1	Life cycle assessment of electricity generation from sugarcane bagasse
2	hydrochar produced by microwave assisted hydrothermal carbonization
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16	Abstract
17	Microwave assisted hydrothermal carbonization (MAHTC) is a promising strategy for fuel production
18	from sugarcane bagasse. Even though microwave heating has been reported to facilitate product
19	(hydrochar) yield, energy properties and heating efficiency during hydrothermal carbonization (HTC)
20	process, the environmental consequences of MAHTC treatment were still not clear. This study
21	evaluated the environmental impact associated with 1 MJ electricity generation from sugarcane
22	bagasse hydrochar through life cycle assessment (LCA) method, focusing on the critical role of
23	process-based parameters to provide insights for optimization of MAHTC treatment. Specifically, two
24	different allocation factors (energy content and economic value) and five environmental impact
25	categories (climate change, freshwater eutrophication, freshwater ecotoxicity, human toxicity and
26	fossil depletion) were assessed in this study. The LCA results revealed significant contribution of
27	MAHTC process on climate change and fossil depletion because of large energy consumption used to
28	maintain the system at designed temperature. Discharge of liquid phase from MAHTC process
29	resulted in severe eutrophication impact especially when economic value was used as allocation

30	factor. Gas emission from hydrochar combustion caused most toxicity related impacts indicating
31	essential requirement of further investigation to quantify different gaseous composition. Based on
32	LCA results, sensitivity analysis indicated hydrochar yield and carbon content as the top two
33	influential factors on total environmental consequences. Comparison study with other fuel sources
34	were further conducted to identify the influence of substituted energy sources. The overall results
35	suggested MAHTC as promising method for bagasse utilization and energy retention efficiency as
36	important indicator for optimization of MAHTC treatment or the sake of high-quality products and
37	good environmental performance.
38	Key words
39	Life cycle assessment, Microwave assisted hydrothermal carbonization, Sugarcane bagasse,
40	Environmental impact, Hydrochar production

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42 **1. Introduction**

China, as the third largest sugar producer in the world, is manufacturing 10.25 million tonnes of 43 sugar annually (Peng et al., 2014) and over 90% of those are cane sugar. Sugarcane bagasse, the 44 45 fibrous residue after juice extraction of sugarcane, thus becomes the largest agro-industrial residue 46 and by-product of sugar industry (Sindhu et al., 2016). Various technologies have been applied on 47 utilising sugarcane bagasse including direct combustion for electricity generation, pulp and paper production, and fermentation for value-added products (Ahmed and Gupta, 2012). In most cases, 48 49 bagasse is used directly for co-combustion to generate heat and power (Sindhu et al., 2016) and 50 then used as primary fuel source for sugar mills (Peng et al., 2014).

51 The high carbon content of sugarcane bagasse, reported as 47% on a moisture and ash free basis 52 (Wright et al., 2016), is quite desirable for fuel production. With over 50% water content in the 53 original bagasse (Erlich et al., 2005), thermochemical processes are usually required as pre-54 treatment for fuel production (Bajpai, 2016). Hydrothermal carbonization (HTC), a novel 55 thermochemical process to transform biomass with high water content into carbon-rich solid 56 product (Kim et al., 2016), has been proposed as suitable fuel production method for biomass 57 feedstocks like sugarcane bagasse (Melo et al., 2017). As the suitability of biomass waste for fuel 58 production largely depends on its heating value (Wright et al., 2016), numerous studies have been 59 focus on modification of HTC process parameters to obtain fuel product, so-called hydrochar, with 60 better energy properties from different feedstocks, such as municipal waste streams (Berge et al., 61 2011), pure cellulose (Lu et al., 2013) and lignocellulosic biomass (Xiao et al., 2012). The 62 conventional HTC treatment using sugarcane bagasse as feedstock resulted in only 20.3% increase in 63 calorific value at 180 $^\circ$ C for 30 minutes (Chen et al., 2012). A more satisfying result of 34.0% increase 64 in energy content of hydrochar was observed (Hoekman et al., 2013), however it required large 65 amount of energy to maintain the reaction at $295\,^\circ$ C. Microwave heating has been reported with 66 enhancement effects on process kinetics and considerable reduction of energy consumption during

HTC of pure cellulose (Zhang et al., 2018). Through a comparison between HTC and microwave
assisted hydrothermal carbonization (MAHTC) of lignocellulosic waste material (Elaigwu and
Greenway, 2016a), hydrochars with similar energy properties were generated with 10 times
reduction in reaction time under MAHTC. MAHTC, as a potential strategy to improve the fuel
properties of biomass feedstock with less energy consumption (Elaigwu and Greenway, 2016b), has
risen a particular interest of its accessibility for fuel production from sugarcane bagasse.

73 Although there are several studies suggesting advantages of MAHTC previously, a systems level 74 analysis is still needed to identify and quantify environmental consequences (Berge et al., 2015) for 75 further application. It is also essential to understand the environmental performance of electricity 76 production from hydrochar, as it frequently account for a major portion of the total environmental 77 burdens (Ramjeawon, 2008) identified in the life cycle of sugarcane bagasse. Life cycle assessment 78 (LCA) is a comprehensive method for analysis of a process/system and optimization of industry 79 (Chauhan et al., 2011). It has been successfully applied in conducting system level analysis and 80 addressing environmental sustainability (Borrion et al., 2012). There are few existing studies using 81 LCA to estimate the environmental impacts of electricity generation from hydrochar. Berge et al. 82 (2015) evaluated 9 types of environmental impacts associated with HTC of food waste and the 83 subsequent combustion of the generated hydrochar for energy production using lab scale data. The 84 LCA results indicated that treatment of process water HTC process and the energy sources being 85 replaced by hydrochar should be concerned to reduce environmental loads. Another study 86 conducted by Owsianiak et al. (2016) assessed 15 impact categories from the International 87 Reference Life Cycle Data System (ILCD) handbook of HTC using four types of wet biomass waste 88 streams as feedstock at industry-relevant scale. Although the application of HTC treatment with 89 hydrochar as solid fuel shown to be environmentally attractive (Owsianiak et al., 2016), their results 90 also suggested a range of factors like plant size and geographic location could largely influence the 91 overall environmental performance.

92 Previous LCA studies of HTC have shown interests in comparing environmental impacts of HTC and 93 other waste treatment methods (Berge et al., 2015) or HTC treatment using different feedstock 94 (Owsianiak et al., 2016). Not enough attention was offered on HTC process itself and the parameters 95 related with hydrochar production when conducting LCA on HTC. Only hydrochar moisture and 96 energy content have been discussed in LCA study of HTC using olive mill waste as feedstock 97 (Benavente et al., 2017). More process related parameters should be considered in LCA to reveal 98 how these parameters contribute to system environmental impacts, which could also provide 99 guidance for optimization of current MAHTC treatment. Moreover, the function of allocation 100 method in the previous LCA studies of HTC was ignored and required further investigation to 101 uncover its impact. As the whole electricity generation from hydrochar production by MAHTC 102 involves several multifunctional stages in the life cycle, allocation is an important issue in this 103 biorefinery system (Karka et al., 2015). Since sugarcane bagasse is the major by-product from sugar 104 industry, it is also essential to combine the environmental impact of MAHTC process with initial 105 bagasse production and the final use of hydrochar to understand the whole system instead of 106 focusing on the process itself.

107 A process based LCA was conducted on the MAHTC of sugarcane bagasse and electricity generation 108 from subsequent hydrochar to evaluate environmental performance and facilitate optimization of 109 MAHTC treatment. As far as we know, evaluation of environmental impacts associated with 110 hydrochar from MAHTC treatment for electricity generation has not been carried out. The aim of this 111 paper is to conduct LCA of electricity generation from sugarcane bagasse hydrochar to evaluate the 112 system environmental impact with two different allocation methods and identify critical MAHTC 113 process parameters on environmental burdens for future optimization. This work has three specific 114 objectives, (1) to investigate the environmental hotspots of electricity production from sugarcane 115 bagasse hydrochar generated from MAHTC with allocation of economic benefit and energy content 116 respectively; (2) to understand the effects of different process related parameters on LCA results to 117 minimize environmental impacts for optimization of MAHTC treatment; (3) to evaluate the feasibility

118 of sugarcane bagasse for fuel production through MAHTC treatment from environmental 119 perspective with comparison of other fuel sources. The novelty of this research is to establish the 120 connection between detailed MAHTC process design with systematic environmental impact of 121 electricity generation from bagasse hydrochar. By expanding the lab-scale result into practical 122 system level, this study presents a holistic understanding of the environmental consequences of the 123 implement of MAHTC as a promising waste-to-energy treatment for other biomass waste like 124 sugarcane bagasse. The LCA results from this study can also provide useful insights for both biomass 125 waste treatment and sugarcane bagasse biorefinery.

126 2. Methods

LCA modelling was performed based on ISO 14040 and 14044 standards (Finkbeiner et al., 2006) to evaluate the environmental impact associated with material balance and energy flow of electricity generation from sugarcane bagasse hydrochar. This study comprises four main phases as goal and scope, life cycle inventory (LCI) analysis, life cycle impact assessment (LCIA) and interpretation.

131 2.1 Goal and scope

132 2.1.1 Goal of the study

133 The goal of the present research is to quantify environmental impacts of electricity production from

134 sugarcane bagasse hydrochar generated via MAHTC, identify critical process parameters from

135 sensitivity analysis and compare the environmental consequences with other fuel resources.

136 2.1.2 Scope of the study and functional unit

The system boundary of this study is illustrated in Fig.1. Two scenarios (S1 and S2) are discussed based on two allocation methods and different treatment of liquid phase product from MAHTC of bagasse. In both scenarios the system starts from sugarcane collection and transportation to the electricity generation from combustion plant, including three main processes as sugarcane milling, MAHTC of sugarcane bagasse and combustion of hydrochar. Sugarcane production process is not

- 142 included in the system since it has been discussed previously in detail in LCA studies focusing on
- sugar industry from Australia (Renouf et al., 2010), South Africa (Mashoko et al., 2010) and



144 worldwide (Chauhan et al., 2011).



An MAHTC plant is assumed to treat 4,000 t/y sugarcane bagasse on wet basis after sugarcane 149 150 milling. The plant with two reactors system is assumed to run 300 d annually, which resembles to 151 the capacity of an in-used HTC plant in Germany run by HTCycle company (HTCycle, 2018). Each 152 reactor contains four batches with capacity of treating 200 kg sugarcane bagasse on wet basis. Eight 153 batches of experiment are assumed to run per day to produce 3.35 t energy-enriched hydrochar. 154 The HTCycle case is selected as reference in this study because there is no existing in-used HTC plant 155 in China up to now and HTCycle is cooperating with several Chinese companies to implement their 156 technology (HTCycle, 2019), such as Huizhou Tongyong, Beijing Tempro Technologies Inc etc. After MAHTC treatment the hydrochar will be made into briquette (Liu et al., 2017) and sent to 157 incineration plant to produce 6.53×10^{10} J electricity per day. 158

159 The functional unit in this study is generation of 1 MJ (10⁶ J) electricity from sugarcane bagasse

160 hydrochar produced from MAHTC treatment. As hydrochar being assumed as potential fuel, using

161 electricity as functional unit would benefit the comparison of environmental impact between162 different fuel sources.

163 2.1.3 Allocation method

164 Table 1 lists the value of allocation factors used in two scenarios. The allocation factors for 165 sugarcane milling process are adopted from previous research and experimental results are used for 166 allocation in MAHTC of sugarcane bagasse. Both economic benefit (Gnansounou et al., 2015) and 167 energy content (Renó et al., 2011) have been used as allocation factors for sugarcane based 168 biorefineries. Bagasse is generally regarded as a processing waste with limited value in sugarcane industry, in which way the economic performance is sensitive to sugarcane milling process. As 169 170 presented in Table 1, allocation factor is calculated by economic value in S1, where the liquid phase 171 generated after MAHTC process without further treatment has no actual economic value and is 172 assumed as direct discharge to wastewater treatment facility. However, the distribution in bagasse 173 and juice as well as solid and liquid products from MAHTC treatment would largely change when the 174 energy content is taken as allocation factor. In S2, the allocation is based on the energy content of 175 products from both milling process and MAHTC treatment, where the liquid phase was taken as by-176 product with further extraction treatment to recover organic compounds. The environmental impact 177 associated with further treatment of liquid phase after MAHTC treatment is not discussed in the designed system boundary as the focus of this study is on application of solid products. 178

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Table 1. Allocation factors in two different scenarios

Scenario		Sugarca	ne milling	MAHTC of sugarcane bagasse		Reference
S1	Products	Bagasse	Juice	Hydi	Hydrochar (Gr	
(Economic value)	Allocation factor	17%	83%	100% et		et al., 2015)
S2	Products Bag		Juice	Hydrochar	Liquid products	(Renó et al., 2011); Experimental
(Energy content)	Allocation factor	49%	51%	74.2%	25.8%	result

181 2.2. Life cycle inventory analysis

Existing literature with LCA of HTC treatment and other energy recovery methods of sugarcane
bagasse is used to provide secondary data for sugarcane milling and hydrochar combustion
processes. For the MAHTC treatment of sugarcane bagasse, our lab-scale experiment of sugarcane
bagasse is used as primary source to model the scale-up plant. Life cycle inventory data from
Ecoinvent 3.0 database is employed in this study.

187 2.2.1 LCI calculation

188 2.2.1.1 Sugarcane milling

189 The sugarcane milling process is adopted from the ED scenario in the research of Gnansounou et al. 190 (2015), where bagasse and juice are the only products from this process. Tap water is used for 191 feedstock washing and imbibition. Sugarcane juice and bagasse are separated after going through 192 steam for dirt and impurities removal. A small amount of lime is added in the end to adjust the pH 193 value of sugarcane juice. Data of bagasse production are summarized from sugarcane based biorefinery researches in Brazil (Gnansounou et al., 2015), Thailand (Nguyen et al., 2008), Mauritius 194 195 (Ramjeawon, 2008) and China (Peng et al., 2014) and results in consistence that the yield of 196 sugarcane bagasse is about 30% with 50% water content. Emission of sugarcane milling process is 197 ignored in this study since the main emission is generated during further treatment of sugarcane 198 juice (Ensinas et al., 2007).

199 2.2.1.2 HTC treatment of sugarcane bagasse

Energy and mass balance of MAHTC process is calculated based on lab experimental results. The hydrochar sample using sugarcane bagasse as feedstock was obtained at 240°C for 30 minutes in the same batch reactor described in our previous study (Zhang et al., 2018), resulting in 38.9% increase in carbon content and 36.3% increase in calculated high heating value (HHV) comparing to raw

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bagasse. The optimum condition is selected from temperature range of 200°C to 250°C and time
range of 0 to 90 minutes and identified as the condition when the highest energy retention
efficiency (ERE) is reached (Wang et al., 2018). ERE is an indicator usually used to evaluate the extent
of HTC treatment in previous research (Lu et al., 2013), which is calculated by solid yield multiply
energy densification ratio as a measure of the fraction of feedstock energy retained within
generated hydrochar. The properties of sugarcane bagasse and hydrochar generated by MAHTC and
conventional HTC from research of Hoekman et al. (2013) are listed in Table 2.

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Table 2. Physicochemical properties of sugarcane bagasse and generated hydrochar

		(dry mass)						
	Content	e ————————————————————————————————————						
	(%)	С	Ν	0	Н	HHV (IVIJ/Kg)	ERE (%)	
Sugarcane bagasse	50.3	44.2	0.4	35.0	6.4	17.9	-	
Hydrochar MAHTC, 240°C, 30 min	-	61.4	0.9	18.4	5.3	24.4	84.8	
Hydrochar Conventional HTC, 255°C, 30 min	-	62.3	1.1	18.5	4.7	23.4	56.33	

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213 According to Bermúdez et al. (2015), when the microwave assisted process is scaled-up from 5 g to 214 100 g, there is a decrease in specific energy consumption of 90% to 95% with both water heating and 215 carbon heating methods and will further decrease with lager capacity on the basis of model 216 prediction. In this study we assume the material input per unit is assumed to be not affected by 217 scaling-up while the energy input per unit is assumed to decrease accordingly. The energy 218 consumption would therefore decrease 95% per kg with increasing capacity from 10 g bagasse at 219 lab-scale to 200 kg per batch for industrial plant. The comparison between parameters of lab-scale 220 batch equipment and the hypothetical scale-up plant is listed in Table 3.

	Lab-scale	Scale up
Batch Capacity	0.1 kg (dry mass)	200 kg (wet mass)
Deionized water	300.0 mL	2.9 m ³
Hydrochar yield		62.5%
Electricity for microwave heating	414 MJ/kg	20.7 MJ/kg

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223 The liquid phase after MAHTC treatment can be easily separated by gravity drainage. Several organic 224 compounds in liquid phase were detected including various sugars and organic acids, where 5-HMF 225 and furfural were detected as main products by GC-MS analysis as the degradation products of 226 sugars. Although it has been suggested that the liquid phase could be utilized by recovery of valuable 227 organic chemicals, without practical extraction method such a complex mixture would possess no economic value. The liquid phase is therefore usually regarded as waste aqueous phase that should 228 229 be treated before discharge (Pala et al., 2014). According to our experimental results, the liquid 230 phase still represents more than 25% energy content of initial feedstock at selected condition. 231 Whether to take the liquid phase as by-products or wastewater is therefore depends on the chosen 232 allocation method, which has been discussed in section 2.1.3. 233 The composition of gas phase was not detected since less than 5% of carbon in sugarcane bagasse 234 has been transferred into gas phase according to our experimental result. It is well known that CO₂ is 235 the dominant gaseous species for most biomass waste (Hoekman et al., 2011), which accounts for 236 over 95% of the total gases quantified from HTC treatment of sugarcane bagasse (2013). In this 237 study, the gas phase discharged into atmosphere from HTC process in this study is assumed to be 238 CO₂ emission.

239 2.2.1.3 Hydrochar combustion

In the recent research of Liu et al. (2017), a boiler was used for co-combustion of hydrochar and coal
fines at different mixing ratios. The energy consumption data of the boiler is adopted in this study
under the circumstance that hydrochar being used as the only fuel source. The gaseous pollutant
emission from 1 kg hydrochar combustion is assumed as the same amount of the emission from 1 kg
MSW incineration since combustion of hydrochar could resemble the same process of municipal
solid waste (MSW) according to Berge et al. (2015).

246 2.2.1.4 Transportation

247 Collection and transport distance of sugarcane from field to the milling plant is calculated following the logistics of Wang et al. (2014) according to the average haul distance equation (1), where C_p (10³ 248 t of sugarcane crushed per year) is 4000 as the assumed capacity of milling plant and Y is the 249 250 sugarcane yield with a value of 60.9% as the average sugarcane yield in china from 2004 to 2013 (Li 251 and Yang, 2015). Other relative parameters in the equation are adopted from the mentioned 252 reference, including α (the proportion of distillery catchment area covered by sugarcane), σ (the 253 actual distance travelled to the straight-line distance) and n (the harvest area would be constrained 254 to 1/n as a result of geography).

$$d = 2 \times \sigma \times \frac{2}{3} \sqrt{\frac{n \times C_p}{100 \times \alpha \pi Y}}$$
(1)

The average distance to waste treatment facility in Ecolnvent database is 10 km (Milà i Canals et al., 2007) which is also adopted in the current research as the distance from sugarcane milling plant to MAHTC plant. For the transportation distance of hydrochar briquette to the combustion plant by train is assumed as 160.93 km (100 miles) as suggested by Liu et al (2017).

260 2.2.2 Life cycle inventory

Table 4 summarizes all the material flow data in the system boundary referring to 1 MJ electricity
 generation from sugarcane bagasse hydrochar.

	Value	Units	Data source
Sugarcane milling			
Input			
Sugarcane (wet mass)	0.653	kg	(Ensinas et al., 2007; Gnansounou et al., 2015; Nguyen et al., 2008; Ramjeawon, 2008)
Transport sugarcane by agricultural tractor and trailer	0.015	tkm	Calculated based equation (Wang et al., 2014)
Tap water (used for washing)	0.245	kg	(Gnansounou et al., 2015)
Tap water (used for imbibition)	0.042	kg	(Gnansounou et al., 2015)
Electricity from natural gas	0.059	MJ	(Nguyen et al., 2008)
Steam consumption	0.163	kg	(Nguyen et al., 2008)
Lime hydrated	0.001	kg	(Gnansounou et al., 2015)
MAHTC of sugarcane bagasse			
Input			
Deionized water for MAHTC reaction	2.842	kg	Experiment result
Transport sugarcane bagasse by unspecified lorry	0.002	tkm	Ecolnvent database
Electricity from natural gas for MAHTC reaction	0.749	MJ	Experiment result and calculation based on equation (Bermúdez et al., 2015)
Electricity from natural gas for briquette production	0.341	MJ	(Liu et al., 2017)
Output			
Aqueous phase	0.003	m³	Experiment result
CO ₂ emission	0.003	kg	Experiment result
Combustion of hydrochar briquette			
Input			
Transport briquette to combustion plant using train	0.042	tkm	(Liu et al., 2017)

	Electricity from natural gas for heater	0.073	MJ	(Liu et al., 2017)
	Output			
	Gaseous pollutants emission	0.051	kg	Data of MSW incineration process from EcoInvent database
264				

265 2.3 Impact assessment and interpretation

Five environmental impact categories, i.e. Climate Change (as Global warming Potential, GWP),
Freshwater Eutrophication (FEP), Freshwater Ecotoxicity (ET), Human Toxicity (HT) and Fossil
Depletion (FD) are evaluated based on the ReCiPe midpoint methodology. The interpretation phase
of this study discusses the LCA results of two scenarios with different allocation methods and
provided critical and feasible advices for technical improvement from environmental perspective.

271 3. LCA results

The LCA results from this study presents the potential environmental impacts of bagasse generation, MAHTC treatment of sugarcane bagasse and hydrochar combustion processes under two scenarios with different allocation factors. For each individual impact, the inputs and outputs have been classified into four categories as emission, transport, fuel, and material to discuss their different influence on causing environmental burdens. Based on LCA results, sensitivity analysis and comparison with other fuel sources are further investigated to identify the influence from key process parameters and substituted energy sources on the results.

LCA results of GWP, FEP, ET, HT and FD impacts under S1 and S2 are listed in Table 5. All five impacts under of evaluated scenarios have resulted in similar value, confirming both allocation methods are suitable from systematic view. The GWP, HT and FD impacts are slightly lower in S2 than S1, as a small portion of impact related with energy consumption initially associated with hydrochar generation would be allocated in liquid by-product when energy content is used for allocation. While the FEP and ET impacts slightly increase, implying increasing amount of freshwater use in S2. The 285 GWP impact is 0.264 and 0.231 kg/MJ CO₂ eq under S1 and S2, both of which are less than half of 286 the value reported by Liu et.al (2017) using conventional HTC treatment (0.149 kg/kwh CO_2 eq for 287 100% hydrochar). Although the former research used wood chips as feedstock, several parameters 288 are similar to the current study including the water content of feedstock and hydrochar properties. 289 This comparison result suggests MAHTC as an environmental-friendly way for fuel production from 290 biomass waste comparing to conventional HTC. To better understand the contribution of 291 environmental burdens from individual process, each environmental impact will be broken down 292 into three main processes and analysed in detail with inputs and outputs classified into four 293 categories.

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Table 5. Impact categories and the corresponding results under with two scenarios

Impact category	Abbreviation	Unit/MJ	S1	S2
Climate change	GWP	kg CO₂ eq	2.64 x 10 ⁻¹	2.31 x 10 ⁻¹
Freshwater eutrophication	FEP	kg P eq	1.13 x 10 ⁻⁵	1.18 x 10 ⁻⁵
Freshwater ecotoxicity	ET	kg 1,4-DB eq	8.13 x 10 ⁻³	8.14 x 10 ⁻³
Human toxicity	HT	kg 1,4-DB eq	2.15	2.01
Fossil depletion	FD	kg oil eq	8.82 x 10 ⁻²	7.57 x 10 ⁻²

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296 4.1 LCA results of Scenario 1

Results from LCA analysis of S1 are presented in Fig.2. The percent contribution of three main
processes in Fig.2(a) – bagasse generation from sugarcane milling, MAHTC of sugarcane bagasse and
hydrochar combustion for electricity generation are compared on each individual impact category.
Sugarcane milling has the least contribution on all the selected impact categories comparing to the
other two processes. For GWP, FD, FEP and HT impacts, the LCA results indicate MAHTC of bagasse
have the most significant roles in causing environmental loads, where ET impact is shown to be
mainly affected by hydrochar combustion process.

In order to investigate the influence of different factors related to each process, the impact potential is presented with absolute value in Fig.2(b)-(f). As shown in Fig.2(b) and (f), the GWP and FD impacts have similar distribution that majority of the burden comes from fuel consumption while a small contribution from hydrochar combustion emission is not neglectable on GWP impact. It is not surprising to find out that MAHTC process represents the largest contribution in these two fuel related categories because of large amount of heat consumption during MAHTC to maintain the system at designed temperature.

311 Two types of freshwater related impacts, i.e. FEP and ET impacts, are illustrated in Fig.2(c) and (d) 312 with large distinction in process distribution. There is a great complexity in the attributed factors in 313 FEP impact. In sugarcane milling process, water use is shown to be the main factor since large 314 amount of water is consumed for steam generation during milling process (Lobo et al., 2007). When 315 it comes to the MAHTC process, material and fuel use represent half contribution with the other half 316 comes from discharge of the liquid product after wastewater treatment. For the last process when 317 hydrochar is combusted for electricity, there is a reasonable finding that gaseous pollutants emission 318 from hydrochar combustion has largest impact on eutrophication. These gaseous pollutants also 319 represent the cause of over 94% ET impact presented in Fig.2(d), indicating the major ET impact is 320 caused by traditional combustion process. Another 3% of ET impact is observed to be caused by fuel 321 consumption during MAHTC treatment, which is originated from electricity generation plant.

The last but the most intuitive category discussed in this research is HT impact as shown in Fig.2(f). Sugarcane milling process represents the least contribution on HT impact with fuel use as the main cause. Unlike the results from Berge et al. (2015), the contribution of MAHTC treatment on human toxicity in this study mainly comes from fuel use rather than water emission since the liquid product is assumed to be treated prior to discharge. The largest HT impact among all the concerned factors is again associated with gaseous pollutants emission from hydrochar combustion process. Since the impact of gas emission from hydrochar combustion is assumed as same as the impact from MSW combustion, it is suggested less impact would associate with real hydrochar combustion condition
 because of relatively simple composition of the combust. The negligible environmental impact
 associated with gaseous pollutants on toxicity related impact categories indicates essential
 requirement of further investigation to quantify different composition in gas emission from
 hydrochar combustion.



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335 Figure 2. LCA results of five environmental impact categories under S1, (a) percentage contribution of different 336 processes and impacts from emission, transportation, fuel, and material use on (b) GWP; (c) FEP; (d) ET; (e) HT and (f) FD

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338 4.2 LCA results of Scenario 2

339 As discussed previously in Table 5, the total environmental impact value of each categories under 340 two scenarios are approximately similar. Yet obvious differences are revealed between these two 341 scenarios as environmental impacts being allocated into each process. Comparing the contribution 342 of three main processes in Fig.3(a) with Fig.2(a), a significant increase of the environmental burden 343 caused by sugarcane milling process (displayed in black) can be observed in S2 with simultaneous 344 decrease in the burden caused by MAHTC process (displayed in dark grey). The change of process 345 distribution in two scenarios is mainly caused by two different allocation methods that have changed 346 impact allocated to different products. The distribution of hydrochar combustion process is almost 347 unaffected since no allocation is involved in this process.

348 In the two main products from sugarcane milling process, bagasse possesses very little economic 349 value comparing to juice but occupies half of the energy content because of its large volume. When 350 the environmental impacts are allocated by energy content in S2, the impact associated with 351 bagasse generation would increase and therefore result in higher contribution of sugarcane milling 352 process than S1. Another process being largely affected by allocation method is MAHTC treatment of 353 bagasse to produce hydrochar. In S1 when the liquid product from MAHTC process is regarded as 354 wastewater, the environmental impact largely associates with treatment and discharge process, 355 while in S2 the liquid phase is regarded as a by-product of MAHTC process with appropriate 356 treatment to retrieve energy. As mentioned before, the liquid product from lab-MAHTC process 357 possesses 25% of total carbon contribution in initial feedstock that cannot be ignored when energy 358 content is used for allocation. It has been discussed in previous research that the recovery of liquid 359 phase is critical for the environmental performance of HTC treatment from both technical (Pala et al., 2014) and environmental (Berge et al., 2015) perspective. Hoekman et al. (2013) suggested a
 multistep HTC process for recovery of valuable organic chemicals to maximize them at different
 temperature ranges, however studies on successful extraction of organic compounds from the
 complex aqueous products are barely published.

364 The impact potential of five categories in S2 related with emission, transportation, fuel and material 365 use contribution in different processes are presented in Fig.3 (b)-(f). The high energy consumption of 366 MAHTC process still makes it as the biggest contribution of GWP and FD impacts, though the 367 increased contribution from sugarcane milling process indicates fuel use during bagasse generation 368 is also an important influential factor on these two impacts. The increasing load represented by 369 impact potential of sugarcane milling process is also observed in FEP, ET and HT impacts. A visible 370 difference in the contribution of FEP impact between two scenarios in Fig.2(c) and Fig.3(c), indicating 371 this impact category is sensitive to process with high amount of water use and wastewater 372 treatment. The increasing environmental impact from water use during bagasse generation makes 373 sugarcane milling process become the dominant contribution for eutrophication impact in this 374 scenario. It also explains the desirable decrease of FEP impact in S2 because of liquid phase recovery 375 from MAHTC process. These results not only indicate the unavoidable environmental impact from 376 fuel use in sugarcane milling process, but also highlight the importance of environmental burdens 377 caused by material use, i.e., water and sugarcane. As for ET and HT, gaseous emission from 378 hydrochar combustion is still playing a determining role in both toxicity related impacts in 379 accordance with the environmental performance of S1.



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Figure 3. LCA results of five environmental impact categories under S2, (a) percentage contribution of different
 processes and impacts from emission, transportation, fuel, and material use on (b) GWP; (c) FEP; (d) ET; (e) HT
 and (f) FD

384 4.3 Sensitivity analysis

385 Sensitivity analysis is conducted with ten potential parameters based on previous analysis to

386 understand how changes of parameters in different processes would influence environmental

- 387 impacts. For sugarcane milling process, bagasse yield and energy use are selected as primary factors
- 388 since both material and fuel use have large impact on the total environmental impact. Transport

distance from sugarcane field to milling plant is also selected because there could be an uncertainty during the calculation of the distance under real circumstance. The influence of energy use during MAHTC treatment on the environmental impact is proved critical to the system, thus electricity use during MAHTC process. Other three parameters related with MAHTC reaction including hydrochar yield, water/solid ratio and carbon content of hydrochar are also selected for sensitivity analysis. For combustion of hydrochar, the emission is evaluated for its essential impact as well as fuel use and railway transportation.

396 Sensitivity ratio (SR) is introduced and calculated to quantify the influential extent of each parameter. 397 It is defined as the percent change of the result divided by the percent change of parameter (Berge 398 et al., 2015). LCA results from S1 is selected as base case since the economic value is more 399 frequently used as the allocation method in biorefinery research and the total environmental impact 400 value from the two scenarios are similar. For all the selected parameters, simulations are conducted 401 in 20% increasing and 20% decreasing of the value in base case for comprehensive analysis (except 402 for water/solid ratio used in MAHTC process, 15/1 and 40/1 are selected as the same one used for 403 test in lab-scale experiment), which are recorded as '+20%' and '-20%' respectively to facilitate the 404 discussion. Fig.4 and Fig.5 illustrate the correlation of three most sensitive parameters (selected with 405 top three absolute value of SR) in each impact category under two circumstances with their 406 corresponding value listed in the table.

2					
1					
-1					
-2	Climate change	Freshwater eutrophication	Freshwater ecotoxicity	Human toxicity	Fossil depletion
■ 1	Carbon content	Carbon content	Carbon content	Carbon content	Carbon content
■ ⊥	-1.608	-1.170	-1.654	-1.624	-1.627
2	Hydrochar yield	Hydrochar yield	Emission to gas from combustion	Hydrochar yield	Hydrochar yield
	-1.107	-1.084	0.888	-0.777	-1.215
3	Fuel use in HTC	Bagasse yield	Hydrochar yield	Fuel use in HTC	Fuel use in HTC
	0.743	-0.264	-0.118	0.488	0.842

408 Figure 4. Top three sensitive factors for each impact category with corresponding sensitivity ratio (+20%)



Figure 5. Top three sensitive factors for each impact category with corresponding sensitivity ratio (-20%)

412 It is not surprising to discover carbon content of hydrochar as most influential factor which has 413 negative correlation on all the five impact categories in this study. With hydrochar being used for 414 electricity generation, the change of carbon content closely relates with the demanded quantity of 415 hydrochar and thus all the parameters associated with hydrochar generation, especially material and 416 energy use in MAHTC process. Gaseous emission from combustion is also highly influenced by the 417 amount of provided hydrochar. Hydrochar yield lists second influential for GWP, FEP, HT and FD 418 impacts in the +20% simulation as shown in Fig.4. It is also observed as second influential factor for 419 GWP, FEP and HT impacts in the -20% simulation as shown in Fig.5. As explained before, there is no 420 doubt with the importance of hydrochar generation related parameters as the critical role of MAHTC 421 process plays in total environmental impact. Results from sensitivity analysis highlight the 422 importance of increasing product yield and carbon content of hydrochar simultaneously during 423 MAHTC treatment, which has been ignored in previous LCA studies since energy consumption is 424 always recognized as the most influential factor in the case of environmental performance. As a 425 combined indicator to evaluate carbon content and hydrochar yield in the same time, ERE is not only 426 an indicator to improve energy properties of hydrochar, but also can be used to evaluate the 427 potential environmental performance of hydrochar for electricity production. 428 As for the electricity use during MAHTC treatment, sensitivity analysis shows that the results are 429 quite sensitive with GWP, HT and FD impacts. GWP and FD represent the impacts from energy use 430 thus are tightly related with the electricity use. However, it should be highly noticed that the HT 431 impact also obtains considerable influences from energy consumption. Another sensitive parameter 432 is gaseous emission from hydrochar combustion as illustrated by the analysis, especially for ET 433 impact. The explanation of this result could be the toxic gas emission from conventional combustion, 434 as the impact of hydrochar combustion is assumed as the same of MSW incineration. Gas emission 435 from hydrochar combustion and bagasse yield respectively list as the third influential factor for FEP 436 impact under +20% and -20% simulation. Discrepancy in results from two simulation scenarios is due 437 to the uneven percentage each process contributed to the total environmental impact.

438 4.4 Comparison with other electricity generation method

The importance of substituted energy sources has been highlighted in previous LCA studies using 439 food waste (Berge et al., 2015) and biomass waste (Owsianiak et al., 2016) as feedstocks when the 440 441 electricity production from hydrochar is used to offset other types of electricity generation. It is 442 critical to understand at what level was the environmental impact caused by electricity generation 443 from hydrochar combustion when compared with other fuel sources. Fig.6 presents the comparison of environmental impact of 1 MJ electricity generation from bagasse hydrochar combustion in two 444 445 scenarios and six other electricity sources with impact potential data adopted from EcoInvent 446 database, including the average high voltage electricity generation (CN, China), electricity from cogeneration of sweet sorghum bagasse (GLO, Global), incineration of MSW (RoW, Rest of the world), 447 448 conventional natural gas power plant (CN), hard coal combustion (CN) and lignite combustion (RoW).



450

Figure 6. Comparison of environmental impacts from different electricity sources

451 As illustrated in Fig.6, electricity production from hydrochar combustion generally causes more 452 environmental burdens in GWP, ET and FD impacts. On the other hand, less impact is caused by hydrochar combustion in FEP and HT categories than other fuel sources. Comparing with combustion 453 of lignite, hard coal, and average high voltage electricity generation, hydrochar combustion results in 454 lower GWP impact, indicating it possesses large potential to substitute traditional coal. The results 455 456 are not ideal comparing with electricity production from sorghum bagasse, MSW and natural gas, 457 suggesting hydrochar for combustion is not competitive with other biomass or equivalent clean fuel 458 sources. This result is consistent with the conclusion of Berge et al. (2015) that electricity generated

459 from hydrochar combustion would result in environmental savings when offsetting coal-based 460 energy sources but biomass. Disappointingly, FD impacts under both S1 and S2 appear to be even 461 larger than that of hard coal and lignite combustion. This observation is mainly caused by the large 462 amount of electricity consumption during heating and reaction period of MAHTC process, which 463 could be the main obstacle for its industrial application.

464 For other three impact categories, a big disappointing fact is the high impact potential of ET impact 465 associated with electricity production from hydrochar combustion. Under both scenarios, this 466 impact is shown to be far larger than electricity generated from other sources except lignite. This 467 result is mainly caused by the assumption that the gaseous pollutants from hydrochar combustion 468 was as same as MSW incineration, which has a great influence on toxicity related environmental 469 impacts. The low value of FEP and HT impacts associated with electricity production from hydrochar 470 generation when compared to other sources indicates the attracting aspects for implementation of 471 MAHTC treatment for fuel production from sugarcane bagasse.

472 **5. Discussion and implications**

473 The LCA results from this study indicates the critical position of MAHTC treatment and hydrochar 474 combustion process of causing environmental burdens in the system boundary, suggesting technical 475 improvement should be done in these two processes when using hydrochar for electricity 476 generation. Although the change of process-related parameters is micro under lab-scale, it could 477 result in variable differences in material balance and energy consumption from the view of life cycle. 478 For the first time, results from the sensitivity analysis have highlighted both solid yield and carbon 479 content of generated hydrochar have adequate critical influences on the environmental 480 performance of HTC treatment. As a combined index for carbon content and hydrochar yield, ERE is 481 thus suggested as useful indicator to minimize potential environmental burdens. In the life cycle 482 analysis of conventional HTC of two-phase olive mill waste, Benavente et al. (2017) also observed 483 desirable environmental benefits being obtained at highest ERE, which was described as "the

percent of energy initially present in the olive mill waste that is recovered in the hydrochar reached
maximum". Our recent research (Zhang et al., 2018) has proved that using microwave heating during
HTC treatment could further enhance both solid yield and energy content of hydrochar and result in
high ERE value, indicating MAHTC could provide attracting improvement in environmental
performance comparing with conventional HTC.

489 The largest portion of environmental burdens associated with electricity generation from bagasse 490 hydrochar still appears to be energy consumption even with microwave assisted heating during HTC 491 treatment. As energy consumption has been proved to decrease with reduction of heating time 492 during MAHTC, the environmental burdens mainly attribute to the energy source being used to 493 generate heat. For conventional HTC, the heat is usually generated from direct combustion of 494 natural gas or other fuel. Though there could be heat loss due to limited transfer efficiency, 495 environmental burdens associated with direct combustion is still less than that from electricity. 496 Microwave dielectric heating requires significantly more energy than conventional heating 497 techniques as a consequence of the comparably low transfer efficiency of magnetrons in converting 498 electrical to microwave energy (Razzaq and Kappe, 2008). When microwave is used as the only 499 energy supply, MAHTC process could result in more environmental load than conventional HTC using 500 electricity or heat as sources. Considering both process enhancement effect and low transfer 501 efficiency of magnetrons associated with microwave heating, an ideal option could be using natural 502 gas as heating source to reach the designed reaction temperature first and then adding microwave 503 as enhancement method to obtain hydrochar with desirable properties.

Substitutional energy is critical for industrial expansion of sugarcane bagasse hydrochar for
electricity generation. Comparing with other energy sources in China, electricity generation from
sugarcane bagasse hydrochar could result in environmental savings when it is used to replace
traditional coal combustion and average electricity generation. Nevertheless, it is not as competitive
as other energy source with less environmental loads such as biomass or natural gas. Water

509 emission from MAHTC treatment remains a big issue for industrial application of MAHTC treatment 510 from environmental perspective. If the liquid phase from MAHTC process could be partially recycled, 511 all the environmental impact categories discussed in the study especially FEP impact can be largely 512 reduced. If the economic value of liquid product could increase with proper recovery, a small portion of impact initially associated with hydrochar generation would be allocated with the final liquid 513 514 product thus results in less impact with hydrocar. As for hydrochar combustion process, further investigations should be conducted with gaseous composition to understand the unneglectable 515 516 toxicity related impacts to benefit the industrial use of hydrochar as alternative fuel.

517 6. Conclusion

518 A comprehensive life cycle assessment (LCA) has been conducted in this study to evaluate 519 environmental impacts associated with microwave assisted hydrothermal carbonization (MAHTC) of 520 sugarcane bagasse and electricity generation from subsequent hydrochar to provide practical 521 advices for future optimization of MAHTC treatment. Both economic benefit and energy content are 522 suitable for allocation with slightly difference in total environmental impacts. Comparing the two 523 scenarios discussed in this article, the Climate change, Human toxicity, and Fossil depletion impacts 524 are slightly lower when energy content was used for allocation, as a small portion of impact initially 525 associated with hydrochar generation would be allocated in liquid product. Results indicate MAHTC 526 as an environmental-friendly treatment for fuel production from biomass waste comparing to 527 conventional hydrothermal carbonization in the case of Climate change impact for both scenarios. 528 Environmental burdens associated with electricity generated from hydrochar combustion are mainly 529 attributed liquid phase treatment and fuel use from MAHTC treatment process and gaseous 530 emission from hydrochar combustions. Comparing with other electricity sources in China, hydrochar 531 combustion could result in environmental savings when substitutes traditional coal and average 532 electricity generation. It is important to note, however, using hydrochar for fuel is not as competitive 533 as biomass and natural gas.

534 LCA Results from this study has successfully established linkage between the environmental 535 performance and the process parameters from MAHTC treatment when hydrochar is used for fuel 536 production. As suggested by sensitivity analysis, carbon content and hydrochar yield have 537 neglectable influences on environmental performance that have been overlooked previously. Both 538 hydrochar properties and environmental performance can be optimized with high energy retention 539 efficiency indicator which can be effectively improved with assistance of microwave heating. MAHTC 540 is suggested as an environmentally friendly way for fuel production from biomass waste like 541 sugarcane bagasse to substitute traditional coal. It is recommended that multistep treatment for 542 recovery of valuable organic chemicals from liquid phase and mixed heating method are favourable 543 to reduce associated environmental impacts for future optimizations of hydrothermal carbonization 544 process.

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552 References:

1. Ahmed, I.I., Gupta, A.K., 2012. Sugarcane bagasse gasification: Global reaction mechanism of syngas

554 evolution. Applied Energy 91 (1), 75–81. doi.org/10.1016/j.apenergy.2011.07.001

555 2. Bajpai, P., 2016. Pretreatment of Lignocellulosic Biomass, in: Pretreatment of Lignocellulosic Biomass for

556 Biofuel Production. Springer Singapore, Singapore, pp. 17–70. doi.org/10.1007/978-981-10-0687-6

- 557 3. Benavente, V., Fullana, A., Berge, N.D., 2017. Life cycle analysis of hydrothermal carbonization of olive
- 558 mill waste: Comparison with current management approaches. Journal of Cleaner Production 142, 2637–
- 559 2648. doi.org/10.1016/j.jclepro.2016.11.013
- 560 4. Berge, N.D., Li, L., Flora, J.R.V., Ro, K.S., 2015. Assessing the environmental impact of energy production
- 561 from hydrochar generated via hydrothermal carbonization of food wastes. Waste Management 43, 203–
- 562 217. doi.org/10.1016/j.wasman.2015.04.029
- 5. Berge, N.D., Ro, K.S., Mao, J., Flora, J.R. V, Chappell, M.A., Bae, S., 2011. Hydrothermal Carbonization of
- 564 Municipal Waste Streams. Environmental Science & Technology 45 (13), 5696–5703.
- 565 doi.org/10.1021/es2004528
- 566 6. Bermúdez, J.M., Beneroso, D., Rey-Raap, N., Arenillas, A., Menéndez, J.A., 2015. Energy consumption
- 567 estimation in the scaling-up of microwave heating processes. Chemical Engineering and Processing:
- 568 Process Intensification 95, 1–8. doi.org/10.1016/j.cep.2015.05.001
- 569 7. Borrion, A.L., McManus, M.C., Hammond, G.P., 2012. Environmental life cycle assessment of bioethanol
 570 production from wheat straw. Biomass and Bioenergy 47, 9–19. doi.org/10.1016/j.biombioe.2012.10.017
- 571 8. Chauhan, M.K., Varun, Chaudhary, S., Kumar, S., Samar, 2011. Life cycle assessment of sugar industry: A
- 572 review. Renewable and Sustainable Energy Reviews 15 (7), 3445–3453.
- 573 doi.org/10.1016/j.rser.2011.04.033
- 574 9. Chen, W.H., Ye, S.C., Sheen, H.K., 2012. Hydrothermal carbonization of sugarcane bagasse via wet
- 575 torrefaction in association with microwave heating. Bioresource Technology 118, 195–203.
- 576 doi.org/10.1016/J.BIORTECH.2012.04.101
- 577 10. Elaigwu, S.E., Greenway, G.M., 2016a. Microwave-assisted and conventional hydrothermal carbonization
- 578 of lignocellulosic waste material: Comparison of the chemical and structural properties of the hydrochars.
- 579 Journal of Analytical and Applied Pyrolysis 118, 1–8. doi.org/10.1016/J.JAAP.2015.12.013

- 580 11. Elaigwu, S.E., Greenway, G.M., 2016b. Microwave-assisted hydrothermal carbonization of rapeseed husk:
- 581 A strategy for improving its solid fuel properties. Fuel Processing Technology 149, 305–312.
- 582 doi.org/10.1016/J.FUPROC.2016.04.030
- 583 12. Ensinas, A. V., Nebra, S.A., Lozano, M.A., Serra, L.M., 2007. Analysis of process steam demand reduction
- 584 and electricity generation in sugar and ethanol production from sugarcane. Energy Conversion and
- 585 Management 48 (11), 2978–2987. doi.org/10.1016/j.enconman.2007.06.038
- 586 13. Erlich, C., Öhman, M., Björnbom, E., Fransson, T.H., 2005. Thermochemical characteristics of sugar cane
 587 bagasse pellets. Fuel 84 (5), 569–575. doi.org/10.1016/j.fuel.2004.10.005
- 588 14. Finkbeiner, M., Inaba, A., Tan, R.B.H., Christiansen, K., Klüppel, H.J., 2006. The New International
- 589 Standards for Life Cycle Assessment: ISO 14040 and ISO 14044. The International Journal of Life Cycle
- 590 Assessment 11 (2), 80–85. doi.org/10.1065/lca2006.02.002
- 591 15. Gnansounou, E., Vaskan, P., Pachón, E.R., 2015. Comparative techno-economic assessment and LCA of
- selected integrated sugarcane-based biorefineries. Bioresource Technology 196, 364–375.
- 593 doi.org/10.1016/j.biortech.2015.07.072
- Hoekman, S.K., Broch, A., Robbins, C., 2011. Hydrothermal Carbonization (HTC) of Lignocellulosic Biomass.
 Energy & Fuels 25 (4), 1802–1810. doi.org/10.1021/ef101745n
- 17. Hoekman, S.K., Broch, A., Robbins, C., Zielinska, B., Felix, L., 2013. Hydrothermal carbonization (HTC) of
- selected woody and herbaceous biomass feedstocks. Biomass Conversion and Biorefinery 3 (2), 113–126.
- 598 doi.org/10.1007/s13399-012-0066-y
- 599 18. HTCycle, 2019. Company, HTC panorama. http://htcycle.ag/en/htc-panorama_44 (Accessed 01
 600 September 19)
- 19. HTCycle, 2018. A Chinese delegation of executives visits HTCycle. https://htcycle.ag/en/article/a-chinese-
- 602 delegation-of-executives-visits-htcycle_30 (Accessed 01 September 19)
- 603 20. Karka, P., Papadokonstantakis, S., Hungerbühler, K., Kokossis, A., 2015. Life Cycle Assessment of
- 604 Biorefinery Products Based on Different Allocation Approaches. Computer Aided Chemical Engineering 37,
- 605 2573–2578. doi.org/10.1016/B978-0-444-63576-1.50123-0
- 606 21. Kim, D., Lee, K., Park, K.Y., 2016. Upgrading the characteristics of biochar from cellulose, lignin, and xylan
- 607 for solid biofuel production from biomass by hydrothermal carbonization. Journal of Industrial and
- 608 Engineering Chemistry 42, 95–100. doi.org/10.1016/j.jiec.2016.07.037

22. Li, Y.R., Yang, L.T., 2015. Sugarcane Agriculture and Sugar Industry in China. Sugar Tech 17, 1–8.

610 doi.org/10.1007/s12355-014-0342-1

- 611 23. Liu, X., Hoekman, S., Farthing, W., Felix, L., 2017. TC2015: Life cycle analysis of co-formed coal fines and
- 612 hydrochar produced in twin-screw extruder (TSE). Environmental Progress & Sustainable Energy 36, 668-
- 613 676. doi.org/10.1002/ep.12552
- 614 24. Lobo, P.C., Jaguaribe, E.F., Rodrigues, J., da Rocha, F.A.A., 2007. Economics of alternative sugar cane
- 615 milling options. Applied Thermal Engineering 27 (8-9), 1405–1413.
- 616 doi.org/10.1016/j.applthermaleng.2006.10.023
- 617 25. Lu, X., Pellechia, P.J., Flora, J.R.V., Berge, N.D., 2013. Influence of reaction time and temperature on
- 618 product formation and characteristics associated with the hydrothermal carbonization of cellulose.

619 Bioresource Technology 138, 180–190. doi.org/10.1016/j.biortech.2013.03.163

- 620 26. Mashoko, L., Mbohwa, C., Thomas, V.M., 2010. LCA of the South African sugar industry. Journal of
- 621 Environmental Planning and Management 53 (6), 793–807. doi.org/10.1080/09640568.2010.488120
- 622 27. Melo, C.A., Junior, F.H.S., Bisinoti, M.C., Moreira, A.B., Ferreira, O.P., 2017. Transforming Sugarcane
- 623 Bagasse and Vinasse Wastes into Hydrochar in the Presence of Phosphoric Acid: An Evaluation of Nutrient

624 Contents and Structural Properties. Waste and Biomass Valorization 8 (4), 1139–1151.

625 doi.org/10.1007/s12649-016-9664-4

- 626 28. Milà i Canals, L., McLaren, S., Muñoz, I., Miguel, B., 2007. LCA Methodology and Modelling Considerations
 627 for Vegetable Production and Consumption. CES Working Papers 02/07. (Accessed 01 March 18)
- 628 29. Nguyen, T.L.T., Gheewala, S.H., Garivait, S., 2008. Full chain energy analysis of fuel ethanol from cane

629 molasses in Thailand. Applied Energy 85 (8), 722–734. doi.org/10.1016/j.apenergy.2008.02.002

- 630 30. Owsianiak, M., Ryberg, M.W., Renz, M., Hitzl, M., Hauschild, M.Z., 2016. Environmental Performance of
- 631 Hydrothermal Carbonization of Four Wet Biomass Waste Streams at Industry-Relevant Scales. ACS
- 632 Sustainable Chemistry & Engineering 4 (12), 6783–6791. doi.org/10.1021/acssuschemeng.6b01732
- 633 31. Pala, M., Kantarli, I.C., Buyukisik, H.B., Yanik, J., 2014. Hydrothermal carbonization and torrefaction of
- 634 grape pomace: A comparative evaluation. Bioresource Technology 161, 255–262.

635 doi.org/10.1016/j.biortech.2014.03.052

- 636 32. Peng, L., Jackson, P.A., Li, Q. wei, Deng, H. hua, 2014. Potential for Bioenergy Production from Sugarcane
- 637 in China. Bioenergy Research 7 (3), 1045–1059. doi.org/10.1007/s12155-013-9403-7

- 638 33. Ramjeawon, T., 2008. Life cycle assessment of electricity generation from bagasse in Mauritius. Journal of
- 639 Cleaner Production 16, 1727–1734. doi.org/10.1016/j.jclepro.2007.11.001
- 640 34. Razzaq, T., Kappe, C.O., 2008. On the Energy Efficiency of Microwave-Assisted Organic Reactions.
- 641 ChemSusChem 1, 123–132. doi.org/10.1002/cssc.200700036
- 642 35. Renó, M.L.G., Lora, E.E.S., Palacio, J.C.E., Venturini, O.J., Buchgeister, J., Almazan, O., 2011. A LCA (life
- 643 cycle assessment) of the methanol production from sugarcane bagasse. Energy 36 (6), 3716–3726.
- 644 doi.org/10.1016/j.energy.2010.12.010
- 64536.Renouf, M.A., Wegener, M.K., Pagan, R.J., 2010. Life cycle assessment of Australian sugarcane production
- 646 with a focus on sugarcane growing. The International Journal of Life Cycle Assessment 15 (9), 927–937.
- 647 doi.org/10.1007/s11367-010-0226-x
- 648 37. Sindhu, R., Gnansounou, E., Binod, P., Pandey, A., 2016. Bioconversion of sugarcane crop residue for
- 649 value added products An overview. Renewable Energy 98, 203–215.
- 650 doi.org/10.1016/j.renene.2016.02.057
- 38. Wang, L., Quiceno, R., Price, C., Malpas, R., Woods, J., 2014. Economic and GHG emissions analyses for
- 652 sugarcane ethanol in Brazil: Looking forward. Renewable and Sustainable Energy Reviews 40, 571–582.
- 653 doi.org/10.1016/j.rser.2014.07.212
- 654 39. Wang, T., Zhai, Y., Zhu, Y., Li, C., Zeng, G., 2018. A review of the hydrothermal carbonization of biomass

655 waste for hydrochar formation: Process conditions, fundamentals, and physicochemical properties.

656 Renewable and Sustainable Energy Reviews 90, 223–247. doi.org/10.1016/j.rser.2018.03.071

- 40. Wright, M., Lima, I., Bigner, R., 2016. Microbial and physicochemical properties of sugarcane bagasse for
- potential conversion to value-added products. International Sugar Journal 118 (1410), 10-18.

659 https://www.researchgate.net/publication/305005860_Microbial_and_physicochemical_properties_of_s

- 660 ugarcane_bagasse_for_potential_conversion_to_value-added_products. (Accessed 24 April 18)
- 41. Xiao, L.P., Shi, Z.J., Xu, F., Sun, R.C., 2012. Hydrothermal carbonization of lignocellulosic biomass.
- 662 Bioresource Technology 118, 619–623. doi.org/10.1016/j.biortech.2012.05.060
- 42. Zhang, J., An, Y., Borrion, A., He, W., Wang, N., Chen, Y., Li, G., 2018. Process characteristics for
- 664 microwave assisted hydrothermal carbonization of cellulose. Bioresource Technology 259.
- 665 doi.org/10.1016/j.biortech.2018.03.010

666