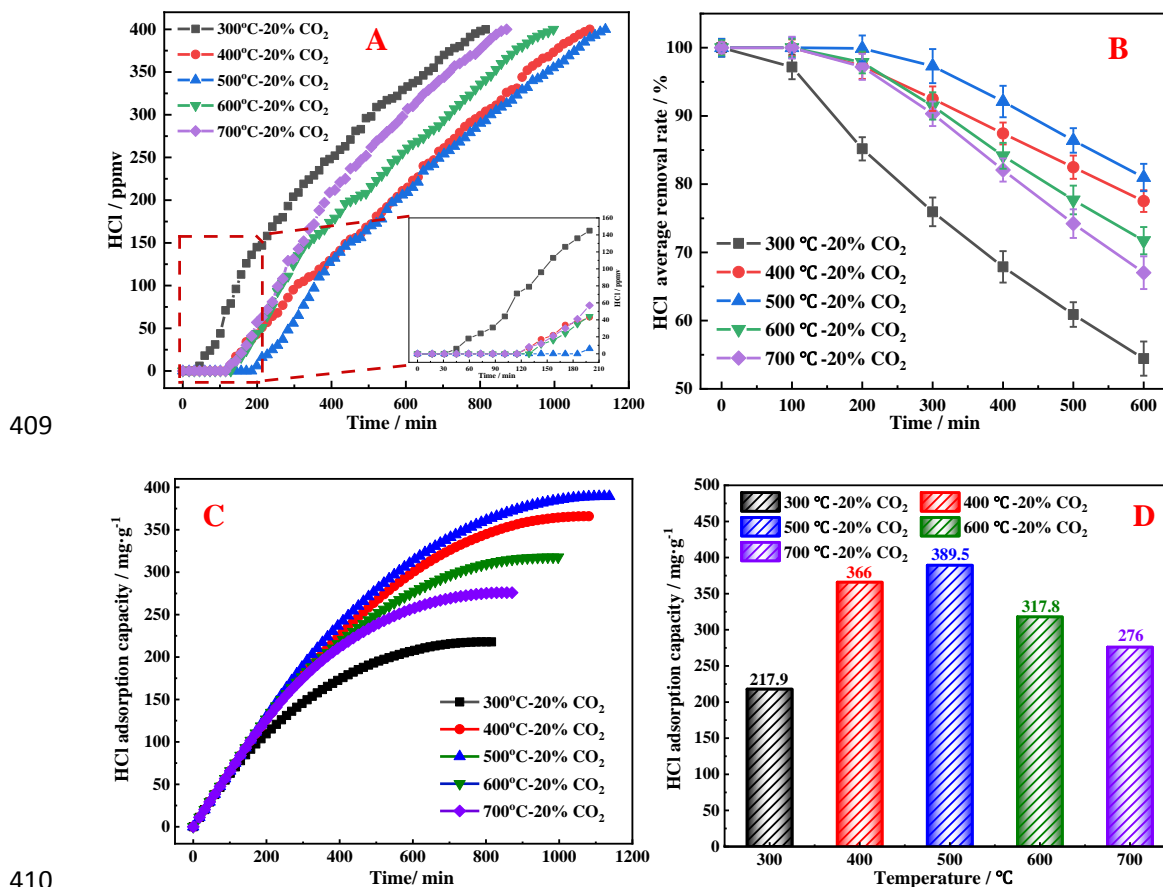


386 specifically focusing on the curve for 20 % CO<sub>2</sub>. It can be seen that the concentration of CO<sub>2</sub>  
387 exhibited a slow increase with increasing reaction time, indicating that CO<sub>2</sub> can also be adsorbed  
388 by Ca-Mg-Al-CO<sub>3</sub>. Especially within 450 min, the concentration of CO<sub>2</sub> remained around 19.65 %.  
389 The result implied that CO<sub>2</sub> also could be adsorbed by Ca-Mg-Al-CO<sub>3</sub> and reacted with partly Ca-  
390 oxide to generate CaCO<sub>3</sub>. The presence of CaCO<sub>3</sub> reduced the adsorption activity of HCl, thereby  
391 decreasing the performance of HCl removal. Despite the negative impact of CO<sub>2</sub> on the  
392 performance of HCl removal by Ca-Mg-Al-CO<sub>3</sub>, it is noteworthy that the overall performance of  
393 Ca-Mg-Al-CO<sub>3</sub> was better than that of Mg-Al-CO<sub>3</sub>.

### 394 **3.3.2 The effect of reaction temperature on HCl removal performance**

395 In Fig. 8, the change curves depict the effect of different reaction temperatures on the HCl  
396 removal performance of Ca-Mg-Al-CO<sub>3</sub>. The experimental conditions included a CO<sub>2</sub>  
397 concentration of 20 %, an initial HCl concentration of 400 ppmv, and reaction temperatures  
398 ranging from 300 to 700 °C in intervals of 100 °C. As shown in Fig.8A, the breakthrough curves  
399 shifted to the left first and then to the right with the increase of reaction temperature, the optimal  
400 HCl adsorption performance of HCl adsorption was observed at 500 °C, with a breakthrough time  
401 of approximately 190 min. This represented a significant increase of 496.8 % compared to the  
402 optimal performance of Mg-Al-CO<sub>3</sub>. When reaction temperature increased from 300 °C to 500 °C,  
403 the reaction time increased from 801min to 1123min, indicating an increase of 28.67 %. However,  
404 as the temperature further increased from 500 to 700 °C, the reaction time decreased from 1123  
405 min to 857 min, representing a decrease of 23.69%. Fig. 8B illustrates the change tendencies of  
406 average HCl removal rates, which aligned with the patterns observed in the breakthrough curves.  
407 The best HCl removal performance was achieved at 500 °C, with an average HCl removal rate

408 exceeding 80 % within 600 min.



411 Fig. 8 The performance of HCl removal using Ca-Mg-Al-CO<sub>3</sub> at different CO<sub>2</sub> atmosphere (A:  
412 the breakthrough curves; B: the average removal of HCl rate; C: the adsorption chlorine capacity  
413 curves; D: the maximum breakthrough chlorine capacity)

414 Fig. 8C presents the change curves of the adsorption chlorine capacity at different temperatures.  
415 The adsorption chlorine capacities exhibited an initial increase followed by a decrease as the  
416 reaction temperature was raised. This behavior is further supported by Fig.8 D, which indicates  
417 that the maximum breakthrough chlorine capacity of Ca-Mg-Al-CO<sub>3</sub> increased from 217.9 mg/g to  
418 389.5 mg/g as the reaction temperature increased from 300 to 500 °C, then decreased from 389.5  
419 mg/g to 276 mg/g as the reaction temperature further increased from 500 to 700 °C. These findings  
420 emphasize the significant influence of reaction temperature on the HCl removal performance of

421 Ca-Mg-Al-CO<sub>3</sub> under CO<sub>2</sub> atmosphere, with the optimal performance observed at 500 °C, reaching  
422 389.5 mg/g.

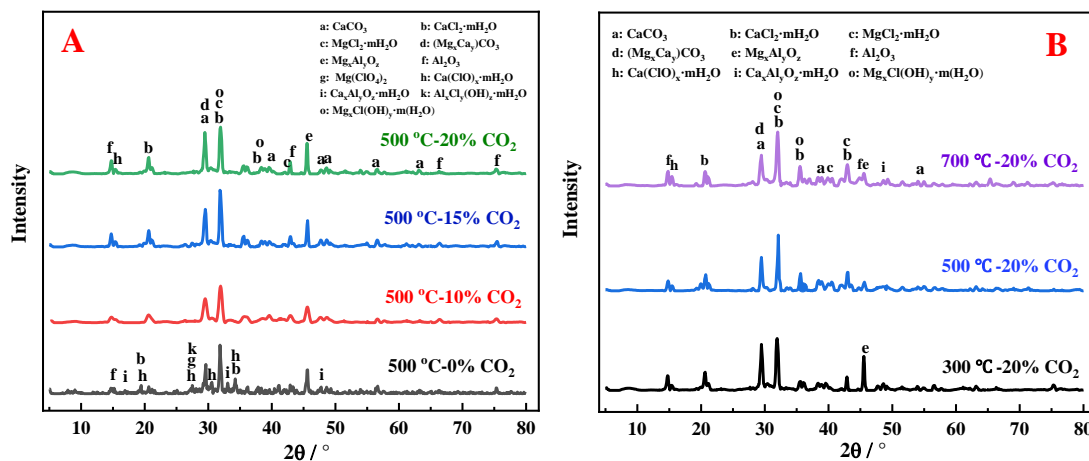
423 The observed results can be attributed to several factors. Firstly, the increase in reaction  
424 temperature promotes the development of the pore structure in Ca-Mg-Al-CO<sub>3</sub>, leading to a larger  
425 specific surface area and an increase in the number of basic active sites. These changes enhance  
426 the adsorption and reaction of HCl with the adsorbent. The presence of Ca<sup>2+</sup> in the hydrotalcite  
427 structure plays a crucial role in determining the optimal condition at 500 °C. At this temperature,  
428 the Ca-Mg-Al-CO<sub>3</sub> exhibits the highest performance in HCl removal. However, when the  
429 temperature exceeds 500 °C, the porous structure of the hydrotalcite can be destroyed, the effective  
430 component of HCl removal decreased and more substances that were not easy to react with HCl  
431 were generated (such as spinel). These transformations prevent the regeneration of the hydrotalcite  
432 structure and lead to a decrease in the number of active sites on the surface. Consequently, the  
433 reaction paths involving Cl<sup>-</sup> replacing CO<sub>3</sub><sup>2-</sup> and the reaction of Cl<sup>-</sup> with Ca<sup>2+</sup> are hindered,  
434 resulting in a reduction in the performance of HCl removal. Therefore, the optimal temperature for  
435 efficient HCl removal by Ca-Mg-Al-CO<sub>3</sub> was determined to be 500 °C, while temperatures above  
436 this threshold were detrimental to the adsorption capacity and overall performance.

437 The results of the HCl removal using Mg-Al-CO<sub>3</sub> and Ca-Mg-Al-CO<sub>3</sub> in the presence of CO<sub>2</sub>  
438 indicate different effects of CO<sub>2</sub> on their performance. In the case of Mg-Al-CO<sub>3</sub>, CO<sub>2</sub> has a  
439 positive impact on HCl removal, whereas in the case of Ca-Mg-Al-CO<sub>3</sub>, CO<sub>2</sub> has a negative effect.  
440 Despite the adverse effect of CO<sub>2</sub> on Ca-Mg-Al-CO<sub>3</sub>, the overall performance of HCl removal  
441 using Ca-Mg-Al-CO<sub>3</sub> was better than that of Mg-Al-CO<sub>3</sub>. When the Mg-Al-CO<sub>3</sub> was modified by  
442 introducing Ca<sup>2+</sup>, the optimum reaction temperature was 500 °C, and the maximum breakthrough

443 chlorine capacity increased to 389.5 mg/g, representing a 103.7 % increase compared to the  
 444 highest breakthrough chlorine capacity of Mg-Al-CO<sub>3</sub>. The presence of Ca<sup>2+</sup> not only improved  
 445 the performance of HCl removal by Mg-Al-CO<sub>3</sub>, but also expanded the optimal reaction  
 446 temperature range for HCl removal using hydrotalcite in a CO<sub>2</sub> atmosphere. This expanded  
 447 temperature range holds great significance for the utilization of remaining heat in subsequent  
 448 stages of the process. The improved performance can be attributed to several key factors. Firstly,  
 449 calcium exhibits superior metallicity when compared to magnesium, leading to changes in  
 450 chemical bonding upon the addition of Ca. The introduction of Ca<sup>2+</sup> and the formation of new  
 451 chemical bonds facilitate the interaction and combination with Cl<sup>-</sup>. Secondly, the SEM results  
 452 from Fig.4 B and E reveals significant modifications in the hydrotalcite structure upon the  
 453 incorporation of Ca. Notably, the layer spacing and size of Ca-Mg-Al-CO<sub>3</sub> was larger in  
 454 comparison to Mg-Al-CO<sub>3</sub>. This structural alteration allows for easier permeation of HCl into the  
 455 interior of the absorbent, enabling more efficient adsorption and removal of HCl.

### 456 3.4 The characterization and mechanism

457 To further analysis the mechanism underlying the reaction between Ca-Mg-Al-CO<sub>3</sub> and HCl in  
 458 the presence of CO<sub>2</sub>, the XRD patterns of the products were obtained and are presented in Fig. 9.

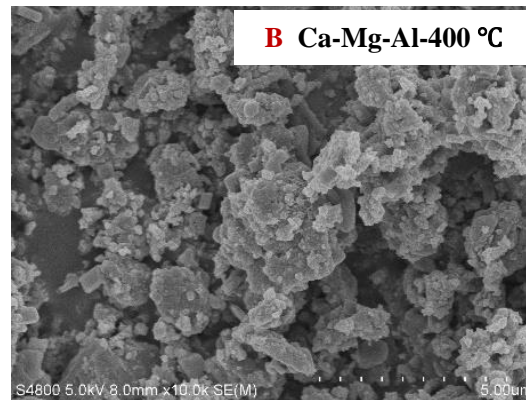
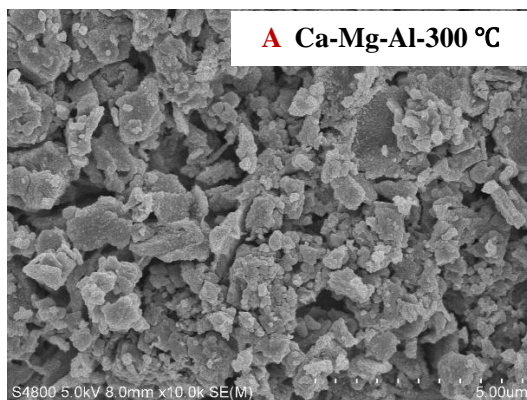


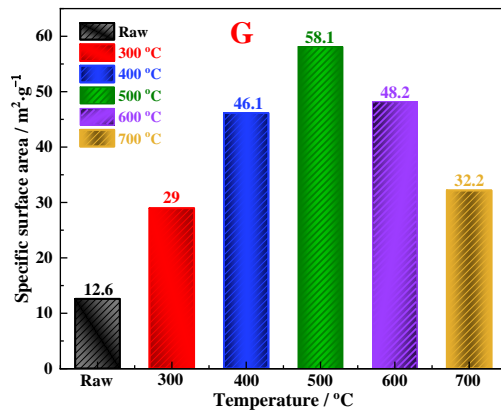
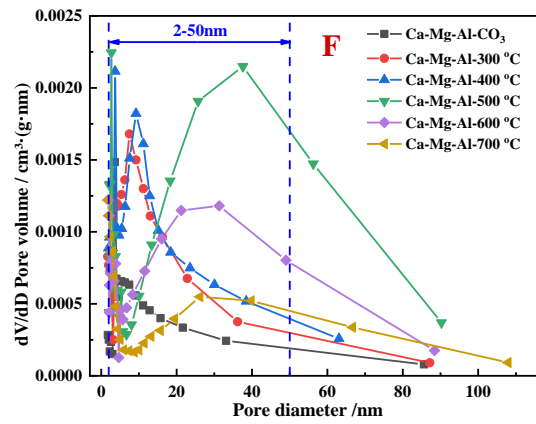
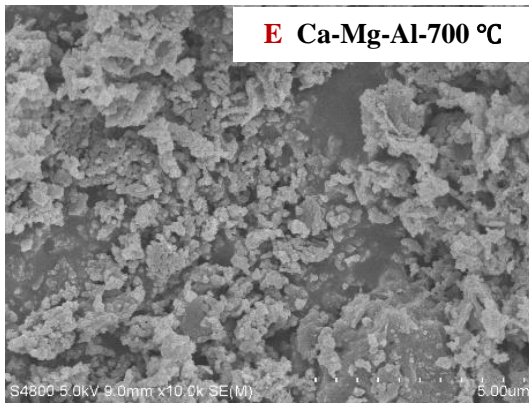
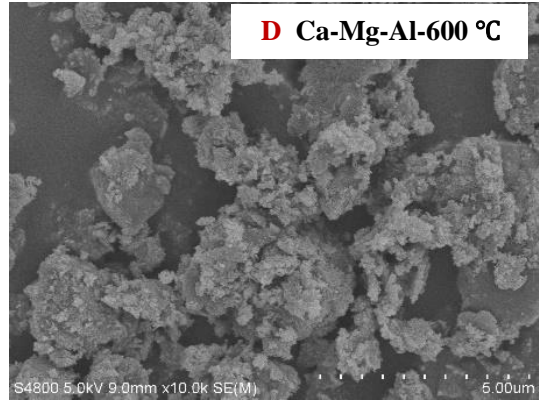
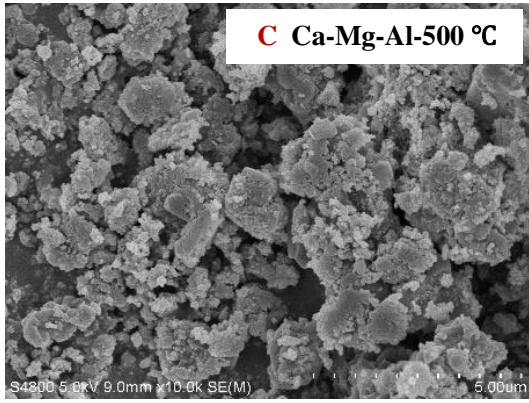
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460 Fig. 9 The XRD patterns of reaction products at different CO<sub>2</sub> concentrations and temperatures (A:  
461 the different CO<sub>2</sub> concentrations; B: the different temperatures)

462 The XRD patterns of the products obtained at 300 °C under different CO<sub>2</sub> concentrations are  
463 shown in Fig. 9A. A comparison with the XRD patterns of raw Ca-Mg-Al-CO<sub>3</sub> from Fig. 4D  
464 reveals a significant decrease in the characteristic diffraction peak of hydrotalcite, indicating its  
465 transformation during the reaction. The main products observed were chloride and a mixture of  
466 oxidate compounds containing calcium, magnesium, and aluminum, such as MgCl<sub>2</sub>, CaCl<sub>2</sub>, Al<sub>2</sub>O<sub>3</sub>,  
467 Mg<sub>x</sub>Al<sub>y</sub>O<sub>z</sub>, Ca<sub>x</sub>Al<sub>y</sub>O<sub>z</sub>. The presence of metal oxides indicated that they were difficult to react with  
468 Cl. The characteristic diffraction peak of CO<sub>3</sub><sup>2-</sup> containing substance decreased, and in the absence  
469 of CO<sub>2</sub>, the reaction products included not only Cl<sup>-</sup>, but also ClO<sub>x</sub><sup>-</sup>. With the increase of CO<sub>2</sub>  
470 concentration, the characteristic diffraction peaks of Mg (ClO<sub>4</sub>)<sub>2</sub>, Ca (ClO)<sub>x</sub>, and Al<sub>x</sub>Cl<sub>y</sub> (OH)<sub>z</sub>  
471 became weak and eventually disappeared, CaCl<sub>2</sub> and MgCl<sub>2</sub> became the main products of chloride,  
472 indicating that CO<sub>2</sub> influenced the formation of the products during the reaction between HCl and  
473 adsorbent. Furthermore, the characteristic diffraction peak of CO<sub>3</sub><sup>2-</sup> containing substance became  
474 stronger, especially that of CaCO<sub>3</sub>, became stronger, suggesting that Ca<sup>2+</sup> reacted with CO<sub>2</sub> to  
475 generate large particles of CaCO<sub>3</sub>. This reaction occupied the active sites on the surface of the  
476 adsorbent, leading to the formation of a dense layer of CaCO<sub>3</sub>. Consequently, the diffusion  
477 resistance of HCl increased, and the reaction rate between CaCO<sub>3</sub> and HCl was slower compared  
478 to that between Ca<sup>2+</sup> and HCl[56, 57]. This phenomenon contributed to the decrease in the  
479 performance of HCl removal by the adsorbent. Fig. 9B presents the XRD patterns of the products  
480 obtained at a CO<sub>2</sub> concentration of 20 % under different reaction temperatures: 300, 500 and  
481 700 °C. It can be seen from Fig. 9B that the intensity of the characteristic diffraction peaks

482 corresponding to carbonates decreased with the increase of temperature, suggesting that the  
483 carbonates were prone to decomposition, and the generation of  $\text{CaCO}_3$  became challenging at high  
484 temperatures. Additionally, the peak intensity of  $\text{Mg}_x\text{Al}_y\text{O}_z$  became weak, indicating the  
485 conversion of  $\text{Mg}_x\text{Al}_y\text{O}_z$  into stable compounds such as  $\text{MgAl}_2\text{O}_4$  and  $\text{Al}_2\text{O}_3$ , and these substances  
486 didn't react with HCl. Notably, the intensity of the characteristic diffraction peaks of chlorides was  
487 more pronounced at 500 °C compared to other temperatures. This finding suggests that the  
488 adsorption performance of HCl removal was enhanced at 500 °C. The surface morphology, pore  
489 size distribution, and specific surface area of Ca-Mg-Al- $\text{CO}_3$  at different temperatures were shown  
490 in Fig.10. As shown in the SEM images of Figure 10A-E, the surface morphology of Ca-Mg-Al-  
491  $\text{CO}_3$  becomes increasingly porous as the temperature increases. However, a sintering phenomenon  
492 occurs when the temperature exceeds 500 °C. In Fig.10F, the pores of Ca-Mg-Al- $\text{CO}_3$  are mainly  
493 mesopores. With the increase of temperature, the number of pores first increases and then  
494 decreases, and pores develop into larger pores, the optimal pore structure is at 500 °C. Therefore,  
495 as shown in Fig. 10G, the specific surface area also first increases and then decreases, the specific  
496 surface area was larger at 500 °C, 58.1  $\text{m}^2/\text{g}$ . The large specific surface area provides more active  
497 sites, facilitating the availability of more active sites. Above results further illustrated that why the  
498 Ca-Mg-Al- $\text{CO}_3$  exhibits excellent performance in removing HCl at 500 °C.





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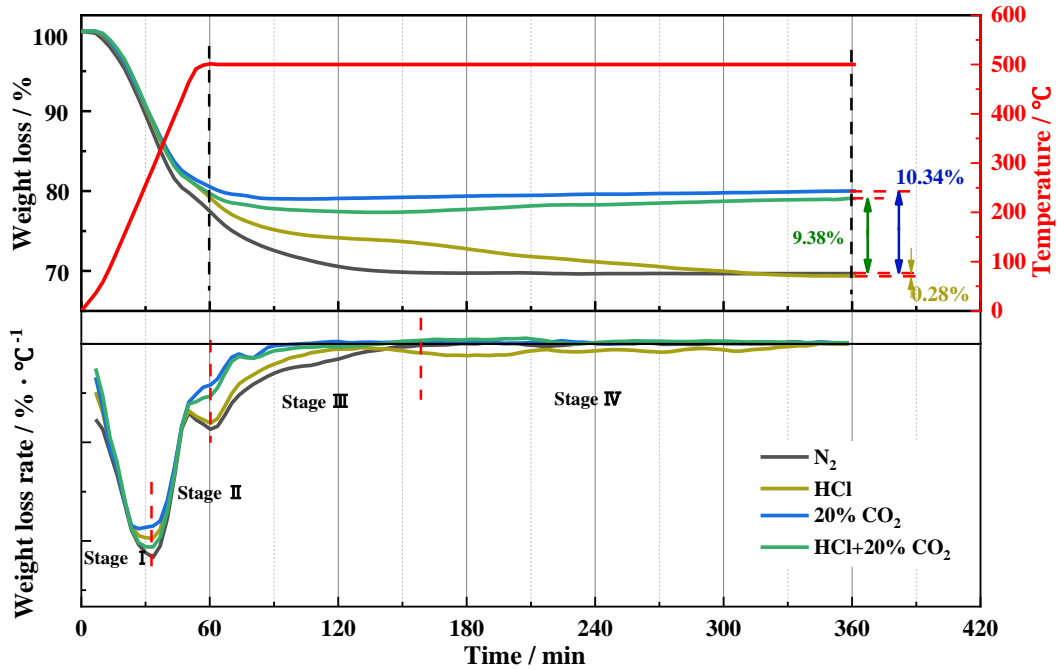
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503 Fig. 10 The surface morphology, pore size distribution, and specifal surface area of Ca-Mg-Al-  
 504 CO<sub>3</sub> at different temperatures (A-E: Surface morphology; F: Pore size distribution; G: Specific  
 505 surface area)

506 **3.5 Thermogravimetric experiments**





507

508 Fig. 11 Adsorption thermogravimetric experiments of HCl and CO<sub>2</sub> by Ca-Mg-Al-CO<sub>3</sub>

509 Fig. 11 displays the TG/DTG curves obtained from - TGA tests conducted on Ca-Mg-Al-CO<sub>3</sub>,  
 510 involving N<sub>2</sub>, 400 ppmv HCl, 20 % CO<sub>2</sub>, and a mixture of 400 ppmv HCl and 20 % CO<sub>2</sub>. The  
 511 duration of the reaction was 5 hours. During the heating process, the thermogravimetric analyzer  
 512 followed a linear heating curve, and the temperature stabilized at 500 °C after 60 min.  
 513 Consequently, the reaction gases were introduced at the 60-minute mark, and the reaction  
 514 continued for 5 h. As shown in the TG curve of N<sub>2</sub>, a weight loss of 30.32% was observed. This  
 515 weight loss primarily consisted of interlayer crystal H<sub>2</sub>O and a portion of CO<sub>3</sub><sup>2-</sup>. The DTG curve  
 516 of N<sub>2</sub> revealed three distinct stages of weight loss. In stage I , which occurred at temperatures  
 517 below 250 °C, only interlayer moisture was lost, without affecting the structure of the material.  
 518 This stage exhibited the highest weight loss peak, with Mg-Al-CO<sub>3</sub> losing 12.7 % of its weight. In  
 519 stage II , as the temperature increased to 250~500 °C, CO<sub>2</sub> was generated, resulting in the  
 520 appearance of the second weight loss peak, corresponding to a weight loss of 22.6 %. Stage III  
 521 commenced when the heating temperature reached 500 °C, with continuous loss of CO<sub>3</sub><sup>2-</sup> and a



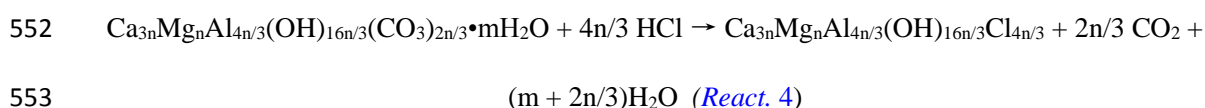
522 stable weight loss for Ca-Mg-Al-CO<sub>3</sub>[41].

523 Upon introducing HCl gas, an increase in weight was observed compared to the N<sub>2</sub> condition,  
524 indicating a rapid adsorption stage for HCl. In the DTG curve, four weight loss peaks were  
525 observed in the stage III and IV. This phenomenon can be explained by the reaction of HCl with  
526 the metals present on the adsorbent's surface, resulting in an increase in weight. Consequently, the  
527 weight loss rate slowed down due to these chemical reactions. The chemical reaction equations for  
528 this process are shown in *React. 2 and 3* (M represents metal). as the reaction time increased, the  
529 active sites available for HCl adsorption became occupied. At this stage, Cl<sup>-</sup> entered the interior of  
530 the adsorbent through the pores, replacing undecomposed CO<sub>3</sub><sup>2-</sup>, and this process is illustrated by  
531 *React. 4*, which represents the main reaction pathway in this stage. As the molecular weight of Cl<sup>-</sup>  
532 was less than that of CO<sub>2</sub>, the weight loss decreased rapidly. Consequently, the overall weight loss  
533 reached 30.6%, which represented a decrease of 28 % compared to the TG curve of N<sub>2</sub>.

534 In the presence of 20 % CO<sub>2</sub>, an increase in weight was observed upon introducing CO<sub>2</sub>. The  
535 weight increase was greater than that observed in the HCl gas condition, and it was the highest  
536 among the four TG curves, indicating a stronger adsorption capacity for CO<sub>2</sub>. In the DTG curve,  
537 three peaks of weight increase were observed in the stage III and IV, providing further evidence of  
538 the CO<sub>2</sub> adsorption process. Comparing it with the TG curve of N<sub>2</sub>, the ultimate weight increased  
539 by 10.34 %. This increase can be attributed to the reaction between the adsorbent and CO<sub>2</sub>,  
540 leading to the formation of new carbonates, primarily CaCO<sub>3</sub>, as shown in *Fig. 9*.

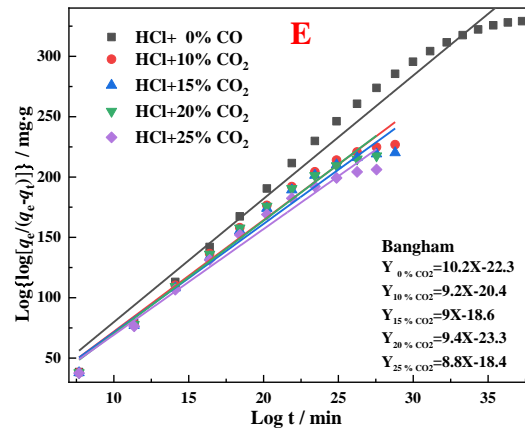
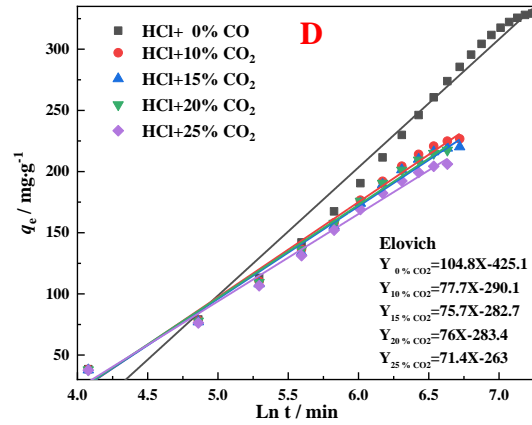
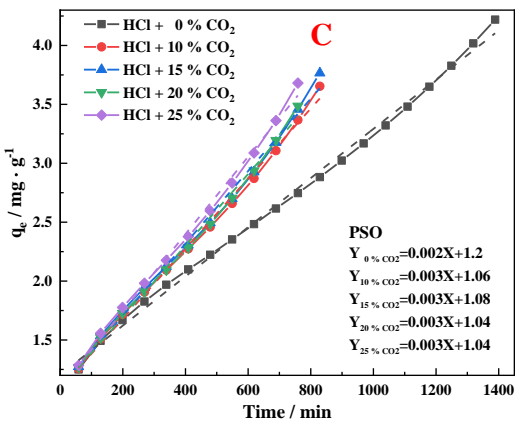
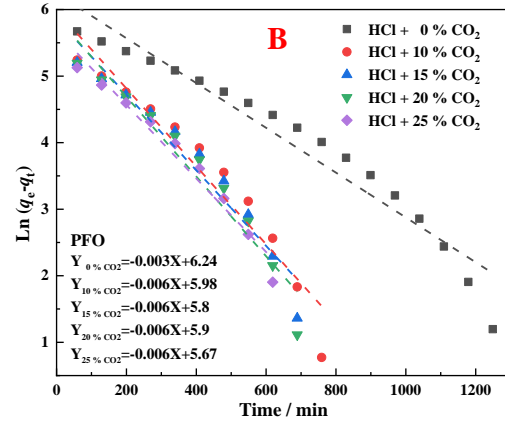
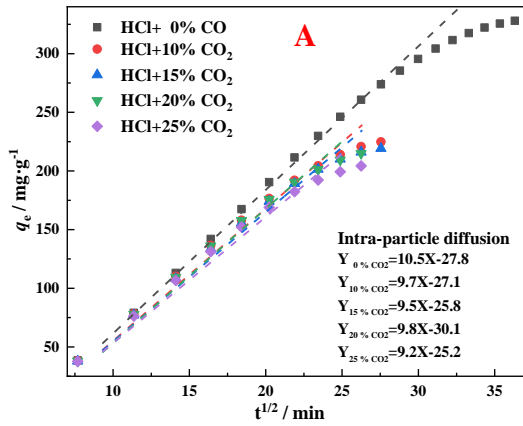
541 In the presence of 20 % CO<sub>2</sub> and 400 ppmv HCl, the TG curve showed a final weight increase  
542 of 9.38 % compared to the product in N<sub>2</sub>. In the DTG curve, there was a loss weight peak in stage  
543 III, mainly caused by HCl gas, as observed in the DTG curve of 20 % CO<sub>2</sub>. Subsequently, there

544 were two weight increase peaks in stage IV, caused by CO<sub>2</sub> gas. These results indicated a  
 545 competitive relationship between HCl and CO<sub>2</sub> during the reaction process, the reaction rate of the  
 546 adsorbent with HCl was faster than that of CO<sub>2</sub> in the early stage, and the adsorption capacity for  
 547 CO<sub>2</sub> was stronger than for HCl. This can be attributed to the significantly higher concentration of  
 548 CO<sub>2</sub> compared to HCl. This competitive relationship explains the performance of HCl removal  
 549 using Ca-Mg-Al-CO<sub>3</sub> in the presence of CO<sub>2</sub>.



### 554 3.6 Apparent adsorption kinetic

555 The apparent adsorption kinetic models mainly include kinetic control type and diffusion control  
 556 type. The diffusion control type includes Intra-particle diffusion model, which describes gas film  
 557 or intra-particle diffusion as a rate-controlling step. The commonly used kinetic control models are  
 558 Pseudo-first-order model (PFO), Pseudo-second-order model (PSO), Elovich model and Bangham  
 559 model, all of which describe the kinetic adsorption process of surface adsorption or surface  
 560 chemical reactions as rate-controlling steps. The experimental results of HCl removal by Ca-Mg-  
 561 Al-CO<sub>3</sub> at different CO<sub>2</sub> concentrations were analyzed using the above models, and the fitting  
 562 results of adsorption kinetics are presented in [Fig. 12](#).



563

564

565

566 Fig. 12 The fitting results of different adsorption kinetic model (A: Intra-particle diffusion model;

567 B: Pseudo-first-order model (PFO), C: Pseudo-second-order model; D: Elovich model; E:

568 Bangham model)

569 As depicted in Fig. 12 and the correlation coefficients ( $R^2$ ) summarized in Table 2, based on the

570 fitting result of the intra-particle diffusion model, the fitted line does not pass through the origin

571 and did not exhibit a linear relationship, the  $R^2$  values are more than 0.98, indicating that diffusion

572 control is not the solely governed by single rate-determining steps. The fitting results of 4 kinetic  
 573 control types showed a relatively best agreement with the PSO model, the  $R^2$  values exceed 0.99.  
 574 The result illustrates that the chemical adsorption is main reaction mechanism in whole reaction  
 575 process, and chemical adsorption involves electron sharing or electron transfer between Ca-Mg-  
 576 Al-CO<sub>3</sub> and HCl, CO<sub>2</sub>. Compared to PSO model. Although the  $R^2$  values of other 3 adsorption  
 577 kinetic models were slightly lower than that of the PSO model, all models exhibited correlation  
 578 coefficients above 0.97, as shown in Fig. 12B, D and E, there a good fitting result in the early  
 579 stage of the reaction, indicating it is controlled by multiple control steps in the early stage of the  
 580 reaction.

581 **Table. 2 The correlation coefficients ( $R^2$ ) of different adsorption kinetic models**

Model	0% CO <sub>2</sub> +HCl	10% CO <sub>2</sub> + HCl	15% CO <sub>2</sub> + HCl	20% CO <sub>2</sub> +HCl	25% CO <sub>2</sub> +HCl
Intra-particle diffusion	0.983	0.985	0.983	0.987	0.985
PFO	0.94	0.933	0.946	0.943	0.968
PSO	0.996	0.996	0.994	0.996	0.996
Elovich	0.98	0.988	0.987	0.989	0.99
Bangham	0.977	0.976	0.972	0.978	0.974

#### 582 4. Conclusions

583 In this study, the performance of HCl removal using Mg-Al-CO<sub>3</sub> and Ca-Mg-Al-CO<sub>3</sub>, as  
 584 modified sorbent, in the presence of CO<sub>2</sub> were investigated.

585 The Ca-Mg-Al-CO<sub>3</sub> was successfully synthesized with a larger layer spacing  $d_{003}$  and particle  
 586 size compared to Mg-Al-CO<sub>3</sub>. CO<sub>2</sub> had a positive impact on the removal of HCl using Mg-Al-CO<sub>3</sub>,  
 587 but a negative impact on the removal of HCl using Ca-Mg-Al-CO<sub>3</sub>. However, the performance of  
 588 HCl removal using Ca-Mg-Al-CO<sub>3</sub> was superior to that of Mg-Al-CO<sub>3</sub>. The main reason for this  
 589 difference is that CO<sub>2</sub> can react with Mg-Al-CO<sub>3</sub> to generate new hydrotalcite intercalated anions  
 590 through a regeneration process, thereby enhancing the performance of HCl removal. In contrast,

591 Ca-Mg-Al-CO<sub>3</sub> reacted with CO<sub>2</sub> to form large particles of CaCO<sub>3</sub>, which not only occupied the  
592 active sites but also hindered the reaction between HCl and the adsorbent, leading to a decrease in  
593 the performance of HCl removal. The main products of reaction between Ca-Mg-Al-CO<sub>3</sub> and HCl  
594 were CaCl<sub>2</sub> and MgCl<sub>2</sub>. The presence of CO<sub>2</sub> plays a competitive role in the removal of HCl. The  
595 adsorption rate of HCl was initially faster than that of CO<sub>2</sub>, but the adsorption capacity of CO<sub>2</sub> was  
596 stronger. This can be attributed to the significantly higher concentration of CO<sub>2</sub> compared to HCl.  
597 The adsorption reactions of HCl and CO<sub>2</sub> by Ca-Mg-Al-CO<sub>3</sub> was controlled by multiple rate-  
598 controlling mechanisms in the early stage of the reaction, and follow a PSO model in the whole  
599 process. The study demonstrates that Ca-Mg-Al-CO<sub>3</sub> exhibits excellent performance in removing  
600 HCl, which is of significant importance for achieving high-quality synthesis gas and further  
601 promoting pyrolysis gasification technology. The findings have academic and industrial  
602 significance in advancing waste-to-energy conversion technologies and contributing to the  
603 reduction of carbon emissions.

#### 604 **Credit author statement**

605 Songshan Cao: Investigation, Methodology, Writing - original draft. Jun Cao: Writing - review &  
606 editing, Funding acquisition. Hualun Zhu: Writing - review & editing. Yaji Huang: Supervision.  
607 Baosheng Jin: Supervision, Funding acquisition. Massimiliano Materazzi: Supervision, Writing -  
608 review & editing.

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**Author statement**

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