

In-cylinder Polycyclic Aromatic Hydrocarbons (PAHs) Sampled during Diesel Engine Combustion

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Abstract

Polycyclic aromatic hydrocarbons (PAH) are potentially carcinogenic pollutants emitted by diesel engines, both in the gas-phase and adsorbed onto the surface of particulate matter (PM). There remains limited understanding of the complex and dynamic competing mechanisms of PAH formation, growth and oxidation in the gas-phase and their adsorption onto soot, and how these processes impact on the abundance and composition of exhaust PAH. Therefore, this paper presents analysis of gas and particulate samples taken from the cylinder and exhaust of a diesel engine during combustion of fossil diesel with the 16 US-EPA priority PAH species identified and quantified.

In-cylinder results showed that gas phase PAHs were more abundant than soot-bound PAHs in the engine cylinder. The in-cylinder PAHs included 2- to 6-ring PAHs, however, 6-ring PAHs were not observed in the soot samples collected from the engine exhaust. Levels of both PM and the total in-cylinder PAHs decreased following a peak at 10 CAD ATDC, but subsequently increased significantly during the late combustion phase. The B[a]P equivalence of PM in the engine cylinder increased during the period of early diffusion to late combustion phase, following an initial decrease during the period of premixed to early diffusion combustion.

Keywords : Polycyclic aromatic hydrocarbon (PAH), fossil diesel combustion, particulates toxicity, PAH growth mechanisms, US-EPA priority PAH, gas phases PAH, soot-bound PAH, in-cylinder sampling, exhaust sampling

1. Introduction

Internal combustion (IC) engines used in road transportation continue to be a major anthropogenic source of both greenhouse gas emissions and pollutants that impact on air quality (1) (2). Combustion processes in compression-ignition and spark-ignition engines produce and emit gaseous pollutants including NO_x, CO, total hydrocarbons and also soot, the permissible levels of which are strictly regulated (1) (3) (4). Soot particles also carry substances through adsorption onto the particle surface, including polycyclic aromatic hydrocarbons (PAHs) which are known mutagens and are of immense concern because of their high toxicity to human health (5) (6) (7) (8) (9) (10) (11). Over 32 PAH species have been found in diesel engine emissions, but 16 of these species were selected by the US Environmental Protection Agency (US-EPA) as priority pollutants primarily on the basis of their higher toxicity potential, their possibility of human exposure and environmental occurrence, therefore, it is important to measure them. Seven of the 16 PAHs have been deemed possible human carcinogens and have been underscored in Table S 1 (see supporting information) (11) (12) (13). While many researchers have extensively studied the emission of particulate matter (PM) and the 16 PAH species prioritized by the US-EPA in the exhaust of a diesel engines fuelled with 100% fossil diesel (14) (15) (16) (17) (18) (19), very few studies have provided insights into the composition and abundance of these exhaust emissions by way of investigating PAH inside the engine cylinder during the course of combustion. An early study of fossil diesel combustion undertaken by Barbella R, Bertoli C, Ciajolo A, D'anna A (20) using in-cylinder samples reported that soot concentration and PAH fraction in weight per weight (%wt/wt) increased after fuel injection to a peak at around 10 CAD ATDC (indicating increased soot formation rates) and decreased afterwards during the remainder of combustion (due to oxidation), in agreement with the findings of Malmberg VB, Eriksson AC, Shen M, Nilsson P, Gallo Y, Waldheim B, Martinsson J, Andersson O, Pagels J (21). PAH species present in the fossil diesel fuel were compared by Barbella R, Bertoli C, Ciajolo A, D'anna A with the PAH fraction of all samples collected from the cylinder and they were found to be identical, indicating that no combustion-formed PAH species were detected. The authors attributed this observation to the faster formation kinetics of these species relative to the temporal resolution of the sampling valve used (20).

Later experimental in-cylinder studies have instead utilised single-component fuels as a surrogate for diesel (22) (23), with the addition of an aromatic to replicate fossil diesel combustion. One such study by Narushima T, Morishima A, Moriwaki H, Kusaka J, Daisho Y (23) reported that during *n*-heptane combustion the level of in-cylinder gaseous cyclic compounds, including PAHs with less than five rings, was reduced as the air-fuel mixture became increasingly lean, due to oxidation. Wang X, Song C, Lv G, Song J, Li H, Li B (22) also found from in-cylinder studies that the addition of toluene to *n*-heptane promoted soot formation, reduced the total EPA PAH level, but increased the amounts of PAHs that were equal in size or larger than Phenanthrene (PHE) relative to that found during *n*-heptane combustion. Furthermore, Wang X, Song C, Lv G, Song J, Li H, Li B found that naphthalene (NAP) was the most abundant PAH present at all sampling times during the combustion of *n*-heptane and *n*-heptane-toluene blend, with three-ring PAHs, acenaphthene (ACN), acenaphthylene (ACNY), and fluorene (FLU) also present in significant quantities. Like Wang X, Song C, Lv G, Song J, Li H, Li B, Ciajolo A, D'Anna A, Barbella R, Bertoli C (24) also studied the influence of fuel aromaticity on in-cylinder PAHs, but by the addition of α -Methylnaphthalene to *n*-tetradecane. While the addition of aromatics to the diesel surrogate fuel was found to promote soot formation in both studies, Wang X, Song C, Lv G, Song J, Li H, Li B observed a decrease in the total mass of in-cylinder PAHs in contrast to an increase observed by Ciajolo A, D'Anna A, Barbella R, Bertoli C. Of these previous in-cylinder studies, only two utilised fossil diesel, and the discordant results seen in the utilisation of diesel surrogates makes further investigation of fossil diesel itself imperative, because fossil diesel fuel is known to contain various hydrocarbons including paraffins, olefins, cycloparaffins and aromatics (25) which are absent in single-component surrogates and can therefore significantly impact on the PAH process. In the study by Barbella R, Bertoli C, Ciajolo A, D'anna A, quantitative analysis of individual PAH species was not undertaken, and so changing levels of total and individual PAH species during combustion was not investigated. The authors utilized ice traps to condense heavy hydrocarbons during

104 samples collection, and gaseous PAH species might have condensed during this process, thus
105 making it difficult to ascertain the separate proportions in the gaseous and solid PAH phases. The
106 diesel fuel utilised by Barbella R, Bertoli C, Ciajolo A, D'anna A contained an excessively high PAH
107 content (33.6 %wt), which is no longer representative of the PAH content of modern diesel fuel,
108 now typically below 5% (26). In the study undertaken by Malmborg VB, Eriksson AC, Shen M,
109 Nilsson P, Gallo Y, Waldheim B, Martinsson J, Andersson O, Pagels J, though quantitative, only
110 soot-bound PAH, comprised majorly of 4- to 6-ring PAH species, were investigated.

111 PAHs are known precursors to formation of soot particles in engines following pyrolysis of
112 injected fuel (27), and their levels during combustion are influenced by competing and dynamic
113 mechanisms of formation and consumption, either through growth to larger species or oxidation.
114 The evolution of gaseous and soot-bound in-cylinder PAHs during fossil diesel combustion in the
115 cylinders of modern direct injection diesel engines and their relationship to the PAH concentrations
116 in the exhaust are still not well understood. There has not previously been such a complete study
117 of analysis of PAHs in both gas and particulate phases. In this paper, analysis of PM and PAHs in
118 samples of gas and PM extracted from the engine cylinder and exhaust gases is presented, with
119 individual PAH levels in the gas- and soot-bound phases quantified. Further, the impact of the PAH
120 levels on the PAH toxicity (B(a)P equivalence) is quantified and discussed.

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124 2. Experimental Approach

125 2.1. Experimental Setup

126 Please note that Table S 1 - Table S 6 and Figure S 1 - Figure S 9 are shown in the supporting
127 information. Figure S 1 shows a schematic diagram of the experimental facility, utilised for all the
128 engine tests reported in this study, which was comprised of a direct injection compression-ignition
129 research engine (see Table S 2), a common fuel rail system which supplied diesel to the engine at
130 the injection pressure, and particulate matter and gaseous PAH collection systems (filter holder
131 housing and resin holder housing) located 0.65m from the exhaust valves. The particulate
132 collection system comprised separate filter and resin holder housings, installed in series, into which
133 the filter and XAD-2 resin were housed, respectively. A vacuum pump drew sample gas through
134 the filter and resin and a gas meter at the exit of the vacuum pump measured the cumulative flow
135 of sample gas. A fast-acting in-cylinder gas sampling valve (IGSV), which was developed and
136 extensively tested by Talibi M, Hellier P, Balachandran R, Ladommatos N (28) to determine
137 consistency of operation and define various operational parameters, including valve opening and
138 closing instants, consistency of sampling windows, penetration of poppet valve into the in-cylinder
139 charge well beyond the boundary layer etc, enabled the extraction and sampling of the contents
140 (PM and combustion-generated gases) present in the combustion chamber (further details of is
141 the IGSV are presented in Section 6.3 of supporting information). In-cylinder sampling was
142 undertaken at four points during the combustion cycle namely 10 CAD ATDC (post premixed
143 combustion dominated), 25 CAD ATDC (diffusion combustion dominated), 40 CAD ATDC (late
144 diffusion combustion) and 55 CAD ATDC (late combustion). In order to collect sufficient gas and
145 particulate sample for analysis, the sampling valve was operated once every three engine cycles
146 until sufficient large samples accumulated for subsequent analysis. The filter holder housing, resin
147 holder housing and all the ¼" stainless-steel sampling pipes to both the IGSV and exhaust pipe
148 were heated by insulated PID-controlled (Proportional Integral Derivative) electric tape heaters
149 maintained at 250 °C to prevent condensation of water vapour and gaseous hydrocarbons. No
150 exhaust after-treatment devices were utilised, and the experiments were conducted without
151 exhaust gas recirculation (EGR).

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2.2. Engine Test Procedure and Conditions

For the collection of in-cylinder and exhaust PM, glass microfibre filter papers were chosen due to the absence of PAH in the filter material, heat resistance properties, wide-spread use in past studies, and recommended use in Method TO-13A provided by the US Environmental Protection Agency (11) (29) (30) (31) (32) (33). Fisher Scientific UK brand glass microfibre filters ($\varnothing = 70$ mm, $0.7\mu\text{m}$ pore size, 75g/m^2 mass, 310 s filtration speed) were used during all engine tests. To prevent the filter from tearing, breaking or deteriorating during sampling as a result of the engine exhaust flow and pressure pulsations, it was sandwiched between two fine, stainless-steel 316 woven wire meshes (aperture = 0.026 mm, 0.025 mm wire diameter (TheMeshCompany U.K) cut to the same diameter as the filter. Both halves of the filter holder firmly gripped the sandwiched filter, thus securing it in place for the duration of sampling. See Section 6.4 of supporting information for further details of the engine warm-up and sampling procedure.

The engine operating conditions for all tests are as follows:

- Engine load (IMEP): 7 bar
- Sample collection duration: 15 mins
- Start of injection: 8.8 CAD BTDC
- End of injection: 2.1 CAD BTDC
- Start of combustion: TDC
- Injection pressure: 450 bar
- Engine speed: 1200 rpm
- Injection duration: $925\mu\text{s}$ (< 6.7 CAD)

Table 1 shows the in-cylinder sampling timings during combustion, the duration of sampling, the in-cylinder pressure, and the total free volume of gas sampled and measured under atmospheric conditions using a gas meter at each of the in-cylinder sampling timings. Sampling timings were chosen to encompass the whole of post premixed combustion. A vacuum pump operating at a flow rate of 43L/min was used to sample from the engine exhaust while the sampling valve gas flow rate from the engine combustion chamber was significantly lower at 1-2 L/min.

Table 1: The in-cylinder sampling timings, duration of sampling windows, in-cylinder gas pressure and sample volume collected for each of the in-cylinder sampling timings

In-cylinder sampling timings (midway of sampling window)	IGSV opening window (CAD)	In-cylinder pressure during sampling window (bar)	Global in-cylinder temperature (K)	Total sample volume (L)
10 CAD ATDC	6	61.8	1222.1	17.95
25 CAD ATDC	10	42.7	1482.5	27.45
40 CAD ATDC	15	22.9	1403.1	19.35
55 CAD ATDC	15	13.3	1277.6	12.2
Exhaust	-	-	-	270

2.3. Fossil diesel properties

Table S 3 shows the fuel properties of the fossil diesel utilised in all engine tests reported in this study. To identify the individual EPA PAH species present in liquid fossil diesel prior to combustion, the total PAH content of fossil diesel was also speciated, and the results are shown in this table.

2.4. Soot-bound and gas phase PAH analysis procedure

All sandwiched filters utilised for collecting particulate matter were desiccated for 12 hours (including supporting mesh) to remove any water accrued due to sampling so that any change in mass of the filter was not attributable to moisture content. The filters were subsequently weighed to determine their new mass (sandwiched filter + particulate), and the gravimetric mass of particulate matter collected was the difference between this new mass and that recorded prior to sampling.

193 2.4.1. Accelerated Solvent Extraction (ASE) system for PAH extraction

194 After the mass of particulate matter collected was determined, PAH species were extracted
195 using dichloromethane solvent from all used sandwiched filters and from the XAD-2 resin, referred
196 to as soot-bound PAHs and gas-phase PAHs respectively throughout this study. The extraction
197 used an accelerated solvent extraction (ASE) system previously optimised by Dandajeh HA,
198 Ladommatos N, Hellier P, Eveleigh A (30) (34) and further described in Section 6.6 (see supporting
199 information). The volume of the extract (solvent + PAH) in the collection bottle at the end of one
200 extraction process was typically between 20 - 30 ml. The PAH extraction method was repeated
201 three times for each engine sample to ensure complete recovery of PAH species, and the three
202 extracts from the three repeat extraction processes of the same engine sample were mixed
203 together. Table S 4 shows the optimised ASE extraction conditions utilised for PAH recovery from
204 all samples collected from the engine. Further steps on the concentration and preparations of ASE
205 extracts are also described in Section 6.6.
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207 2.5. Gas Chromatography-Mass Spectrometer (GCMS) Setup

208 Table S 5 (supporting section) summarizes the operating parameters of the gas
209 chromatograph-mass spectrometer (GCMS) instrument utilized for identifying and quantifying PAH
210 species. Similar to several past studies (14) (35) (22) (11) (32) (36), the GCMS was utilised in SIM
211 mode.
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214 2.6. Quantification of PAH present in the ASE extracts of all engine samples

215 Each deuterated PAH species in the internal standard (IS) was assigned to one or more of
216 the 16 EPA PAH species, as recommended by the EPA (29) and shown in Table S 6. Calibration
217 curves (ratio of PAH concentration to that of assigned IS versus ratio of peak areas of PAH to IS)
218 were plotted for each PAH species from which equations of line of best fit were generated. The
219 concentrations of each PAH species in the engine samples extract were determined using its
220 equation of line of best fit, ratio of peak areas recorded during GCMS run and the final
221 concentration of the assigned internal standard species (0.89 ppm). B[a]P concentration was
222 calculated, using Eqn. 1, as the sum of the products of each PAH species' B[a]P toxicity equivalent
223 factor (TEF) (30) and its concentration, for all 16 PAHs quantified. The TEF indicates PAH toxicity
224 relative to B[a]P and values were taken from the work of Nisbet and LaGoy (37).
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$$B[a]P_{equivalence} = \sum_{i=1}^{16} (TEF \times PAH \text{ concentration}) \quad \text{Eqn. 1}$$

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3. Results

This results of this study are presented in this section while the discussion of the results are presented later in Section 4.

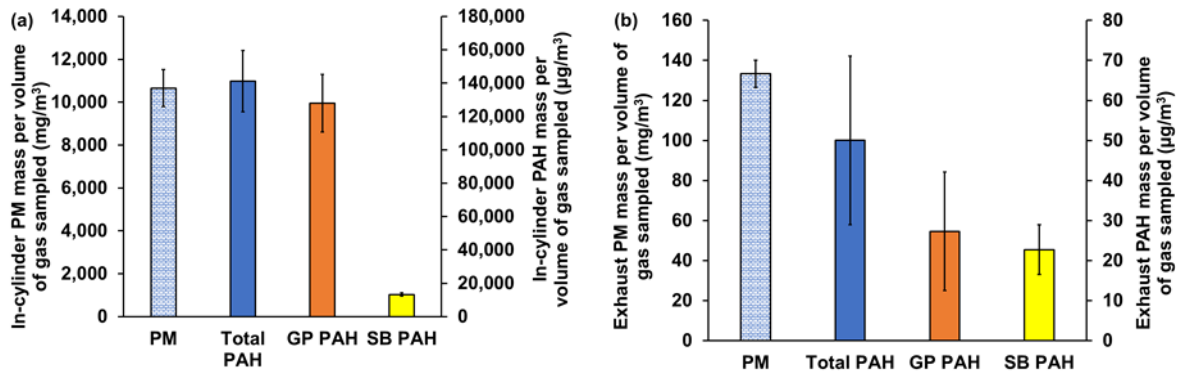
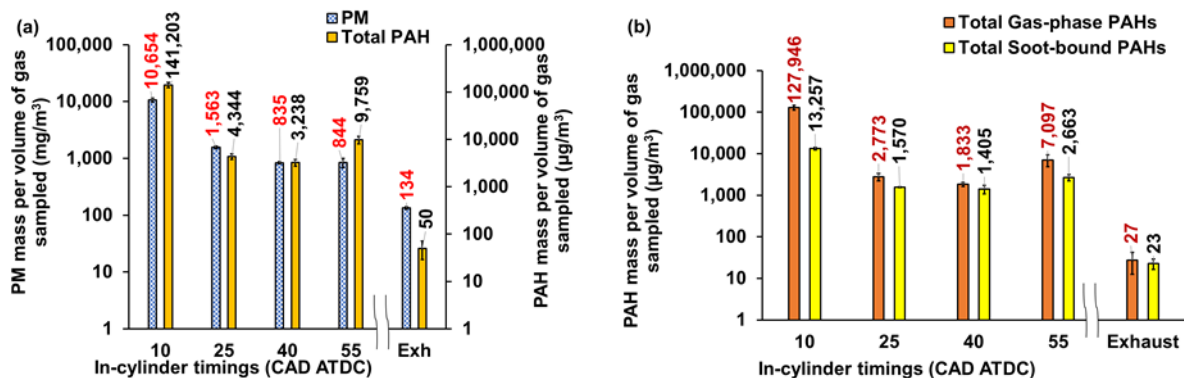


Figure 1: PM, total PAH, gas-phase and soot-bound PAH concentrations present in (a) the engine cylinder immediately following the premixed combustion phase at 10 CAD ATDC (b) the exhaust gas

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Figure 1a shows the PM concentration, total PAH concentration and, gas phase PAH and soot-bound PAH concentrations sampled from the engine cylinder during the premixed combustion phase at 10 CAD ATDC (combustion commenced at TDC) while Figure 1b shows those in the exhaust gas. The concentrations of PM and PAH in the engine cylinder and in the exhaust can be seen in Figure 1a and b respectively. The error bars shown in Figure 1a and b, and where shown in subsequent figures throughout this paper, are plus and minus one standard deviation from the mean of the values measured; three repeat experiments were undertaken for the measurement of PM mass collected with two filters further analysed for determination of PAH concentrations. Sources of variability include cycle to cycle variability in the actual concentration present in the engine due to the turbulent and complex nature of combustion, variability in solvent extraction



process from PM samples, variability in preparing the calibration standard for GCMS runs.

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Figure 2: (a) Masses of particulate and total PAH per volume of gas sampled at 10, 25, 40, 55 (all in CAD ATDC) and exhaust (Exh) (b) Total gas-phase and soot-bound PAH measured at 10, 25, 40, 55 CAD ATDC and the exhaust.

Figure 2a shows the mass of PM and total PAH (gas phase + soot bound phase) per volume of gas measured in the in cylinder gas samples extracted during combustion and from the exhaust gas. Figure 2b shows the total gas phase and total soot bound PAH species on a mass per volume of gas basis at the four in-cylinder sampling timings and from the exhaust gas. The PM and PAH profiles during combustion can be seen in Figure 2a while the gas-phase and soot-bound PAH compositions are shown in Figure 2b.

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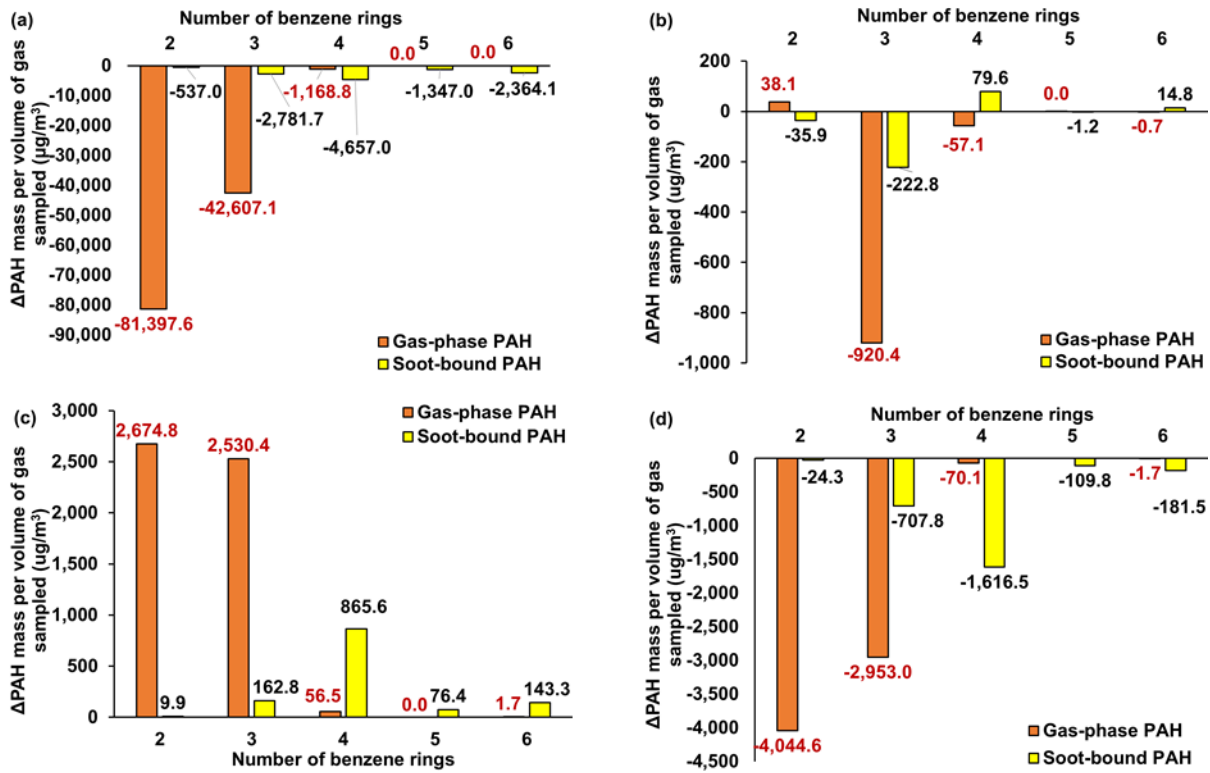


Figure 3: Change in the masses per volume of gas (concentration) of gas-phase and soot-bound 2- to 6-ring PAH during combustion (a) 10 – 25 CAD ATDC (b) 25 - 40 CAD ATDC (c) 40 – 55 CAD ATDC (d) 55 CAD ATDC - Exhaust

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Figure 3a-d show the changes in the total concentrations of gas-phase and soot-bound 2- to 6-ring PAHs during combustion from 10 CAD ATDC through the other in cylinder sampling timings to the exhaust. The changes from one sampling timing to the next through to the exhaust in the concentrations of 2- to 6-ring PAH present in the gas- and soot-bound phases can be seen in Figure 3a-d.

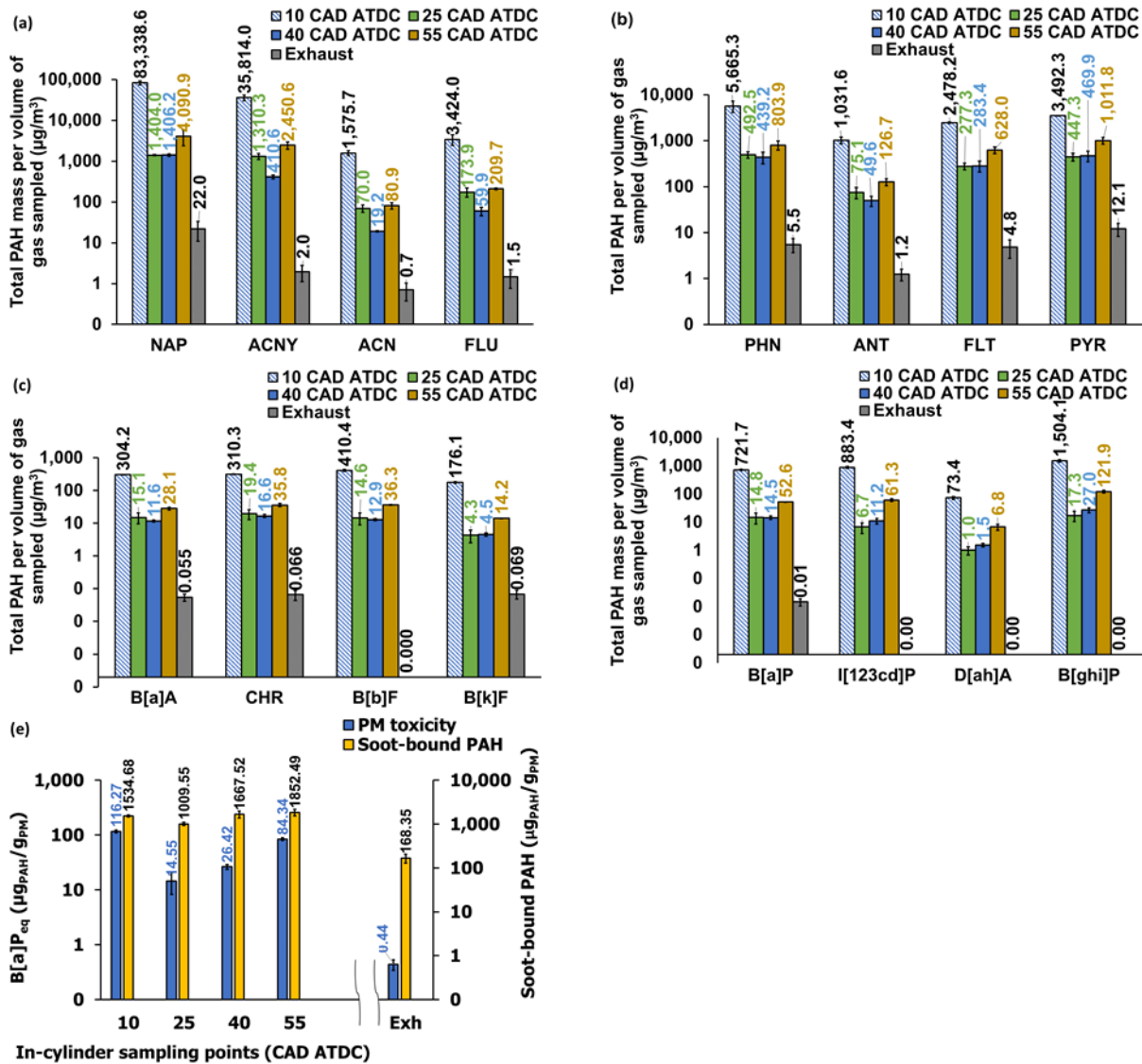


Figure 4: Total measured amounts (gas phase + soot-bound) of each of the 16 EPA priority PAH species at each of the sampling points from 10 CAD to 55 CAD ATDC (a) naphthalene, acenaphthylene, acenaphthene and fluorene (b) phenanthrene, anthracene, fluoranthene and pyrene (c) benzo[a]anthracene, chrysene, benzo[b]fluoranthene, benzo[k]fluoranthene (d) benzo[a]pyrene, indeno[123cd]pyrene, dibenz[ah]anthracene and benzo[ghi]perylene (e) PM toxicity (B[a]P equivalence) and total soot-bound EPA PAH levels of particulate matter collected during combustion at 10 CAD, 25 CAD, 40 CAD, 55 CAD ATDC and exhaust

283 Figure 4a-d shows the abundance of each of the 16 EPA priority PAH species present in
 284 the engine cylinder at the sampling timings. Figure 4e shows the level of PM toxicity and soot-
 285 bound PAH present during combustion at the timings of 10, 25, 40, and 55 CAD ATDC, and
 286 exhaust. The concentration of each of the 16 EPA priority PAH species during combustion and PM
 287 toxicity can be seen in Figure 4a-d and Figure 4e respectively. It can be seen from Figure 4a-d
 288 that all of the 16 EPA priority PAH species were found at varying quantities during diesel
 289 combustion.

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4. Discussion

4.1. In-cylinder PM and PAH

4.1.1. Abundance of PM, Total PAH, gas-phase and soot-bound PAH concentration during premixed combustion

Firstly, it is immediately apparent in Figure 1a that in-cylinder PM and PAH concentrations during premixed combustion are significantly higher relative to those found in the exhaust (Figure 1b). Previous studies by Wang X, Song C, Lv G, Song J, Li H, Li B (22), Barbella R, Bertoli C, Ciajolo A, D'anna A (20), and Malmborg VB, Eriksson AC, Shen M, Nilsson P, Gallo Y, Waldheim B, Martinsson J, Andersson O, Pagels J (21) all found in-cylinder PM concentrations at around 10 CAD ATDC to be significantly greater than those in the exhaust. To understand why the in-cylinder PM concentration is significantly higher at 10 CAD ATDC, Dec's conceptual model of diesel combustion (38) proposed that soot is formed in the fuel-rich core of the leading portion of the fuel/air jet during the combustion of premixed fuel. Furthermore, as 60% of injected fuel was still unburnt during this time and since the sampling zone is from within the fuel spray (Figure S 2a), significant PM can be expected and attributable to the presence of fuel-rich areas and high temperatures. As the combustion chamber volume increases during the expansion stroke, intense turbulent mixing with in-cylinder air dilutes the rich burning zones, oxidising PM and eventually reducing its concentration to the exhaust level.

It is strikingly apparent from Figure 1a that the total in-cylinder PAH concentration significantly exceeds that measured in the exhaust (Figure 1b) (detailed exhaust measurements are available in Section 6.10 of supporting information) and this finding is also consistent with previous studies by Barbella R, Bertoli C, Ciajolo A, D'anna A (20) and Wang X, Song C, Lv G, Song J, Li H, Li B (22). PAHs are known soot precursors and so given the likely presence of conditions suitable for PM formation and the elevated levels of PM present at 10 CAD ATDC relative to the exhaust, an elevated level of PAH might also be expected (Figure 1a). It is immediately clear from Figure 1a that the total in-cylinder PAH concentration is dominated by gaseous PAH (90.6%), which was also observed in the case of the exhaust emission (Figure 1b) but with a much-reduced disparity between levels of gas phase relative to soot-bound PAH. Richter and Howard (1) proposed that the inception of particulate matter begins from heavy PAH molecules, which are typically soot-bound PAHs. It is tentatively suggested that some of the in-cylinder PM concentration measured at 10 CAD ATDC may be as a result of the consumption of soot-bound PAH, and thus much-reduced concentrations relative to the gas phase PAH levels at this time. Interestingly, it can be seen in Figure S 6 that 6-ring PAH were detected in the soot-bound phase, which were not detected in either the exhaust PAH emission (Figure S 4) or in the reference fossil diesel fuel itself. Additionally, as was seen in the case of the exhaust PAH emission, as the number of benzene rings increases, PAH become more abundant in the soot-bound phase, with 5- and 6-ring species detected only in this phase, potentially suggesting that PAH growth continues after adsorption onto soot particles. The computational study of Indarto A, Ghigo G, Maranzana A, Tonachini G on PAH formation following adsorption onto soot particles suggests growth to larger PAH species via cyclisation to be dominant (39). It is also clearly seen from Figure S 6 that the total in-cylinder PAH concentration (Figure 1a) is dominated by 2 and 3-ring PAH species with a significantly greater proportion of these PAH species detected in the gaseous phase. It is suggested that the significant levels of gaseous 2 and 3-ring PAHs could have resulted from the vapourised unburnt fuel likely present at this time. For clarity, 5 and 6-ring PAHs were not detected in the gaseous phase and therefore the total PAH concentration for these species equated to their soot-bound concentration. Each of the 16 EPA PAH species was also quantified at 10 CAD ATDC, shown in Figure S 7a-d (supporting information) and discussed .

346 4.1.2. Abundance of PM, total PAH, gas-phase and soot-bound PAH concentrations from
347 premixed combustion to exhaust

348 The level of PM present at any time during combustion represents a net amount due to
349 competing combustion processes of formation and oxidation. Figure 2a shows that significantly
350 higher levels of both PM and total PAH are present at 10 CAD ATDC than at any other sampling
351 point, and it also indicates PM levels present to have dropped by a factor of 7 between 10 CAD
352 ATDC and the second sampling point of 25 CAD ATDC, in agreement with the findings of Duggal
353 VK, Priede T, Khan IM (40). Between the two sampling timings of 10 and 25 ATDC CAD,
354 combustion was predominantly diffusion-controlled (Figure S 3a)–, during which high rates of PM
355 oxidation can be expected (21). Malmborg VB, Eriksson AC, Shen M, Nilsson P, Gallo Y, Waldheim
356 B, Martinsson J, Andersson O, Pagels J. investigated levels of in-cylinder soot and PAH in a diesel
357 engine equipped with a fast gas sampling valve, at an engine speed of 1200 rpm, and also
358 observed a significant drop in the levels of both soot and PAH between 10 CAD and 20 CAD ATDC.
359 Li Z, Song C, Song J, Lv G, Dong S, Zhao Z. (41) investigated the variation in the size of in-cylinder
360 soot particles during diesel combustion, and observed soot particle size to increase up to a
361 maximum during the early diffusion combustion followed by a decrease. The subsequent decrease
362 in particle size was attributed primarily to high soot oxidation rates and a decrease in the net
363 amount of soot formed as a result of high temperature conditions during the early diffusion
364 combustion phase. The sharp decline in PM between 10 CAD ATDC and 25 CAD ATDC (Figure 2a)
365 is likely to be attributable to a combination of the following four reasons: a reduction in fuel-rich
366 zones within the in-cylinder charge as fuel continued to be burnt; an increase in oxidation rates
367 during this period due to increasing in-cylinder temperature (Figure S 3b); fuel air mixing promoted
368 by in-cylinder air motion and turbulence; and lastly in-cylinder large-scale swirl motion which could
369 have moved fuel-rich areas away from the sampling zone.

370 As combustion proceeded beyond 25 CAD ATDC, Figure 2a shows that in-cylinder PM continued
371 to decrease approximately linearly, reaching a minimum in the exhaust, most likely due to
372 continued soot oxidation and, to a lesser extent, reduced particle formation. The reduction in PM
373 levels during the course of combustion is in agreement with an early optical study utilizing the
374 two-colour method (42). Dec (43) observed that the emergence of PAH species coincided with the
375 disappearance of fuel vapour, indicating fuel pyrolysis and synthesis to have occurred. Hiroyasu
376 H, Arai M, Nakanishi K (44) observed that as soot concentration increased, oxygen concentration
377 decreased (and vice versa when oxygen concentration increased due to air entrainment). At the
378 location of maximum soot, the oxygen concentration was found to be minimum, and in the area
379 where soot concentration increased, liquid fuel evaporated. The similar levels of PM measured at
380 40 CAD and 55 CAD ATDC, given the presence of in-cylinder swirl, could indicate attainment of
381 homogenous in-cylinder conditions. Talibi M, Hellier P, Balachandran R, Ladommatos N (28)
382 observed that the levels of NO_x present in the engine cylinder at 40 CAD ATDC were approximately
383 the same when in-cylinder sampling occurred, first, from within a fuel spray and then between
384 sprays, which suggests near-homogenous in-cylinder conditions during the later stages of diesel
385 combustion (45). Considering PAHs now, Figure 2b shows similar overall trends as those for PM.
386 It can be seen in Figure 2b that a peak in the total mass of PAH occurs at 10 CAD ATDC, which
387 is in agreement with the observed peak level of PM at this time, and likely to have arisen as a
388 result of an expected increase in the rate of fuel pyrolysis following significant heat release during
389 the premixed combustion phase. The significant decline in PAH mass, observed at 25 CAD ATDC
390 relative to 10 CAD ATDC (Figure 2a) could be attributed to the same reasons discussed earlier, in
391 the case of PM. The slight decrease in the total PAH mass observed at 40 CAD relative to 25 CAD
392 ATDC could be attributable to the continued significant oxidation rates present due to high global
393 in-cylinder temperatures (Table 1). At 55 CAD ATDC, Figure 2a shows a significant rise in the total
394 in-cylinder PAH mass to around three-times the value observed at 40 CAD ATDC. This rise in PAH
395 mass could be attributable to an accumulation of new PAH species formed from the pyrolysis of
396 fuel during this interval. Accumulation of newly-formed PAH species can occur due to a
397 combination of decreasing oxidation and soot formation rates as in-cylinder temperature further
398 decreases. Malmborg VB, Eriksson AC, Shen M, Nilsson P, Gallo Y, Waldheim B, Martinsson J,
399 Andersson O, Pagels J found an increase in in-cylinder PAH with the utilization of exhaust gas

400 recirculation (EGR), which they attributed to PAH accumulation due to a decrease in soot formation
401 rates as a result of decreased in-cylinder flame temperatures with EGR. It is also possible that the
402 increased PAH mass at 55 CAD ATDC is due to in-cylinder air motion and swirl changing the relative
403 composition of the fuel spray in the sampling zone. Barbella R, Bertoli C, Ciajolo A, D'anna A (46)
404 observed an increase in heavy hydrocarbon concentration in the later stage of the combustion,
405 and in further studies (20), they also observed an increasing trend in percent weight of total PAH
406 species when in-cylinder gas was sampled from the edge of the fuel spray in the later stages of
407 combustion, despite a trend of decreasing PAH earlier in the course of combustion. Also readily
408 apparent from Figure 2a is the significantly lower exhaust PAH levels relative to that at 55 CAD
409 ATDC which might be expected, and likely to be due to both effect of continued oxidation (albeit
410 at a reduced rate) and the absence of further PAH formation following near-complete fuel
411 consumption. PAH species were extracted from surface of PM and much of the PM burns out during
412 the tail end of combustion, so PM in the exhaust is much smaller relative to that in the engine
413 cylinder (as confirmed in Figure 2a) so exhaust PAH is consequently much smaller. Oxidation is a
414 surface phenomenon and although a reduced oxidation rate is expected during late combustion
415 phase relative to diffusion combustion phase due to lower global in-cylinder temperatures (Table
416 1), the exhaust valve opens much later, around 120 CAD after late combustion during which PAH
417 species have more time to undergo oxidation

418
419 Figure 2b shows that the gas-phase PAH is one order of magnitude higher than soot-bound PAH
420 at 10 CAD ATDC while at the other sampling timings, both levels are of the same order of
421 magnitude, though gas phase PAHs exceed those in the soot-bound phase at all the sampling
422 timings. It can also be observed in Figure 2b that the decrease in total PAH between 10 and 25
423 CAD ATDC seen earlier in Figure 2a was dominated by the reduction in the gas phase PAH. As
424 combustion proceeded beyond 25 CAD ATDC, Figure 2b shows that the total gas-phase and soot-
425 bound PAH mass continued to decrease, although at a much slower rate relative to the significant
426 decrease during early diffusion-controlled combustion, reaching a minimum at 40 CAD ATDC and
427 then increased almost four-fold and two-fold respectively at 55 CAD ATDC. It can be seen in Figure
428 2b that the increase in total PAH level seen at 55 CAD ATDC is dominated by gaseous PAH, likely
429 attributable to pyrolysis as fuel continued to be burned. Also, at 55 CAD ATDC which occurs during
430 the expansion stroke, fuels previously unreacted coming off the cylinder wall crevices into the in-
431 cylinder charge can become pyrolyzed and form new PAH species. Additionally, although, a
432 significant amount of the fuel (95%) has been burnt at 55 CAD ATDC as evidenced in Figure S 3a
433 and in-cylinder temperature at this time is 1277.6 K (Table 1), the remaining 5% of the fuel yet
434 to release useful energy will likely also undergo pyrolysis under these conditions, thus, creating
435 additional PAH species. During the interval between 55 CAD ATDC and engine exhaust, gas-phase
436 and soot-bound PAHs decreased by 99.62% and 99.13% respectively. In general, Figure 2b shows
437 that when PAH levels increase or decrease, the proportion of gas phase PAHs affected significantly
438 exceeds that of soot-bound PAHs.
439

440 4.1.3. Changes in the abundance of total gas-phase and soot-bound PAH species during 441 combustion from premixed to exhaust

442 Figure 3a shows that the substantial decrease in gas phase PAH species seen in Figure 2b
443 between 10 and 25 CAD ATDC can primarily be attributed to a significant decrease in 2- and 3-
444 ring gas-phase PAH species. Of all soot-bound PAH species, 4-ring PAH species decreased the
445 most between 10 to 25 CAD ATDC, likely indicating greater rates of consumption to soot particles
446 or growth to larger PAHs than formation and adsorption onto soot particles. The overall reduction
447 seen in soot-bound PAH species on a per volume basis apparent in Figure 3a is in agreement with
448 the reduction in PM mass between 10 CAD and 25 CAD ATDC seen in Figure 2a, but it is interesting
449 to note that the reduction is not uniform across all PAH sizes.

450 Figure 3b shows a decrease in 3-ring PAH species between 25 CAD and 40 CAD ATDC with a more
451 significant reduction in the gas-phase than in the soot-bound phase and a concurrent increase in
452 4-ring soot-bound PAH species. Figure 3b also shows relatively small increases in the levels of

453 smaller gaseous 2-ring PAH and larger soot-bound 4- and 6-ring PAH (<100 µg/m³). It is suggested
454 therefore that significant levels of growth from 3-ring PAHs to larger species at reaction rates
455 higher than those for the formation of 3-ring PAH may occur during this period.

456 Figure 3c shows a significant increase in 2- & 3-ring gas-phase PAH species and 4-ring
457 soot-bound PAH species between 40 CAD and 55 CAD ATDC. The much more significant increase
458 in the smaller gas-phase PAH suggests presence in the in-cylinder sampling volume of previously
459 unreacted fuel during cylinder-gas expansion, possibly due either to in-cylinder air motion and
460 swirl or the release of fuel impinged on cylinder walls. It is hypothesised that the absence of a
461 concurrent significant increase in PM (Figure 2a) or larger 5- and 6-ring PAH (Figure 3c) suggests
462 the breakdown of the previously unreacted fuel to have not commenced significantly before
463 sampling at 55 CAD ATDC. Figure 3d shows a substantial decrease in all PAH species between 55
464 CAD ATDC and the exhaust, especially 2- and 3-ring PAH species present in the gas phase,
465 suggesting a cessation of PAH formation in addition to continued oxidation.

466 It is tentatively suggested that the much greater reduction in gaseous 2- and 3-ring PAH
467 species (Figure 3a) can be attributed to a combination of the following reasons:

- 468 • PAH oxidation
- 469 • fuel PAH consumption
- 470 • consumption of these species through growth to larger PAH species
- 471 • reduced availability of precursors for the formation of the first ring (benzene) with a greater
472 fuel mass fraction burnt relative to the period between the start of combustion at TDC and
473 10 CAD ATDC (Figure S 3a).

474 4.1.4. Changes in the abundance of each individual PAH species during combustion in 475 the engine cylinder

476 Figure 4a confirms that the majority of the decrease found in the 3-ring PAH species
477 (Figure 3a) between 10 CAD and 25 CAD ATDC is due to a decrease in acenaphthylene. The
478 presence of acenaphthylene is significant as it was shown by Shukla and Koshi (47) that it is
479 possible for acenaphthylene to form higher molecular-weight PAH species via the HACA
480 mechanism. Figure 4b and Figure 4c show that the level of pyrene (PYR) present at 10 CAD ATDC
481 was significantly higher than those of the other 4-ring PAH (fluoranthene, FLT;
482 benzo[a]anthracene, B[a]A; and chrysene, CHR), and the decrease in the level of pyrene from 10
483 to 25 CAD ATDC is also the highest. Appel J, Bockhorn H, Frenklach M (48) noted that previous
484 soot modelling studies represented the soot particle nucleation process as collisions between
485 pyrene and larger PAH species, suggesting pyrene may more readily contribute to PM formation
486 than smaller PAH. Figure 4c and d show that the level of D[ah]A (5-ring PAH) at 10 CAD ATDC
487 was the smallest relative to other PAH of the same size (B[b]F, B[k]F and B[a]P). The benzene
488 rings in D[ah]A join together to form a chain which is different from the clustered structure of the
489 other 5-ring PAH species. The significantly lower level of D[ah]A could therefore suggest
490 preferential PAH growth to PAH species with clustered benzene rings. Figure 4c and d show a
491 higher proportional increase in 5- and 6-ring PAH levels at 55 CAD relative to 40 CAD ATDC except
492 for dibenz[ah]anthracene; thus, suggesting that in-cylinder conditions seem to favour the
493 persistence of closely-clustered high molecular-weight PAHs later on during combustion.

495 4.1.4.1. B[a]P equivalence of particulate matter during the course of combustion

496 The B[a]P equivalence of particulate matter indicates the level of its toxicity and was
497 calculated using Equation 2.2. Readily apparent from Figure 4e is the increase in both PM B[a]P
498 equivalence and soot-bound PAH in the engine cylinder, following a decrease earlier at 25 CAD
499 ATDC. The increase in PM B[a]P equivalence is primarily attributable to an increase in the
500 abundance of soot-bound PAH species, especially larger ones belonging to the B2 group whose
501 TEF is relatively high. Thereafter, in the period between 40 CAD and 55 CAD ATDC, continuing
502 combustion resulted in increased levels of toxic PAH species bound onto PM, despite an only slight
503 increase in total soot-bound PAH, and can be attributed to a significant increase in B[a]P and
504 D[ah]A during this period (Figure 4d).

505 4.1.5. Discussion of potential PAH growth mechanisms identified in diesel combustion

506 Between 25 CAD and 40 CAD a decrease in the level of phenanthrene concurred with an
507 increase in the levels of both pyrene and benzo[ghi]perylene (see Figure S 11a for graphical
508 representation). Shukla and Koshi (49) studied possible precursors to major pyrolysis products of
509 benzene, acetylene (C₂H₂) and a mixture of benzene and acetylene in a flow tube reactor and they
510 suggested that benzo[e]pyrene (B[e]P) yields benzo[ghi]perylene (B[ghi]P) via the HACA
511 mechanism. Although, B[e]P was not measured in this study, it is therefore cautiously suggested
512 that the growth of phenanthrene to B[ghi]P (Figure S 9) may have occurred into two stages:

- 513 1. Phenanthrene to B[e]P via Phenyl Addition Cyclisation (PAC mechanism)
- 514 2. B[e]P to B[ghi]P via the HACA mechanism

515 Acetylene is a well-known key PAH precursor and its likely involvement in the formation of both
516 pyrene and benzo[ghi]perylene suggests that it is a possible pyrolysis product of fossil diesel. For
517 instance, Ciajolo A, D'Anna A, Barbella R, Bertoli C (24) found acetylene to be one of the major
518 pyrolysis products in the combustion of tetradecane (a diesel-fuel surrogate) and several soot
519 modelling studies include the pyrolytic production of acetylene as one of the steps in the soot
520 formation process (50), (51), (52). Also apparent during this period, is a decrease in the level of
521 benzo[b]fluoranthene concurred with a greater increase in the level of indeno[123-cd]pyrene
522 (I[123-cd]P) (see Figure S 11b). Shukla and Koshi (47) also suggested that phenylpyrene (phenyl
523 addition to pyrene) is a precursor to the formation of indeno[123-cd]pyrene; this is a possibility
524 given the apparent availability of pyrene and the possibility of the presence of benzene among the
525 pyrolytic products of fossil diesel as suggested by Xanthopoulou (53). It is cautiously suggested
526 that abstraction of two hydrogen atoms from benzo[b]fluoranthene (C₂₀H₁₂) and subsequent
527 addition of acetylene via HACA () could have also contributed to the formation of a greater amount
528 of indeno[123-cd]pyrene (C₂₂H₁₂), in addition to formation via phenylpyrene.

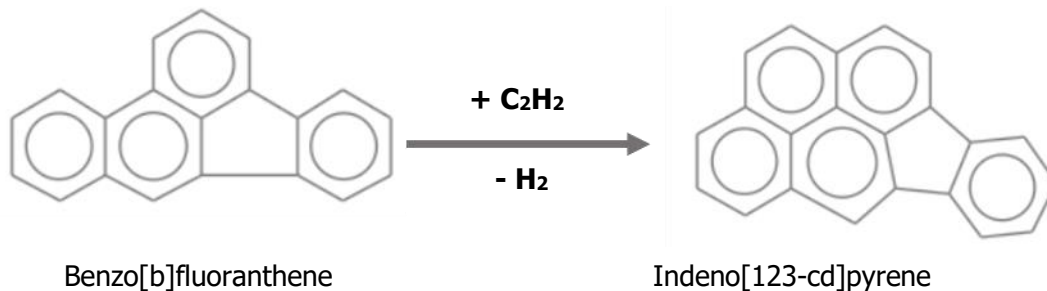


Figure 5: Growth of benzo[b]fluoranthene to Indeno[123-cd]pyrene via HACA mechanism

529 **5. Implications**

530 Exhaust PAH was composed mainly of 2- and 3-ring PAH species predominantly present in the
531 gaseous phase which can increase the toxicity of ambient air to humans. There is evidence in this
532 study that PAH levels reduced, and PM burned out as in-cylinder temperature increased during
533 diffusion combustion while during late combustion when temperature fell, PAH levels rise, including
534 the B2 group PAH. Therefore, a practical implication is to consider design that improves and
535 enhances oxidation and high temperature towards end of combustion to limit PAH rise and
536 consequently further reduce exhaust PAH levels. An environmental implication is that a further
537 reduction in PAH levels during late combustion will further reduce exhaust levels of B2-group PAH
538 and consequently the toxicity of exhaust gas. However, any such combustion strategy to reduce
539 late combustion formation and accumulation of PAH should also consider the potential effects on
540 thermal efficiency and NO_x emissions, and thus further impacts on GHG and air quality emissions,
541 respectively.
542
543

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547

548 Supporting information

549 The 16 EPA priority PAH species measured in this study, Schematic diagram and diesel engine
550 specifications, In-cylinder gas sampling valve description, Fossil diesel fuel properties, Description
551 of the Accelerated solvent extraction and optimised extraction conditions, GCMS configuration,
552 PAH assignment to deuterated PAH species, In-cylinder conditions and sampling timings, Measured
553 concentrations of exhaust PM and PAH, PAH speciation according to the number of benzene rings
554 during combustion, Change in the abundance of total gas-phase and soot-bound PAH species
555 during combustion from premixed to exhaust, Potential PAH growth mechanisms in diesel
556 combustion.

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References

- 601 1. Richter H, Howard JB. Formation of polycyclic aromatic hydrocarbons and their growth
602 to soot—a review of chemical reaction pathways. *Progress in Energy and Combustion*
603 *Science*. 2000 Aug 1;26(4):565–608.
- 604 2. Sprouse C, Depcik C. Review of organic Rankine cycles for internal combustion engine
605 exhaust waste heat recovery. *Applied Thermal Engineering*. 2013 Mar 1;51(1):711–22.
- 606 3. Yue T, Gao X, Gao J, Tong Y, Wang K, Zuo P, Zhang X, Tong L, Wang C, Xue Y. Emission
607 characteristics of NO_x, CO, NH₃ and VOCs from gas-fired industrial boilers based on field
608 measurements in Beijing city, China. *Atmospheric Environment*. 2018 Jul 1;184:1–8.
- 609 4. Mau V, Gross A. Energy conversion and gas emissions from production and combustion
610 of poultry-litter-derived hydrochar and biochar. *Applied Energy*. 2018 Mar 1;213:510–9.
- 611 5. Han F, Guo H, Hu J, Zhang J, Ying Q, Zhang H. Sources and health risks of ambient
612 polycyclic aromatic hydrocarbons in China. *Science of The Total Environment*. 2020 Jan
613 1;698:134229.
- 614 6. Strandberg B, Julander A, Sjöström M, Lewné M, Hatice KA, Bigert C. An improved
615 method for determining dermal exposure to polycyclic aromatic hydrocarbons.
616 *Chemosphere*. 2018 May 1;198:274–80.
- 617 7. Hoppe-Jones C, Beitel S, Burgess JL, Snyder S, Flahr L, Griffin S, Littau S, Jeong KS,
618 Zhou J, Gulotta J, Moore P. Use of urinary biomarkers and bioassays to evaluate chemical
619 exposure and activation of cancer pathways in firefighters. *Occup Environ Med*. 2018 Apr
620 1;75(Suppl 2):A412–3.
- 621 8. Mayer AC, Fent KW, Bertke S, Horn GP, Smith DL, Kerber S, La Guardia MJ. Firefighter
622 hood contamination: Efficiency of laundering to remove PAHs and FRs. *Journal of*
623 *Occupational and Environmental Hygiene*. 2018 Nov 14;1–32.
- 624 9. Chen A, Northcross A, Folger S. Environmental Degradation in Baía de Todos os Santos,
625 Brazil: A Review of the Evidence. 2018;
- 626 10. Geier MC, Chlebowski AC, Truong L, Massey Simonich SL, Anderson KA, Tanguay RL.
627 Comparative developmental toxicity of a comprehensive suite of polycyclic aromatic
628 hydrocarbons. *Arch Toxicol*. 2018 Feb 1;92(2):571–86.
- 629 11. Ballesteros R, Hernández JJ, Lyons LL. An experimental study of the influence of biofuel
630 origin on particle-associated PAH emissions. *Atmospheric Environment*. 2010 Mar
631 1;44(7):930–8.
- 632 12. Nadal M, Schuhmacher M, Domingo JL. Levels of PAHs in soil and vegetation samples
633 from Tarragona County, Spain. *Environmental Pollution*. 2004 Nov 1;132(1):1–11.
- 634 13. Keith LH. The source of US EPA's sixteen PAH priority pollutants. *Polycyclic Aromatic*
635 *Compounds*. 2015;35(2–4):147–160.

- 636 14. He C, Ge Y, Tan J, You K, Han X, Wang J. Characteristics of polycyclic aromatic
637 hydrocarbons emissions of diesel engine fueled with biodiesel and diesel. *Fuel*. 2010
638 Aug;89(8):2040–6.
- 639 15. Tsai J-H, Chen S-J, Huang K-L, Lin Y-C, Lee W-J, Lin C-C, Lin W-Y. PM, carbon, and PAH
640 emissions from a diesel generator fuelled with soy-biodiesel blends. *Journal of Hazardous*
641 *Materials*. 2010 Jul 15;179(1):237–43.
- 642 16. Tsai J-H, Chen S-J, Huang K-L, Lee W-J, Kuo W-C, Lin W-Y. Characteristics of particulate
643 emissions from a diesel generator fueled with varying blends of biodiesel and fossil diesel.
644 *Journal of Environmental Science and Health, Part A*. 2011 Jan 1;46(2):204–13.
- 645 17. Borrás E, Tortajada-Genaro LA, Vázquez M, Zielinska B. Polycyclic aromatic hydrocarbon
646 exhaust emissions from different reformulated diesel fuels and engine operating
647 conditions. *Atmospheric Environment*. 2009 Dec 1;43(37):5944–52.
- 648 18. How HG, Teoh YH, Masjuki HH, Kalam MA. Impact of coconut oil blends on particulate-
649 phase PAHs and regulated emissions from a light duty diesel engine. *Energy*. 2012 Dec
650 1;48(1):500–9.
- 651 19. Lin Y-C, Lee W-J, Hou H-C. PAH emissions and energy efficiency of palm-biodiesel blends
652 fueled on diesel generator. *Atmospheric Environment*. 2006 Jul 1;40(21):3930–40.
- 653 20. Barbella R, Bertoli C, Ciajolo A, D’anna A. Behavior of a fuel oil during the combustion
654 cycle of a direct injection diesel engine. *Combustion and Flame*. 1990 Nov 1;82(2):191–
655 8.
- 656 21. Malmborg VB, Eriksson AC, Shen M, Nilsson P, Gallo Y, Waldheim B, Martinsson J.
657 Evolution of In-Cylinder Diesel Engine Soot and Emission Characteristics Investigated
658 with Online Aerosol Mass Spectrometry. *Environmental Science & Technology*. 2017 Feb
659 7;51(3):1876–85.
- 660 22. Wang X, Song C, Lv G, Song J, Li H, Li B. Evolution of in-cylinder polycyclic aromatic
661 hydrocarbons in a diesel engine fueled with n-heptane and n-heptane/toluene. *Fuel*.
662 2015 Oct 15;158:322–9.
- 663 23. Narushima T, Morishima A, Moriwaki H, Kusaka J, Daisho Y. Experimental and Numerical
664 Studies on Soot Formation in Fuel Rich Mixture [Internet]. Warrendale, PA: SAE
665 International; 2003 May [cited 2018 Dec 5]. Report No.: 2003-01–1850. Available from:
666 <https://www.sae.org/publications/technical-papers/content/2003-01-1850/>
- 667 24. Ciajolo A, D’Anna A, Barbella R, Bertoli C. Combustion of Tetradecane and Tetradecane/
668 α -Methylnaphthalene in a Diesel Engine with Regard to Soot and PAH Formation.
669 *Combustion Science and Technology*. 1993 Jan;87(1–6):127–37.
- 670 25. Kang I, Bae J, Bae G. Performance comparison of autothermal reforming for liquid
671 hydrocarbons, gasoline and diesel for fuel cell applications. *Journal of Power Sources*.
672 2006 Dec 7;163(1):538–46.
- 673 26. Aakko P, Harju T, Niemi M, Rantanen-Kolehmainen L. PAH content of diesel fuel and
674 automotive emissions. VTT Technical Research Centre of Finland Research Report VTT.
675 2006;

- 676 27. Tree DR, Svensson KI. Soot processes in compression ignition engines. *Progress in*
677 *Energy and Combustion Science*. 2007 Jun 1;33(3):272–309.
- 678 28. Talibi M, Hellier P, Balachandran R, Ladommatos N. Development of a fast-acting, time-
679 resolved gas sampling system for combustion and fuels analysis. *SAE International*
680 *Journal of Engines*. 2016;9(2):1102–1116.
- 681 29. US EPA. Compendium of Methods for the Determination of Toxic Organic Compounds in
682 Ambient Air. Environmental Protection Agency, US Federal Register, Compendium
683 Method TO-13A 1-42 [Internet]. 1999 [cited 2018 Apr 6]. Available from:
684 <https://www3.epa.gov/ttn/amtic/files/ambient/airtox/to-13arr.pdf>
- 685 30. Dandajeh HA, Ladommatos N, Hellier P, Eveleigh A. Effects of unsaturation of C2 and C3
686 hydrocarbons on the formation of PAHs and on the toxicity of soot particles. *Fuel*. 2017
687 Apr 15;194:306–20.
- 688 31. Sánchez NE, Callejas A, Millera A, Bilbao R, Alzueta MU. Formation of PAH and soot
689 during acetylene pyrolysis at different gas residence times and reaction temperatures.
690 *Energy*. 2012 Jul 1;43(1):30–6.
- 691 32. Karavalakis G, Boutsika V, Stournas S, Bakeas E. Biodiesel emissions profile in modern
692 diesel vehicles. Part 2: Effect of biodiesel origin on carbonyl, PAH, nitro-PAH and oxy-
693 PAH emissions. *Science of The Total Environment*. 2011 Jan 15;409(4):738–47.
- 694 33. Karavalakis G, Bakeas E, Fontaras G, Stournas S. Effect of biodiesel origin on regulated
695 and particle-bound PAH (polycyclic aromatic hydrocarbon) emissions from a Euro 4
696 passenger car. *Energy*. 2011 Aug 1;36(8):5328–37.
- 697 34. Dandajeh HA, Ladommatos N, Hellier P, Eveleigh A. Influence of carbon number of C1–
698 C7 hydrocarbons on PAH formation. *Fuel*. 2018 Sep 15;228:140–51.
- 699 35. Park SS, Kim YJ, Kang CH. Atmospheric polycyclic aromatic hydrocarbons in Seoul, Korea.
700 *Atmospheric Environment*. 2002 Jun 1;36(17):2917–24.
- 701 36. Kanaujia PK, Singh D, Tripathi D, Konathala LNSK, Saran S, Chauhan RK, Sharma YK,
702 Garg MO. Characterization and Identification of Polycyclic Aromatic Hydrocarbons in
703 Diesel Particulate Matter. *Analytical Letters*. 2015 Sep 22;48(14):2303–18.
- 704 37. Nisbet ICT, LaGoy PK. Toxic equivalency factors (TEFs) for polycyclic aromatic
705 hydrocarbons (PAHs). *Regulatory Toxicology and Pharmacology*. 1992 Dec 1;16(3):290–
706 300.
- 707 38. Dec JE. Advanced compression-ignition engines—understanding the in-cylinder
708 processes. *Proceedings of the Combustion Institute*. 2009;32(2):2727–42.
- 709 39. Indarto A, Ghigo G, Maranzana A, Tonachini G. Polycyclic aromatic hydrocarbon
710 formation mechanism in the ‘particle phase’. A theoretical study. *Physical chemistry*
711 *chemical physics: PCCP*. 2010 Aug 28;12:9429–40.
- 712 40. Duggal VK, Priede T, Khan IM. A Study of Pollutant Formation within the Combustion
713 Space of a Diesel Engine. *SAE Transactions*. 1978;87:987–1000.

- 714 41. Li Z, Song C, Song J, Lv G, Dong S, Zhao Z. Evolution of the nanostructure, fractal
715 dimension and size of in-cylinder soot during diesel combustion process. *Combustion and*
716 *Flame*. 2011 Aug 1;158(8):1624–30.
- 717 42. Aoyagi Y, Kamimoto T, Matsui Y, Matsuoka S. A Gas Sampling Study on the Formation
718 Processes of Soot and NO in a DI Diesel Engine. *SAE Transactions*. 1980;89:1175–89.
- 719 43. Dec JE. A Conceptual Model of DI Diesel Combustion Based on Laser-Sheet Imaging*
720 [Internet]. Warrendale, PA: SAE International; 1997 Feb [cited 2018 Feb 19]. Report
721 No.: 970873. Available from: <http://papers.sae.org/970873/>
- 722 44. Hiroyasu H, Arai M, Nakanishi K. Soot Formation and Oxidation in Diesel Engines. *SAE*
723 *Transactions*. 1980;89:1148–62.
- 724 45. Talibi M. Co-combustion of diesel and gaseous fuels with exhaust emissions analysis and
725 in-cylinder gas sampling [Internet] [Doctoral]. UCL (University College London); 2015
726 [cited 2019 May 1]. Available from: <http://discovery.ucl.ac.uk/1471760/>
- 727 46. Barbella R, Bertoli C, Ciajolo A, D'Anna A, Masi S. In-Cylinder Sampling of High Molecular
728 weight Hydrocarbons From a D.I. Light Duty Diesel Engine. *SAE Transactions*.
729 1989;98:712–21.
- 730 47. Shukla B, Koshi M. A novel route for PAH growth in HACA based mechanisms.
731 *Combustion and Flame*. 2012 Dec 1;159(12):3589–96.
- 732 48. Appel J, Bockhorn H, Frenklach M. Kinetic modeling of soot formation with detailed
733 chemistry and physics: laminar premixed flames of C2 hydrocarbons. *Combustion and*
734 *Flame*. 2000 Apr 1;121(1):122–36.
- 735 49. Shukla B, Koshi M. A highly efficient growth mechanism of polycyclic aromatic
736 hydrocarbons. *Physical Chemistry Chemical Physics*. 2010;12(10):2427.
- 737 50. Belardini P, Bertoli C, Del Giacomo N, Iorio B. Soot formation and oxidation in a di diesel
738 engine: A comparison between measurements and three dimensional computations. In:
739 Fall Fuels and Lubricants Meeting and Exposition, October 18, 1993 - October 21, 1993.
740 SAE International; 1993. (SAE Technical Papers).
- 741 51. Sung N, Lee S, Kim H, Kim B. A numerical study on soot formation and oxidation for a
742 direct injection diesel engine. *Proceedings of the Institution of Mechanical Engineers,*
743 *Part D: Journal of Automobile Engineering*. 2003 May 1;217(5):403–13.
- 744 52. Cheng X, Chen L, Yan F, Dong S. Study on soot formation characteristics in the diesel
745 combustion process based on an improved detailed soot model. *Energy Conversion and*
746 *Management*. 2013;75:1–10.
- 747 53. Xanthopoulou G. Oxide catalysts for pyrolysis of diesel fuel made by self-propagating
748 high-temperature synthesis. Part I: cobalt-modified Mg–Al spinel catalysts. *Applied*
749 *Catalysis A: General*. 1999 Jun 21;182(2):285–95.
- 750 54. Kim K-H, Jahan SA, Kabir E, Brown RJC. A review of airborne polycyclic aromatic
751 hydrocarbons (PAHs) and their human health effects. *Environment International*. 2013
752 Oct;60:71–80.

- 753 55. Karabektas M, Hosoz M. Performance and emission characteristics of a diesel engine
754 using isobutanol–diesel fuel blends. *Renewable Energy*. 2009 Jun 1;34(6):1554–9.
- 755 56. Lu T, Cheung CS, Huang Z. Effects of engine operating conditions on the size and
756 nanostructure of diesel particles. *Journal of Aerosol Science*. 2012 May 1;47:27–38.
- 757 57. Corrêa SM, Arbilla G. Aromatic hydrocarbons emissions in diesel and biodiesel exhaust.
758 *Atmospheric Environment*. 2006 Nov 1;40(35):6821–6.
- 759 58. Barbella R, Bertoli C, Ciajolo A, D’anna A. Soot and Unburnt Liquid Hydrocarbon
760 Emissions from Diesel Engines. *Combustion Science and Technology*. 1988 May 1;59(1–
761 3):183–98.
- 762 59. Tavares M, Pinto JP, Souza AL, Scarmínio IS, Cristina Solci M. Emission of polycyclic
763 aromatic hydrocarbons from diesel engine in a bus station, Londrina, Brazil. *Atmospheric*
764 *Environment*. 2004 Sep 1;38(30):5039–44.
- 765 60. Marr LC, Kirchstetter TW, Harley RA, Miguel AH, Hering SV, Hammond SK.
766 Characterization of Polycyclic Aromatic Hydrocarbons in Motor Vehicle Fuels and Exhaust
767 Emissions. *Environmental Science & Technology*. 1999 Sep;33(18):3091–9.
- 768 61. Zielinska B, Sagebiel J, Arnott WP, Rogers CF, Kelly KE, Wagner DA, Lighty JS, Sarofim
769 AF, Palmer G. Phase and Size Distribution of Polycyclic Aromatic Hydrocarbons in Diesel
770 and Gasoline Vehicle Emissions. *Environmental Science & Technology*. 2004
771 May;38(9):2557–67.
- 772 62. Sánchez NE, Callejas A, Millera Á, Bilbao R, Alzueta MU. Polycyclic Aromatic Hydrocarbon
773 (PAH) and Soot Formation in the Pyrolysis of Acetylene and Ethylene: Effect of the
774 Reaction Temperature. *Energy & Fuels*. 2012 Aug 16;26(8):4823–9.
- 775 63. Tancell PJ, Rhead MM, Pemberton RD, Braven Jim. Survival of Polycyclic Aromatic
776 Hydrocarbons during Diesel Combustion. *Environmental Science & Technology*. 1995
777 Nov;29(11):2871–6.
- 778 64. Williams PT, Bartle KD, Andrews GE. The relation between polycyclic aromatic
779 compounds in diesel fuels and exhaust particulates. *Fuel*. 1986 Aug 1;65(8):1150–8.
- 780 65. Williams PT, Abbass MK, Andrews GE, Bartle KD. Diesel particulate emissions: The role
781 of unburned fuel. *Combustion and Flame*. 1989 Jan 1;75(1):1–24.
- 782 66. Rhead MM, Pemberton RD. Sources of Naphthalene in Diesel Exhaust Emissions. *Energy*
783 *& Fuels*. 1996 Jan;10(3):837–43.
- 784 67. Kislov VV, Sadovnikov AI, Mebel AM. Formation Mechanism of Polycyclic Aromatic
785 Hydrocarbons beyond the Second Aromatic Ring. *J Phys Chem A*. 2013 Jun
786 13;117(23):4794–816.
- 787 68. Dobbins RA, Fletcher RA, Benner BA, Hoefft S. Polycyclic aromatic hydrocarbons in flames,
788 in diesel fuels, and in diesel emissions. *Combustion and Flame*. 2006 Mar 1;144(4):773–
789 81.
- 790
791
792