

1 **Linkage analysis of economic consumption, pollutant emissions and**  
2 **concentrations based on a city-level multi-regional Input–output (MRIO) model**  
3 **and atmospheric transport**

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17 **Abstract:** China is experiencing serious atmospheric pollution, which also exhibits  
18 significant spatial heterogeneity. The Chinese government has implemented targeted  
19 pollution control measures at the city level, emphasizing coordination among cities to  
20 prevent and control air pollution in key regions such as Beijing–Tianjin–Hebei (BTH)  
21 urban agglomeration. This study combined an inter-city multi-regional input–output  
22 (MRIO) model with an air quality dispersion model consisting of a weather research  
23 and forecasting (WRF) model and the CALPUFF model (WRF/CALPUFF) to study  
24 the inter-city economic consumption, pollutant emission and concentration among 13  
25 cities in BTH urban agglomeration. NO<sub>x</sub> is chosen as an example. The combined  
26 effects of economic linkage and atmospheric transport show that NO<sub>x</sub> concentrations  
27 in cities in the BTH urban agglomeration are attributable to three consumption  
28 sources: a local contribution from the target city’s own local economic consumption  
29 (average, 25%), and non-local consumption contributions, including other cities in  
30 the BTH urban agglomeration (average, 36%) and regions outside of BTH (average,  
31 39%). Compared with the contributions to NO<sub>x</sub> concentrations calculated using only  
32 the MRIO model or atmospheric transport stimulation model, the results of this paper  
33 quantify that the consumption outside of a city could provide a greater impact on the  
34 city’s air quality due to the combined effects of economic linkage and atmospheric  
35 transport. To avoid negative impacts of emission reduction targets on economic  
36 consumption, governmental regional pollution control policies should consider the  
37 combined effects of economic linkage and atmospheric transport.

38 **Keywords:** Atmospheric transport; Beijing–Tianjin–Hebei; Concentration; Emissions;  
39 MRIO; WRF/CALPUFF

## 40 1. Introduction

41 The rapid development of China's heavy industry and the intensive use of energy  
42 have caused severe air pollution and negative public health impacts in China over the  
43 past few decades, which has become a significant environmental problem in China  
44 (Zhang et al., 2019). The latest Law of the People's Republic of China on the Prevention  
45 and Control of Atmospheric Pollution emphasizes collaborative efforts across  
46 administrative boundaries for emissions control and pollution prevention [Ministry of  
47 Ecology and Environmental Protection of China (MEPC), 2018]. Such collaborative  
48 pollution prevention efforts require an understanding of pollution at the regional scale,  
49 including the interaction among regions or even cities.

50 Currently, two types of research are being conducted to explore interactions of air  
51 pollution from different sources at the regional level. The first is transboundary  
52 atmospheric transport research, which focuses on how local air quality is affected by  
53 atmospheric transport of pollution from non-local sources (Chang et al., 2018; Hua et al.,  
54 2016; Huang et al., 2015; Li et al., 2015; Kwok et al., 2013; Zheng et al., 2018). This  
55 type of research is based on both **air quality simulation** modeling and measurements of  
56 air pollutants to determine their interactions among regions. The second type of research  
57 emphasizes the transfer of virtual emissions through trade (Weber et al., 2007; Yong et  
58 al., 2017; Zhao et al., 2015; Jiang et al., 2016, 2017; Mi et al., 2017; Xu et al., 2009).  
59 The contribution from non-local final demand to local emissions is a key problem  
60 resulting from the production of goods (and their associated emissions) in one region for  
61 consumption in another region. **Many previous studies have focused on examining the  
62 embodied environmental impacts of trade by input-output (IO) model, such as  
63 wastewater (Zheng et al., 2020), municipal solid waste (Li et al., 2019). But unlike other  
64 wastes, air pollutants emissions will affect other areas along with the wind. Therefore,**

65 the integration of the economic model with the atmospheric transport of pollution model  
66 is necessary.

67 Recently, a third type of research has attempted to combine the multi-regional  
68 input-output (MRIO) model with the atmospheric transport of pollution model, to  
69 determine the concentrations of pollutants produced by emissions driven by  
70 consumption. These studies have focused on the coupling of atmospheric transport and  
71 trade at the country or provincial scales, such as China and the United States (Lin et al.,  
72 2014), among countries worldwide (Zhang et al., 2017) and among multiple Chinese  
73 provinces (Li et al., 2016; Lu et al., 2019; Zhao et al., 2017). Many studies have  
74 considered the impact of atmospheric pollutant transmission and economic consumption,  
75 but most of them are between various countries and provinces. From intercity  
76 perspective, the combined effects of atmospheric pollutant transport and economic  
77 linkages are seldom considered, mainly due to the lack of appropriate city-level data.  
78 Zheng et al. (2019) compiled a city-level MRIO table for Hebei Province of China to  
79 determine a city-level energy footprint; this table is also useful for studying  
80 consumption-driven air quality at the city scale, which could in turn yield accurate  
81 scientific data for studies of the interactions of air pollutants at the city level from the  
82 perspective of economic consumption.

83 The purpose of this study is to examine air pollution interactions among cities  
84 using the city-level MRIO model and an air quality simulation model [a weather  
85 research and forecasting (WRF) model coupled with the CALPUFF model  
86 (WRF/CALPUFF)]. Although there are a lot of research about CALPUFF modeling  
87 (Abdul-Wahab et al., 2011; Dresser et al., 2011; Wu et al., 2018; Shubbar et al., 2019),  
88 few studies have considered the combination of this model and MRIO model to reveal  
89 the contribution on air quality driven by economic consumption. This paper attempts to

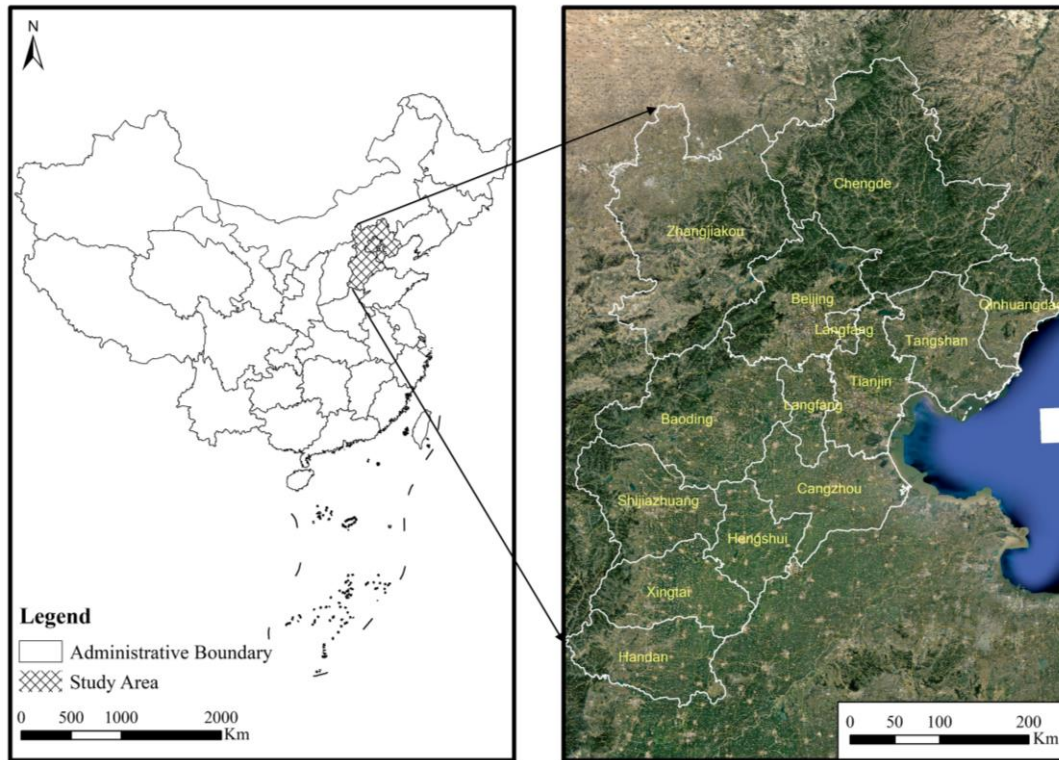
90 combine the above two models to reveal the natural and economic linkages of air  
91 pollution among cities. The air quality simulation model is used to simulate the  
92 diffusion process of pollutants and further calculate the concentration of pollutants in  
93 each region. In addition, this study also focuses on the relationship between the  
94 pollutants and economic flows among these regions. Therefore, the MRIO model is  
95 introduced to estimate the hidden environmental impact of trade. By combining and  
96 comparing the results of these models, we can find the difference of pollutant emission  
97 and economic consumption on pollutant concentration.

98 In this paper, the Beijing–Tianjin–Hebei urban agglomeration (BTH) is chosen as  
99 the study area. Because this region is the most polluted areas in China, containing 5 of  
100 the 10 cities with the worst air quality in the country; 49.5% of days did not satisfy the  
101 national air quality standard in 2018 (MEPC, 2019). The BTH region has become a  
102 hotspot for air pollution research in China, including studies of emissions attributable to  
103 the supply chain by MRIO model (Zhao et al., 2016; Wang et al., 2017a; Chen et al.,  
104 2016; Zhang et al., 2016; Zheng et al., 2016). However, these studies regarded Hebei as  
105 a whole province, and did not analyze 11 cities in Hebei Province. Obviously, the  
106 research between cities can further reveal the emissions of economic linkage attributed  
107 to the supply chains in BTH region. There are also some researches on physical and  
108 chemical atmospheric transport of pollutants among cities in this area (Wang et al.,  
109 2017b; Chang et al., 2019). However, few studies have reported the combined effects of  
110 atmospheric transport and economic linkage on air pollution in the BTH region.

111 In sum, this study considers pollutant concentrations, emissions, and economic  
112 consumption to understand atmospheric pollution at the city level. The results of this  
113 study could facilitate optimization of emission reduction targets for different cities, to  
114 minimize the impact of emission reduction on the economy and living standards.

115 **2. Study area and data sources**

116 The study area was the Beijing–Tianjin–Hebei urban agglomeration (Figure 1), one  
117 of three major urban agglomerations in China. BTH includes Beijing, Tianjin, and 11  
118 cities in Hebei Province.



119

120 **Figure 1.** Location of the Beijing–Tianjin–Hebei (BTH) urban agglomeration. **City abbreviations are**  
121 **BJ, Beijing; TJ, Tianjin; SJZ, Shijiazhuang; TS, Tangshan; QHD, Qinhuangdao; HD, Handan; XT,**  
122 **Xingtai; BD, Baoding; ZJK, Zhangjiakou; CD, Chengde; LF, Langfang; HS, Hengshui.**

123

124 **NO<sub>x</sub>** is chosen as a typical air pollutant to explore economic and atmospheric  
125 **linkages of air pollution among cities.** The control of NO<sub>x</sub> is **one of key steps** in  
126 reducing PM<sub>2.5</sub> and O<sub>3</sub> levels (Sillman et al., 1990). Although there are many sources of  
127 emissions in the BTH urban agglomeration, including industry, residential areas and  
128 transportation, among these, industrial pollution sources account for 70–90% of the total  
129 emissions (MEPC, 2013). Limited by statistical data, **it** only has relatively accurate data  
130 of industrial emissions (including the Electricity & heat sector and Transportation &

131 storage sector) at present. The provincial NO<sub>x</sub> emissions data are obtained from the  
132 2012 National Environmental Statistical Yearbook. The city-level NO<sub>x</sub> emissions data  
133 are derived from the official environmental statistics emission database. NO<sub>x</sub> emissions  
134 are aggregated from 12,929 industrial enterprises in BTH into 22 industry sectors. More  
135 details on 13 cities of BTH region with the 22 industrial sectors and those industrial  
136 enterprises distributions are provided in the Supplemental Information (Table S1, S2  
137 and Figure S1). Because the national MRIO table lacks detailed city-level data, the 2012  
138 city-level MRIO table in Zheng et al. (2019) is used, which describes intermediate trade  
139 flow among 13 BTH cities and 30 provinces. This city-level MRIO table is constructed  
140 with consideration of the flow of domestic intermediate products, re-exports and inter-  
141 city trade flow.

### 142 **3. Methods**

#### 143 *3.1 MRIO analysis*

144 In the 1930s, Leontief proposed an IO method for economic analysis that captured  
145 the relations between sectors and industries (Leontief et al., 1970). There are three types  
146 of IO model: single-region input–output (SRIO), bilateral trade input–output (BTIO),  
147 and MRIO model. These models differ in terms of their system boundaries and research  
148 objectives (Sato et al., 2013). The MRIO model can characterize economic flows  
149 between regions, economic sectors and final demand. Through a combination of  
150 regional and sectoral emission inventories, MRIO can effectively uncover the pollutant  
151 emissions in other regions caused by the consumption of a given region (Wiedmann et  
152 al., 2011). These features make MRIO a popular method to quantify trade activities and  
153 their national or regional environmental impacts (Li et al., 2016; Wang et al., 2017a;  
154 Zhao et al., 2017; Lu et al., 2019; Zhang et al., 2017). Recently, the MRIO model has  
155 been widely used to quantify emissions transfer attributable to trade, because it

156 considers economic output at both the regional and sectorial level, as well as the output  
157 of one region as consumed in another region (Davis et al., 2010; Lenzen et al., 2012;  
158 Wiedmann et al., 2015). The MRIO table is an effective tool for MRIO analysis that  
159 intuitively presents IO monetary flow data in terms of regions and industries.

160 The current study generally uses two parts of the MRIO table: intermediate IO, of  
161 22 sectors among 13 cities located in BTH, and the connections between BTH and 27  
162 other provinces in China. Regarding the latter, we focus on the “final consumption” in  
163 these regions.

164 The MRIO model is constructed based on the connections in the MRIO table. The  
165 balance of trade flow for the entire research system is described by the following  
166 equation (Leontief et al., 1970):

$$167 \quad Q = AQ + Y \quad (1)$$

168 where  $Q$  is the total output of the system;  $A$  is the direct consumption coefficient matrix,  
169 whose elements  $a_{ij}^{rs}$  ( $a_{ij}^{rs} = x_{ij}^{rs} / q_j^s$ ) indicate the amount of intermediate economic input  
170 from sector  $i$  in region  $r$  that produces a unit output for sector  $j$  in region  $s$ ; and  $Y$  is a  
171 vector representing the final demand, including investment, final consumption (i.e.,  
172 household and government consumption) and net export.

173 If  $Y$  is known, then  $Q$  can be calculated using the following equation (Leontief et  
174 al., 1970)

$$175 \quad Q = (I - A)^{-1} Y \quad (2)$$

176 and consumption-based  $\text{NO}_x$  emissions can be obtained as follows:

$$177 \quad E = EI (I - A)^{-1} Y_c \quad (3)$$

178 where  $EI$  is a vector whose elements are defined as the amount of direct  $\text{NO}_x$  emissions  
179 per unit total output,  $(I - A)^{-1}$  is the Leontief inverse matrix, and  $Y_c$  is the final  
180 consumption (Li et al., 2016).



181 Net NO<sub>x</sub> emission flux is calculated as follows:

$$182 \quad netE^{r \rightarrow s} = E^{rs} - E^{sr} \quad (4)$$

183 where  $netE^{r \rightarrow s}$  is the net NO<sub>x</sub> emission from region  $r$  to region  $s$ . When  $netE^{r \rightarrow s} > 0$ ,  
184 regions  $s$  and  $r$  are the receptor and source, respectively, and vice versa (Wang et al.,  
185 2017a).

### 186 3.2 WRF/CALPUFF model

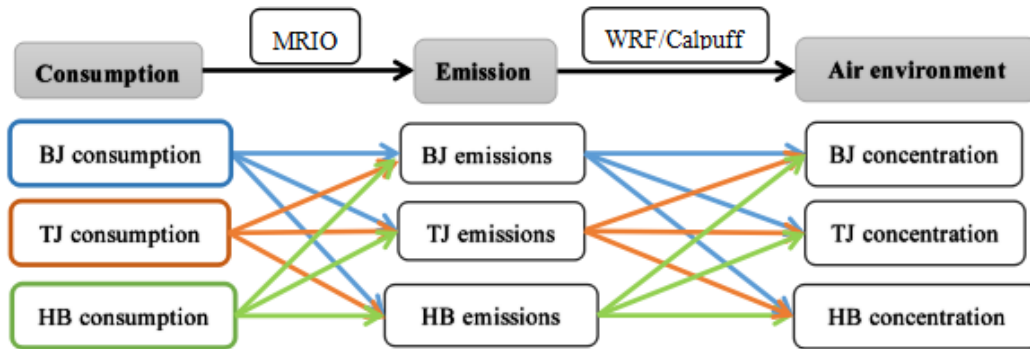
187 WRF/CALPUFF model is used to simulate NO<sub>x</sub> concentrations in the BTH urban  
188 agglomeration. This model is used widely in air quality simulations (Abdul-Wahab et al.,  
189 2011; Dresser et al., 2011; Wu et al., 2018), and has been applied and validated for use  
190 in BTH (Wang et al., 2019). The coupling of WRF and CALPUFF is to simulate the  
191 diffusion process of NO<sub>x</sub> and spatial distribution of NO<sub>x</sub> concentration. The objective of  
192 WRF is to generate the real meteorological field in the simulation area by using the  
193 NCEP/NCAR reanalysis data. The output of WRF is further used as the initial  
194 meteorological field for CALPUFF model. The simulated diffusion process of NO<sub>x</sub> in  
195 CALPUFF model is completed in this meteorological field generated by WRF. Finally,  
196 the spatial distribution of NO<sub>x</sub> concentration is generated according to the set time  
197 interval and spatial accuracy.

198 The data for 2012 January is extracted to represent pollutant diffusion in winter  
199 because winter is the most polluted season (Wang et al., 2014). All parameter settings  
200 and accuracy verification data used in this paper are reported previously (Wang et al.,  
201 2019). The detail parameters setting in WRF/CALPUFF model is provided in  
202 Supporting Information (Table S3 and S4).

### 203 3.3 Combined MRIO and WRF/CALPUFF model analysis

204 The MRIO and WRF/CALPUFF models are combined through the following steps.  
205 For a given city (taking Tianjin as an example), first the MRIO model is used to

206 separate the annual emissions of every cities driven by consumption of Tianjin. And  
 207 then the annual emission data is averaged into hourly emissions. Next, the hourly  
 208 emissions are input to the WRF/CALPUFF model to simulate daily average  
 209 atmospheric pollutant concentrations in each BTH city. A conceptual diagram of the  
 210 MRIO + WRF/CALPUFF model is shown in Figure 2, illustrating the relationships  
 211 among consumption, emissions, and pollutant concentration.



212

213 **Figure 2.** Interactions among economic consumption, NO<sub>x</sub> emissions and concentrations in the BTH  
 214 urban agglomeration. City or province abbreviations are BJ, Beijing; TJ, Tianjin; HB, Hebei.

215

### 216 3.4 Inter-city contributions from a consumption perspective

217 The contribution of consumption-based emissions from other regions to a given  
 218 local region is expressed as follows:

$$219 P^{rs} = E^{rs} / E^s \quad (5)$$

220 where  $P^{rs}$  is the inter-region consumption contribution, referring to the share of  
 221 emissions in region  $s$  originating from consumption in region  $r$ ;  $E^{rs}$  is the emissions in  
 222 region  $s$  originating from consumption in region  $r$ ; and  $E^s$  is the direct emissions  
 223 (production-based) in region  $s$ .

224 After simulating NO<sub>x</sub> concentrations, the contribution of NO<sub>x</sub> concentration is  
 225 calculated as follows:

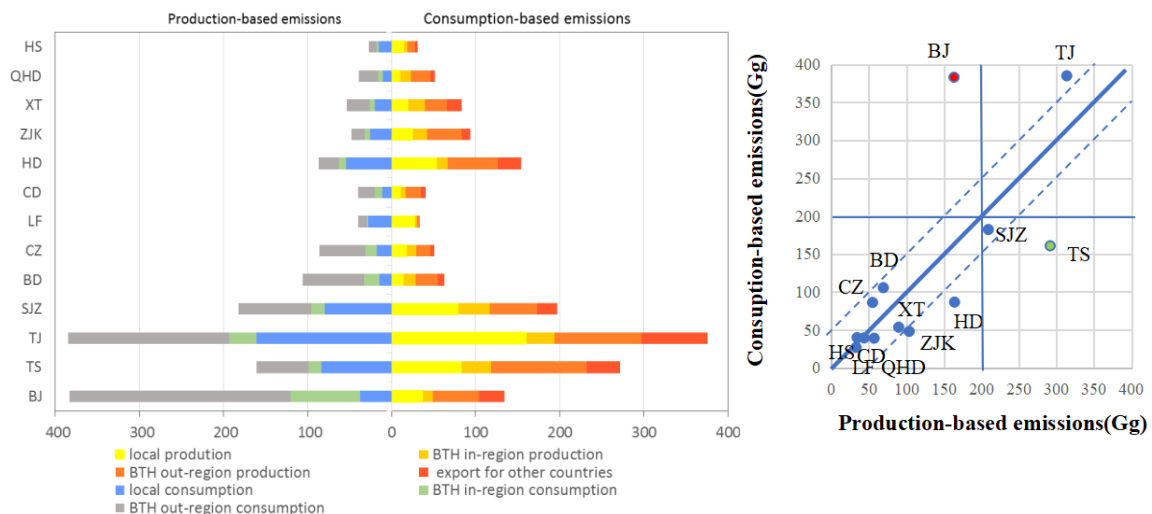
$$226 R^{rs} = C^{rs} / C^s \quad (6)$$

227 where  $R^{rs}$  is the contribution to  $\text{NO}_x$  concentration from interregional consumption,  
 228  $C^{rs}$  is the  $\text{NO}_x$  concentration in region  $s$  due to consumption in region  $r$ , and  $C^s$  is  
 229 the  $\text{NO}_x$  concentration in region  $s$ .

## 230 4. Results

### 231 4.1 Inter-city/province $\text{NO}_x$ emissions virtual transfer only by MRIO model

232 The consumption-based emissions are virtual emissions driven by the  
 233 consumption of local and imported goods, which are derived from final consumption  
 234 output calculated by equation (3). The total production-based emissions amount was  
 235 1,628.06 Gg in BTH, of which originating from BTH in-region consumption accounted  
 236 for 773.12 Gg (47.49%). BTH out-region consumption, from the other 29 provinces of  
 237 China and abroad, accounted for 854.94 Gg of emissions. From a production  
 238 perspective, Tianjin ranked highest, with a total emissions amount of 375.48 Gg,  
 239 followed by Tangshan (271.63 Gg) and Shijiazhuang (197.27 Gg). From a consumption  
 240 perspective, Tianjin and Beijing were the two highest ranking cities in terms of  
 241 consumption-based emissions, due to their huge populations; their consumption-based  
 242 emission data were very similar, at 384.82 and 382.98 Gg, respectively. However,  
 243 Tianjin's proportion of local consumption based on local production reached nearly 50%,  
 244 whereas that of Beijing was only about 20% (Fig. 3a, Table S5).



246

a. Decomposition emissions

b. Scatter plot

247

**Figure 3.** Comparison of production- and consumption-based NO<sub>x</sub> emissions.

248

In total, 8 of the 13 BTH cities were consumption-based emissions receptors; the

249

other 5 cities were consumption-based sources. **However, if these cities are analyzed in**

250

**detail, some new finding can be found.** For most cities, production-based emissions

251

nearly matched consumption-based emissions (Fig. 3b). Take Tianjin and Shijiazhang as

252

examples, the consumption- and production-based emission amounts of these two cities

253

were relatively large due to their large populations and thriving heavy industry sectors.

254

Tianjin is a municipality under the direct control of the Central Government and

255

Shijiazhuang is the capital city of Hebei Province; both cities are consumption- and

256

production-oriented. Only Beijing is a typical consumption-oriented city due to its

257

special economic and political status. In contrast, Tangshan is a typical production-

258

oriented city. Because Tangshan is heavily industrialized and provides other regions

259

with high-emission industrial products (e.g., iron and steel products) (MEPC, 2019b),

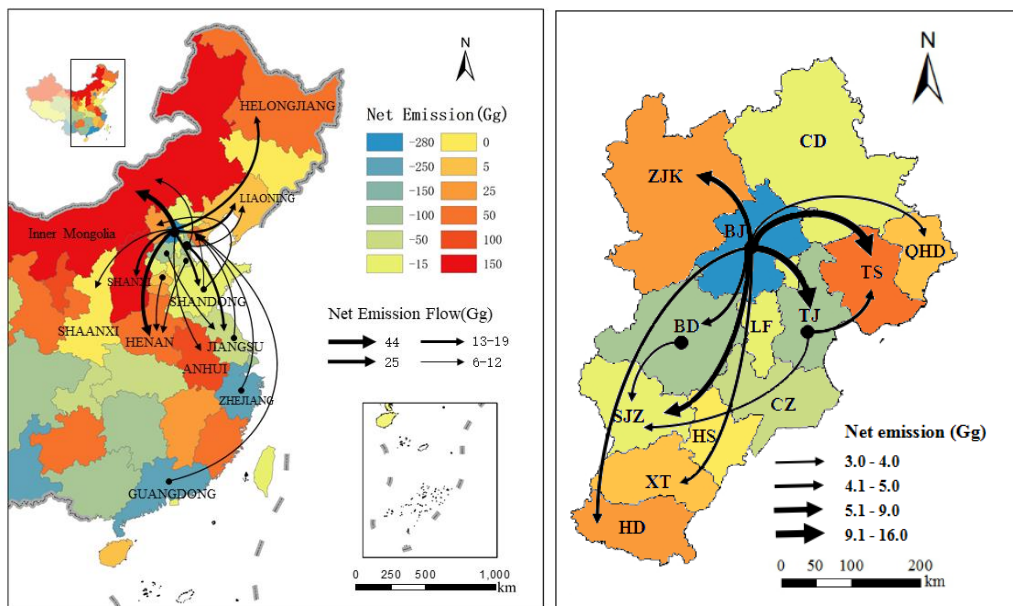
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and its production-based emissions were much larger than its consumption-based

261

emissions (Fig. 3b), with the production-based NO<sub>x</sub> emissions reaching nearly 300 Gg.

262



263

a. Outside BTH region

b. In BTH region

264 **Figure 4.** Consumption-driven virtual NO<sub>x</sub> emission transfer flows. Only flows of > 3Gg between  
265 BTH cities and other provinces are shown.

266 The consumption-driven virtual NO<sub>x</sub> emissions transfer matrix between BTH and  
267 other provinces was calculated using Equation (6) (Table S6). The main net NO<sub>x</sub>  
268 emission flows of BTH cities and surrounding provinces are shown in Fig. 4a. Net  
269 receptor cities, especially Tangshan, were mainly characterized by heavy industry.  
270 Tangshan's industrial activities and associated emissions were influenced by  
271 consumption in other regions, particularly in southeast coastal developed provinces such  
272 as Jiangsu, Zhejiang, and Guangdong. Conversely, Beijing was a major net  
273 consumption-driven emission source. As shown in Fig. 4, the major flows were from  
274 Beijing to other provinces, including Inner Mongolia, Henan, Heilongjiang, Shaanxi,  
275 Shanxi, Liaoning and Shandong. At the same time, some cities in Hebei Province, such  
276 as Baoding, Xingtai and Cangzhou, have transferred virtual emissions to the Central and  
277 East China.

278 Within the BTH region, the major net emission flow was from Beijing to other  
279 cities in the BTH urban agglomeration, such as Tianjin (16 Gg), Tangshan (14.5 Gg),  
280 Baoding (9 Gg) and Shijiazhuang (8 Gg) (Fig. 4b). To satisfy the consumption  
281 requirements of Beijing, these cities emit large amounts of pollutants. Beijing is the  
282 largest consumption-driven source city in BTH, based on virtual emissions.

#### 283 *4.2 Inter-city linkages under the combined effects of economic linkage and atmospheric* 284 *transport*

285 By combining the MRIO and WRF/CALPUFF models, the matrix of  
286 pollutant concentrations induced by consumption is simulated by using Equation (5)  
287 (Fig. 5, Table S9). The contribution to atmospheric environment quality of a city  
288 can be divided into local city itself contribution and non-local contribution (Li et al.,  
289 2016). In this study, the non-local includes "BTH in-region consumption"

290 (consumption in the other 12 cities within BTH) and “BTH out-region  
291 consumption” (consumption in other provinces of China).

292 As Fig. 5 shown, the NO<sub>x</sub> concentrations of most cities are mainly affected by non-  
293 local consumption. According to the combined effect, the 13 cities in BTH region can  
294 be divided into three categories.

295 (1) The I category

296 Tianjin and Shijiazhuang belong to this category. These two cities have the largest  
297 contribution on their local NO<sub>x</sub> concentration among the 13 cities. The reason is the  
298 results of the previous 4.1 section. The two cities have large populations and thriving  
299 heavy industry sectors to induce both large consumption- and production-based  
300 emission.

301 (2) The II category

302 These cities are mainly affected by the inner cities of BTH region, such as  
303 Hengshui, Langfang, Cangzhou and Baoding. It is worth noting that most of these cities  
304 are in the center of Hebei Province. Hengshui was mainly characterized by in-region  
305 consumption (61%), of which Tianjin accounted for 20% and Shijiazhuang 12%. For  
306 Baoding, the in-region consumption was 52%, of which Tianjin accounted for 12%,  
307 Shijiazhuang 10% and Beijing 9%. Tianjin’s consumption was also responsible for 37%  
308 and 28% of the NO<sub>x</sub> in Langfang and Cangzhou, respectively (Fig. 5). The first column  
309 of Fig. 5 indicate that Tianjin’s consumption has not only a great impact on the air  
310 quality of the local environment, but also on the air quality of other cities in the BTH  
311 urban agglomeration. Therefore, Tianjin is an important source city to affect the central  
312 Hebei Province.

313 (3) The III category

314 The third category is opposite to the second category, including Beijing, Tangshan,

315 Chengde, Zhangjiakou, Qinhuangdao, Handan. It is worth noting that most cities locate  
 316 near the border of BTH urban agglomeration, except Beijing and Tangshan. The major  
 317 contributor to their urban air quality is the consumption from BTH out-region due to the  
 318 close distance from the surrounding provinces. For Beijing and Tangshan, as the above  
 319 research shows, the two cities **have** the closest economic ties with the region outside  
 320 BTH. The former is a typical consumption-oriented city, the latter is a production-  
 321 oriented city. As a result, like other border cities, they are more affected by outside.

		Consumption-based source													In-region	Out-region		
		TJ	SJZ	HS	CZ	LF	BD	XT	BJ	TS	CD	ZJK	QHD	HD				
Consumption-based receptor	TJ	47	1	0	1	2	1	0	7	6	0	0	0	0	18	35	I	
	SJZ	4	36	1	2	4	3	2	5	2	1	1	1	2	28	36		
	HS	20	12	30	6	3	2	3	6	3	1	0	0	5	61	9	II	
	CZ	28	5	3	20	7	2	1	5	5	1	1	1	2	61	19		
	LF	37	1	0	4	30	1	1	7	6	0	1	0	1	59	11		
	BD	12	10	1	2	8	13	1	9	4	1	1	1	2	52	35	III	
	XT	6	17	3	2	1	2	15	5	2	1	1	1	9	50	35		
	BJ	13	1	0	1	3	1	0	21	3	0	2	0	0	24	55		
	TS	9	1	0	1	1	1	0	5	28	0	0	1	1	20	52		
	CD	6	1	0	1	1	1	0	7	10	14	3	0	1	31	55		
	ZJK	2	2	0	0	0	1	0	9	2	1	23	0	1	18	59		
	QHD	8	2	0	1	0	2	1	6	13	1	1	12	1	36	52		
	HD	3	5	1	1	0	1	4	3	1	0	0	0	30	19	51		

322  
 323 **Figure 5.** NO<sub>x</sub> concentration contributions of the BTH cities. Each cell in the grid shows the  
 324 contribution of a source city's economic consumption to the NO<sub>x</sub> concentration in the receptor city.  
 325 The unit of numbers in each grid is %. Darker colors indicate greater contributions.

326

### 327 4.3 Comparison of the results

328 The average value of the three kinds of linkages contribution analysis is compared  
 329 (Fig. 6). The first contribution analysis result is only from the economic linkage  
 330 perspective (determined by MRIO only). This result is from our research. The result

331 shows that the main contribution is from the local city and BTH out-region about 40%  
332 respectively. The contribution of BTH in-region cities is smallest (average 20%).

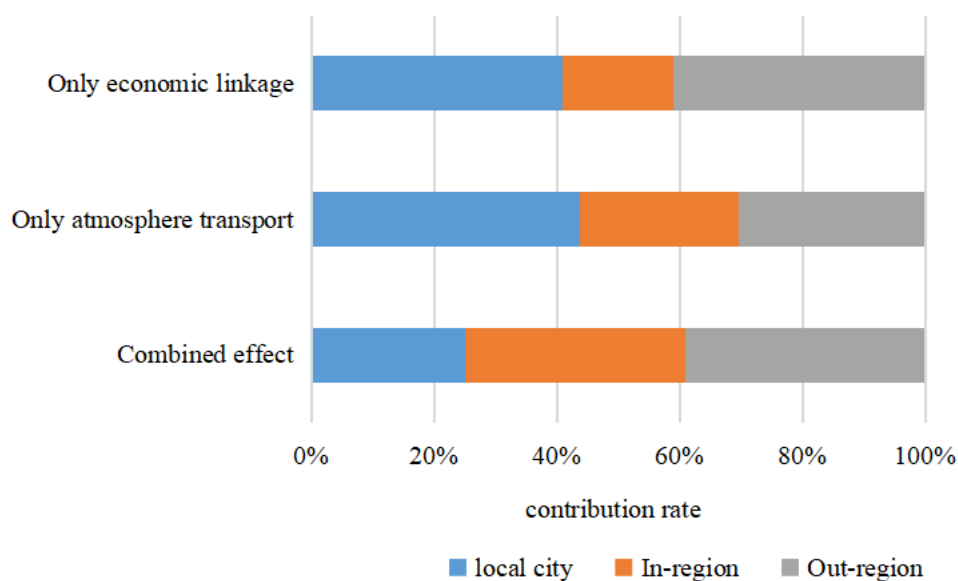
333 The second kind of contribution analysis is only from the atmosphere transport  
334 perspective. This kind contribution analysis result is based on the research results of the  
335 application of CAMx-PSAT air quality model by Wang et al. (2017b). Their results  
336 show that the major contributor to urban air quality is emissions of local cities, about  
337 40% on average. It happens to be similar to the first result of only considering economic  
338 relation. The other research also shows similar results. For example, Li et al. (2015)  
339 used CAMx-PSAT to quantify the contribution of PM<sub>2.5</sub> concentration in the BTH  
340 region in 2006 and 2013. Their results show local emissions make the largest  
341 contribution (40%-60%) for all receptors. Chang et al. (2019) used WRF-CMAQ  
342 modeling system to simulate the air quality in the BTH region. Their results show  
343 annual averaged local contribution ranges from 32% to 63% for the 13 cities in the BTH  
344 region. However, the contribution of BTH in-region is bigger than that of the first result,  
345 about close to 30%. This phenomenon shows that the contribution of BTH in-region  
346 will increase if atmospheric transport is considered due to the close physical distance  
347 between cities in BTH region.

348 The third kind of contribution analysis is the results of combined effect of  
349 economic linkages and atmosphere transport. This result shows that local contributed  
350 less than 30% in most cities. For all cities, more than 50% of the NO<sub>x</sub> concentration  
351 induced by other area contribution. These findings indicate that other area is an  
352 important contributor in the BTH urban agglomeration at the city level. BTH out-  
353 regions were responsible for 39% of the NO<sub>x</sub> concentration. BTH in-region  
354 consumption and city's own consumption made contributions 36% and 25% on average,  
355 respectively. This result indicates that the contribution from local city decreased when



356 combined effect was considered.

357 Therefore, the unique characteristic captured in this paper is when the MRIO and  
358 air quality models are combined, the relative contribution of local urban decreases and  
359 the non-local contribution from the surrounding cities and other area becomes more  
360 important.

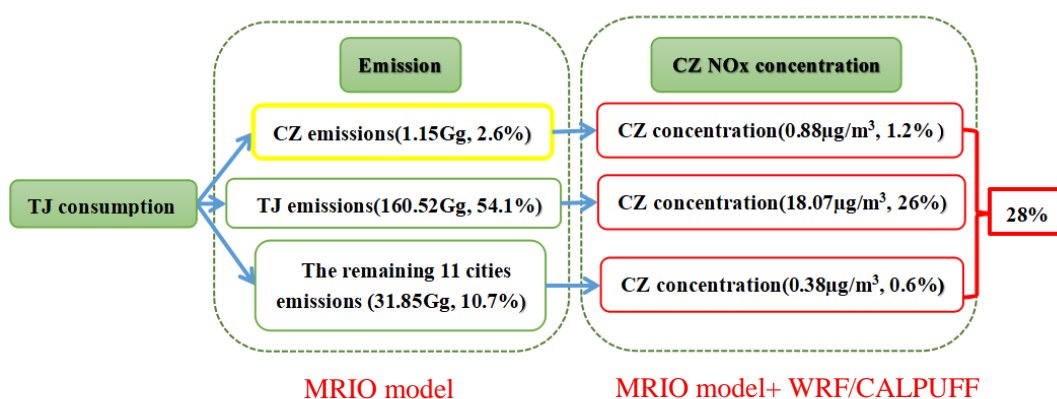


361

362 **Figure 6.** Comparison of the three kinds of contribution analysis. The first kind of contribution  
363 analysis is considered only economic linkage based on MRIO model, the second kind of contribution  
364 analysis is considered only atmosphere transport based on air quality model according to the  
365 reference(Wang et al., 2017b), the third kind of contribution analysis is considered the combined  
366 effect of economic linkage and atmosphere transport.

367 To further explain the atmospheric transport mechanism, the impact of economic  
368 consumption of Tianjin on the pollutant concentration in Cangzhou is assessed as an  
369 example. As shown in Fig. 7, consumption in Tianjin led to production in Cangzhou  
370 (associated with 1.15 Gg NO<sub>x</sub> emissions; 2.6% of all Cangzhou emissions), as well as in  
371 Tianjin (160.52 Gg NO<sub>x</sub> emissions) and other regions (31.85 Gg NO<sub>x</sub> emissions among  
372 the other 11 BTH cities). The contribution of emission in Cangzhou driven by  
373 consumption of Tianjin was only 0.88 μg m<sup>-3</sup>. However, Tianjin's consumption also led

374 to emissions in Tianjin itself (contribution of  $18.07 \mu\text{g m}^{-3}$  to the  $\text{NO}_x$  concentration in  
 375 Cangzhou through atmospheric transport). Similarly, consumption in Tianjin led to  
 376 emissions in the remaining 11 cities, contributing  $0.38 \mu\text{g m}^{-3}$  to the  $\text{NO}_x$  concentration  
 377 in Cangzhou. Cumulatively, these three sources contributed 28% of the  $\text{NO}_x$   
 378 concentration in Cangzhou. This is the real linkages between Tianjin and Cangzhou  
 379 about air pollution.



380

381 **Figure 7.** The impact of economic consumption in Tianjin (TJ) on the  $\text{NO}_x$  concentration in  
 382 Cangzhou(CZ). Yellow box indicates the contribution of consumption in Tianjin on the emissions in  
 383 Cangzhou, i.e., Tianjin consumption→Cangzhou  $\text{NO}_x$  emissions (2.6%) determined using only the  
 384 MRIO model. Red boxes indicate the contribution of consumption in Tianjin to the pollutant  
 385 concentration in Cangzhou, i.e., Tianjin consumption→Cangzhou  $\text{NO}_x$  concentration (28%), as  
 386 calculated using the MRIO and WRF/CALPUFF models. For data on the other cities, see the  
 387 Supporting Information (Figure S2).

388

### 389 5. Conclusion and Policy implications

390 In this paper, take the BTH region as example, when the combined effect of  
 391 economic linkages and atmospheric transport is considered, the result shows the non-  
 392 local contribution is greater than that only considering atmospheric transport or only  
 393 considering economic correlation.

394 These atmospheric and economic relationships should be taken into account in  
 395 air pollution control policies. It suggests that governments should exercise cautious  
 396 when formulating industrial emission reduction targets and policies, because emissions

397 reduction in one city may affect both the air quality and economic consumption in  
398 other cities. Thus, air pollution policies should consider both economic and  
399 atmospheric relations among cities to reduce the negative impacts of emission  
400 reductions on living standards and the economy. From a perspective of regional  
401 collaboration, Beijing and Tianjin, as two metropolises of consumption and production,  
402 are suggested to give support to other cities in Hebei Province in terms of air pollution  
403 control technology and funds. Hebei Province should avoid receiving the polluted  
404 enterprises from Beijing or Tianjin. The government should seize the two-source  
405 control means of new source environmental access and old source backward  
406 production capacity elimination to speed up industrial upgrading. Moreover, it is  
407 suggested to establish the emergency emission reduction scheme for heavy pollution in  
408 winter and the intercity air pollution joint prevention and control plans from the  
409 perspective of differentiation and economic influence. Furthermore, the establishment  
410 of investigation mechanism of long-distance transmission of fixed sources, the  
411 performance evaluation mechanism and financial mechanism are suggested to  
412 guarantee these plans.

413 Equalizing the spatial and temporal resolution of economic and atmospheric data  
414 is difficult; improving the accuracy of this process would improve the quality of  
415 research in this area. The limitation of this paper is the lack of more detailed input data  
416 for the atmospheric transport simulation. Because the basic scientific research on the  
417 mechanism of chemical transformation of air pollutants in BTH area is not mature  
418 enough, this study lacks the atmospheric chemistry simulation. But this article focuses  
419 on the relationship between economy and environment. The method in this paper  
420 represents an improvement of the application of input-output model in air pollution  
421 research. Future studies can apply the methods proposed in this study to explore a

422 wider range of air pollutants, and their sources, to provide a scientific basis to optimize  
423 emissions reduction targets at the city level and formulate rational air pollution control  
424 policies.

425

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431

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