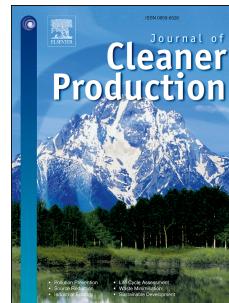


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**Efficiency assessment of rural domestic sewage treatment facilities by a
slack-based DEA model**

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2 **Abstract**

3 In the context of sustainable development, a number of rural domestic sewage
 4 treatment facilities had been built in China to solve the problem of rural domestic
 5 sewage pollution. The comprehensive, quantitative and objective efficiency
 6 assessment of facilities is urgent. This study used a non-radial slacked-based data
 7 envelopment analysis model combined with cluster analysis to construct an index
 8 system covering multiple aspects, including three inputs and four outputs to assess
 9 681 facilities. These samples selected from the biggest demonstration area are the most
 10 representative for and exceed 2/5 of the running facilities all over the country. The
 11 average efficiency score of samples was 0.496 meaning the improvement potential
 12 was about 50.4%. Only 27 samples were relatively effective, scoring 1. The remaining
 13 654 facilities had different levels of input excesses or output shortfalls, which should
 14 be the key objects to improve overall performance. In addition, there was evidence
 15 that output indicators had more room for improvement than input indicators. The
 16 analysis of sensitivity on inputs and outputs confirmed that the idleness and poor
 17 treatment effects of rural sewage treatment facilities should be concerned. Finally,
 18 Kruskal-Wallis non-parametric test verified that technology and load rate of facilities
 19 have significant impacts on efficiency. The performance evaluation results could not
 20 only provide guidance for the local government to strengthen the supervision and
 21 operation of facilities, but also potentially provide reference for the construction,
 22 operation and management of rural sewage treatment facilities in China.

23 *Keywords:* Data envelopment analysis, Efficiency assessment, Rural domestic
 24 sewage, Potential improvement, Sensitivity analysis, Explanatory factors

25 **1 Introduction**

26 In recent years, the water pollution has become a serious challenge to the
 27 development of rural areas. By 2015, the direct emission of rural domestic sewage was
 28 about 20 million tons every day. The annual chemical oxygen demand (COD) emission
 29 was about 10.69 million tons and the annual ammonia nitrogen emission was 0.73
 30 million tons (China Environmental Statistics Annual Report, 2015). Due to economic

31 and geographical factors, the coverage of treatment facilities is extremely low in most
32 rural areas of China. 96 % of rural villages cannot effectively treat sewage (Gu et al.,
33 2016). To control water pollution in rural areas, the central government had proposed
34 an ambitious plan, that the treatment coverage in rural area will reach 33.6% by 2020. A
35 few rural domestic sewage treatment projects have been set up and demonstrated in key
36 river valleys (Chen et al., 2018; Wu et al., 2011). Although certain progress has been
37 made, the existing rural sewage treatment facilities have problems such as scattered
38 locations, jagged technical levels and weak supervision. Thus, it is urgent to evaluate
39 performance of existing facilities and answer which is the best.

40 The environmental performance evaluation proved to be an effective and suitable
41 environmental management tool to find out the problems existing in rural sewage
42 treatment facilities (Alemany et al., 2005; Benedetti et al., 2008; Gallego et al., 2008). It
43 can help the local governments and sewage companies formulate reasonable policies to
44 promote the effective development of rural sewage treatment facilities, and also to
45 provide targeted improvement recommendations. Kalbar et al. (2012) assessed the
46 applicability of 4 common rural sewage treatment technologies in India based on
47 scenario analysis. Xia et al. (2012) evaluated treatment technologies from the economic
48 and technical aspects by the fuzzy advantages and disadvantages coefficient method in
49 a village of Changzhou. Shen et al. (2014) combined the analytic hierarchy process
50 with the entropy method to select 10 advanced technologies from 15 commonly used
51 rural domestic sewage treatment technologies. The existing research mainly focused on
52 the simple evaluation of the treatment technology. Besides, artificially assigning
53 weights to indicators led to subjective errors. More importantly, these methods failed to
54 distinguish inefficient from efficient facilities and quantify the improvement potential.

55 Data envelopment assessment (DEA) has been widely used in the performance
56 evaluation of water sector in recent years (Dong et al., 2017; Hu et al., 2019; Jiang et al.,
57 2020). This method obtains relative efficiency of decision-making units (DMUs) with
58 multiple inputs and multiple outputs based on linear programming (Mostashari-Rad et
59 al., 2019). A significant advantage of the DEA method is that it is not necessary to
60 assume a correlation between input and output indicators (Hosseinzadeh-Bandbafha et

al., 2018). Thus, the evaluation results are objective. Traditional DEA models are radial, which fail to calculate the theoretical target values of inputs and outputs for inefficient plant (Gómez et al., 2017; Lombardi et al., 2019). The slack-based measure (SBM) model proposed by Tone (2002) perfectly solved this problem. On other hand, SBM-DEA model can be combined with clustering analysis to minimize the impact of scale effect on plants performance.

In this context, this study selected SBM-DEA model based on clustering analysis to evaluate the efficiency scores of 681 facilities in rural area of Wuxi district, Jiangsu Province, located in southeastern China. As the biggest demonstration area, these samples are the most representative for and exceed 2/5 of the running facilities all over the country. The purpose of the study is (1) to evaluate the performance efficiency of 681 rural sewage treatment facilities; (2) to identify the improvement potential of inefficient facilities and provide specific improvement suggestions; (3) to identify implicit factors that affect the facility performance. The results can help select out the state-of-art for the construction, operation and management of rural sewage treatment facilities in China, effectively promoting the sustainable development of rural water resources.

2 Methodology

2.1 SBM-DEA model

DEA is a powerful non-parametric comprehensive evaluation method to measure relative efficiency of a large number of decision-making units (DMUs) (Nabavi-Peleesarai et al., 2019). This method selects the efficient DMUs as reference benchmark to identify levels and causes of inefficient DMUs. Different DEA models had been proposed for different purpose. At present, conventional radial models, such as Charnes-Cooper-Rhodes (CCR) and Banker-Charnes-Cooper (BCC) have been widely used. However, these models assume changes of inputs or outputs are proportional, failing to consider the slack of indicators (Carvalho and Marques, 2011).

By comparison, non-radial SBM-DEA model is more suitable for assessing samples with vague interconnections inputs (Thrall, 1996). It considers input excesses and output shortfalls of DMUs further, providing target improvement value for each

inefficient DMU's input and output separately (Castellet and Molinos-Senante, 2016; Wang et al., 2018). What's more, this method can treat environmental impacts as undesirable outputs in the index system to achieve a multi-dimensional assessment of the environment impacts, resources consumption and service value (Guo et al., 2017; Robaina-Alves et al., 2015). Finally, SBM model can be combined with clustering analysis by grouping samples according to the design treatment capacity to evaluate the sample efficiency based on the group-frontier, so as to reduce the impact of scale effect (Jiang et al., 2020).

Based on the above reasons, this study composed an output-oriented SBM-DEA model based on constant scale return (CRS) combined with cluster analysis to evaluate rural sewage treatment facilities. Suppose the number of DMUs is n and each DMU has m inputs and s outputs. The matrices are expressed as $X=[x_{ij}] \in R^{m \times n}$ and $Y=[y_{ij}] \in R^{s \times n}$. The fractional programming form of SBM model is shown as follows:

$$\begin{aligned} \min \rho^* &= \frac{1 - \frac{1}{m} \sum_{i=1}^m s_i^- / x_{i0}}{1 + \frac{1}{s} \sum_{r=1}^l s_r^+ / y_{r0}} \\ &\text{s.t.} \\ &x_0 = X\lambda + s^- \\ &y_0 = Y\lambda - s^+ \\ &\lambda \geq 0, s^- \geq 0, s^+ \geq 0 \end{aligned} \quad (1)$$

where s^- and s^+ represent the input excess and output shortfall, respectively. λ indicates non-negative weight vector. The value of ρ^* ranges from 0 to 1. The higher the value of ρ^* , the better the efficiency of the DMU. When $\rho^*=1$, the DMU is relative efficient means no input excess and output shortfall. Otherwise, the DMU is inefficient. Inefficient DMUs can improve score by decreasing input excesses or making up output shortfalls as follows:

$$x_0 - s^- \rightarrow x'_0, y_0 + s^+ \rightarrow y'_0 \quad (2)$$

Traditionally, DEA model assumed all samples to have the same or similar characteristics when efficiency is evaluated. Therefore, all DMUs were taken as reference set to construct meta-frontiers. In reality, the DMUs not always are homogeneity, which will affect the accuracy of DEA results (Corton and Berg, 2009). Clustering analysis approach can usefully deal with heterogeneous DMUs (Galar et al.,

118 2014). This method divides DMUs into different groups according to certain attributes,
 119 which can maximize the homogeneity of samples in the same cluster to decrease the
 120 effect of heterogeneity on efficiencies. Then, every group takes itself as reference set,
 121 constructing group-frontier separately.

122 The definition of meta-frontier and group-frontier according to output sets and
 123 output distance functions (BATTSESE et al., 2004; O'Donnell et al., 2007) are as
 124 follows. Assume y and x are the output and input vectors of dimension $X \times I$ and $Y \times$
 125 I , respectively. All DMUs make up the meta-technology set:

$$126 T^{meta} = \{(x, y) | x \geq 0; y \geq 0: x \text{ production } y\}$$

127 The corresponding output set P for input vector can be defined as:

$$128 P^{meta}(x) = \{y/(x, y) \in T^{meta}\}$$

129 The upper bound of this set is the meta-frontier. At this time, meta-distance
 130 function can be expressed as:

131 $D^{meta}(x, y) = \inf_{\theta} \{\theta > 0: (y/\theta) \in P^{meta}(x)\}$, if and only if $D^{meta}(x, y) = 1$, the
 132 DMU is efficient.

133 Similarly, if all samples are divided into subgroups according to specific criteria,
 134 the DMUs in the k th group are contained in the group-specific technology set:

$$135 T^k = \{(x, y) | x \geq 0; y \geq 0; x \text{ production } y\}$$

136 The corresponding output set P for input can be defined as:

$$137 P^k(x) = \{y/(x, y) \in T^k\}$$

138 The upper bound of this set is the group-frontier. At this time, group-distance
 139 function can be expressed as:

140 $D^k(x, y) = \inf_{\theta} \{\theta > 0: (y/\theta) \in P^k(x)\}$, if and only if $D^k(x, y) = 1$, DMU is
 141 efficient.

142 $D^{meta}(x, y) \leq D^k(x, y)$, $TE^{meta}(x, y) \leq TE^k(x, y)$, which means the meta-frontier
 143 envelops the group-frontier. The difference between results based on two frontiers can
 144 be measured by technical gap rate (TGR):

$$145 TGR^k(x, y) = TE^{meta}(x, y)/TE^k(x, y) \quad (3)$$

146 The value of TGR ranges from 0 to 1. Assuming that TE^{meta} is 0.6 and TE^k is 0.8,
 147 the TGR would equal 0.75. This means that if the input vector is determined, the

148 maximum output that could be produced by a form group k is 75% of the output that
 149 is feasible when using the meta-frontier as a benchmark. The higher value of TGR, the
 150 smaller gap between the meta-frontier and group-frontier and the smaller gap between
 151 technology used by the DMU and technology frontier.

152 **2.2 Data collection and variables**

153 **2.2.1 Data source**

154 This study investigated 681 rural sewage treatment facilities in Wuxi, Jiangsu
 155 Province. All facilities removed contaminants by conventional secondary treatment,
 156 ensuring the comparability fundamentally. The electricity consumption and water
 157 quality data were sampled once a month. In this study, the monthly average data of
 158 2017 was used as the benchmark. The investment and operational data come from the
 159 information system of Wuxi Wastewater Treatment Authority.

160 **2.2.2 Inputs and outputs**

161 DEA is a data-oriented method, thus, selecting appropriate inputs and outputs is
 162 the key to accurately evaluate relative performance efficiencies of samples. In order to
 163 comprehensively evaluate the performance of rural sewage treatment facilities for
 164 construction, operation and management, an index system should be constructed from
 165 multiple dimensions such as economy, environment and energy consumption. It
 166 should be noted that the more variables, the more difficult to distinguish DMUs
 167 performance because the number of efficient DMUs increases. This study referred to
 168 the indicators selected by the previous researches (Lorenzo-Toja et al., 2015;
 169 Sala-Garrido et al., 2011; Wang et al., 2018) of sewage treatment plants evaluation
 170 and takes into account the availability of data and the characteristics of the selected
 171 model. The minimum number of indicators was selected to ensure the integrity of the
 172 evaluation elements. The units of the input and output variables do not affect the
 173 efficiency score.

174 The necessary inputs had been grouped into three categories: (1) capital cost (x_1 ,
 175 10^4 CNY); (2) operating cost: mainly including labor cost and maintenance cost (x_2 ,
 176 10^4 CNY/year); (3) electricity consumption: the largest energy consumption of

177 operation (x_3 , 10^4 kWh/year). These indicators really reflected resource consumption
 178 of rural sewage treatment facilities.

179 Four operational indicators had been chosen as outputs: (1) treatment capacity
 180 (y_1 , 10^4 ton/year); (2) chemical oxygen demand removed (COD, ton/year) (y_2); (3)
 181 ammonia nitrogen removed (NH_3-N) (y_3 , ton/year); (4) total phosphorus removed (TP)
 182 (y_4 , ton/year). The selection of outputs reflected the service value of rural sewage
 183 treatment facilities to improve the quality of rural water environment by treating
 184 sewage discharged.

185 **2.2.3 Implicit explanatory factors**

186 In addition to the selected three input factors and four output factors, the
 187 performance of the DMUs may also be affected by many other implicit factors. To
 188 further determine the best operating conditions, the next step is to identify the implicit
 189 factors. Based on the reported studies and the available statistical information, another
 190 three factors were considered (Molinos-Senante et al., 2013; Teklehaimanot et al.,
 191 2015; Zeng et al., 2017) : (i) technology, (ii) load rate: expressed as the ratio of the
 192 actual treatment capacity to the designed treatment capacity, (iii) standard of
 193 discharge.

194 **3 Results and discussion**

195 **3.1 Sample description**

196 Previous studies confirmed that scale has significant impacts on the efficiency
 197 scores of sewage treatment facilities: the plants with larger size operate more
 198 effectively (Dong et al., 2017; Hernández-Sancho and Sala-Garrido, 2009). To
 199 minimize scale effect, the DMUs were divided into five groups according to design
 200 treatment scale of facilities: group 1 ([0, 5) t/d), group 2 ([5, 10) t/d), group 3 ([10, 20)
 201 t/d), group 4 ([20, 30) t/d) and group 5 ([30, 80) t/d). A brief description of the inputs
 202 and outputs was listed in Table 1. With the increase of the treatment scale, the average
 203 values of three inputs and four outputs also gradually increased. The degree of data
 204 dispersion (standard deviation) did not show obvious rules.

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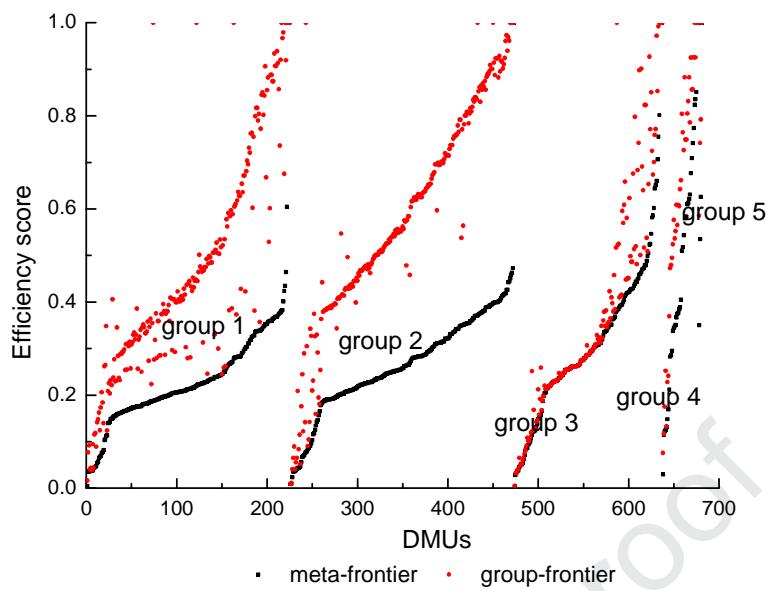
Table 1 The descriptive statistics of the variables for five groups.

group	variables	x_1 (10^4 CNY)	$x_2(10^4$ CNY/year)	$x_3(10^4$ kWh/year)	$y_1(10^4$ t/year)	$y_2(t/year)$	$y_3(t/year)$	$y_4(t/year)$
1	average	4.316	0.979	0.079	0.109	0.181	0.073	0.004
	stdev	1.390	0.027	0.0270	0.039	0.143	0.034	0.003
2	average	10.000	1.074	0.175	0.254	0.3854	0.166	0.008
	stdev	0.000	0.018	0.018	0.033	0.183	0.058	0.003
3	average	17.592	1.335	0.278	0.446	0.632	0.279	0.016
	stdev	2.567	0.181	0.041	0.088	0.326	0.1189	0.020
4	average	26.775	1.620	0.420	0.754	1.192	0.488	0.026
	stdev	1.143	0.047	0.047	0.111	0.731	0.177	0.014
5	average	51.500	2.101	0.901	1.661	2.491	1.246	0.058
	stdev	20.809	0.398	0.398	0.960	1.441	0.720	0.034

207 **3.2 Efficiency analysis and potential improvement**208 **3.2.1 Efficiency scores**

209 In this study, The SBM model based on CRS and group-frontier was established
 210 by MaxDEA Ultra 8 (No 812-182) software. Detailed data and results could be found
 211 in Table S1 in Appendix. The average TGRs of the five groups ranged from 0.477 to
 212 0.898, indicating that the gap between the two frontiers was obvious.

213 Fig. 1 compared the efficiency scores of 681 DMUs under group-frontier with
 214 the scores based on the meta-frontier. Based on the group-frontier, the number of
 215 DMUs with high scores (> 0.5) increased significantly and the number of efficient
 216 facilities (score equals to 1) increased from 10 to 27. This result verified the necessity
 217 of evaluating operating performance of rural sewage treatment facilities under
 218 different scale frontiers. Therefore, the following analysis in the study was all based
 219 on group-frontier.



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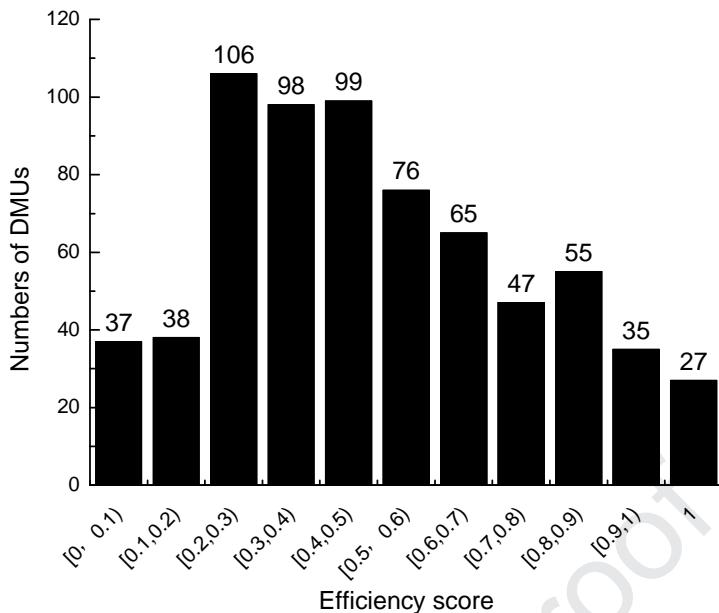
Fig. 1. Efficiency scores of 681 treatment facilities based on meta-frontier and group-frontier

222 respectively.

223

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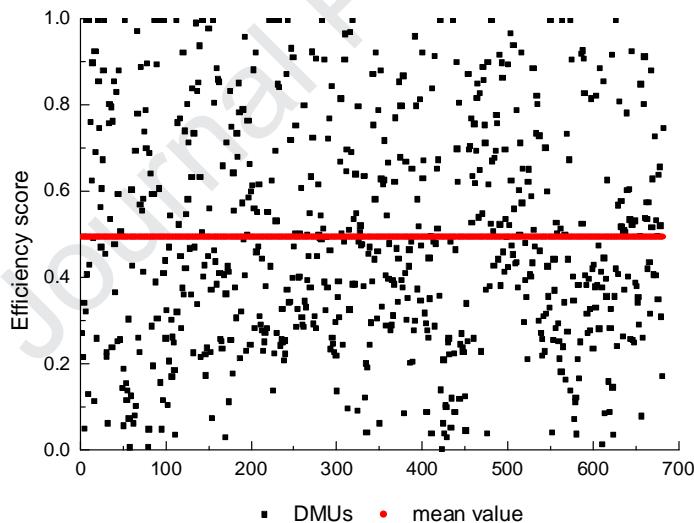
Fig. 2 showed the number of facilities at different subintervals of efficiency scores based on group-frontier. 27 treatment facilities were relatively efficient, meaning that less than 4% of DMUs located on the optimal production frontier, i.e., maximizing outputs. Considering these treatment facilities as the best benchmark, nearly half of samples (305 out of 681) scored less than or equal to 0.5, which meant that there was great room for improvement in the inefficient facilities. Fig. 3 showed that the average score of the samples was 0.496, so the inefficient DMUs had about 50.4% improvement potential. Thus, how to optimize the allocation of inputs and outputs of inefficient DMUs should be the focus to improve the overall efficiency scores of treatment facilities.



233

234

Fig. 2. The number of treatment facilities at different subintervals of efficiency scores based on group-frontier.



236

237

Fig. 3. Efficiency scores of 681 DMUs based on group-frontier.

3.2.2 Potential improvement

As shown in Fig. 4, the difference in the capital and operational costs between inefficient DMUs and efficient DMUs were not significant, showing that the investment of construction and operation for all facilities was overall reasonable. The mean electricity of inefficient DMUs (2379.181 kWh/year) was higher than that of inefficient DMUs (1843.836 kWh/year). Significant output shortfalls existed in

inefficient samples. The average values of four output variables for the efficient DMUs were obviously higher than those for the inefficient DMUs. The average annual treatment capacity of efficient plants was 4,541.963 tons, while that of inefficient plants was only 2,831.661 tons. Furthermore, the pollutants removal of an efficient treatment facility was 2 to 3 times that of an inefficient facility.

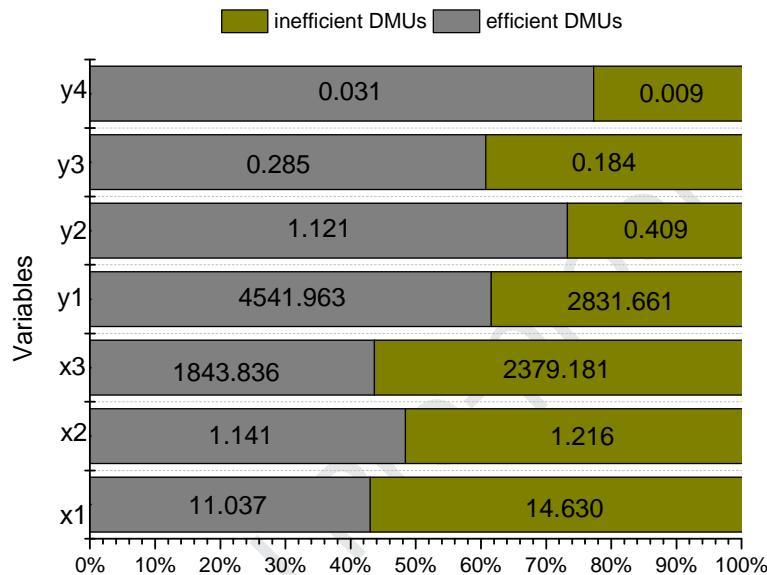
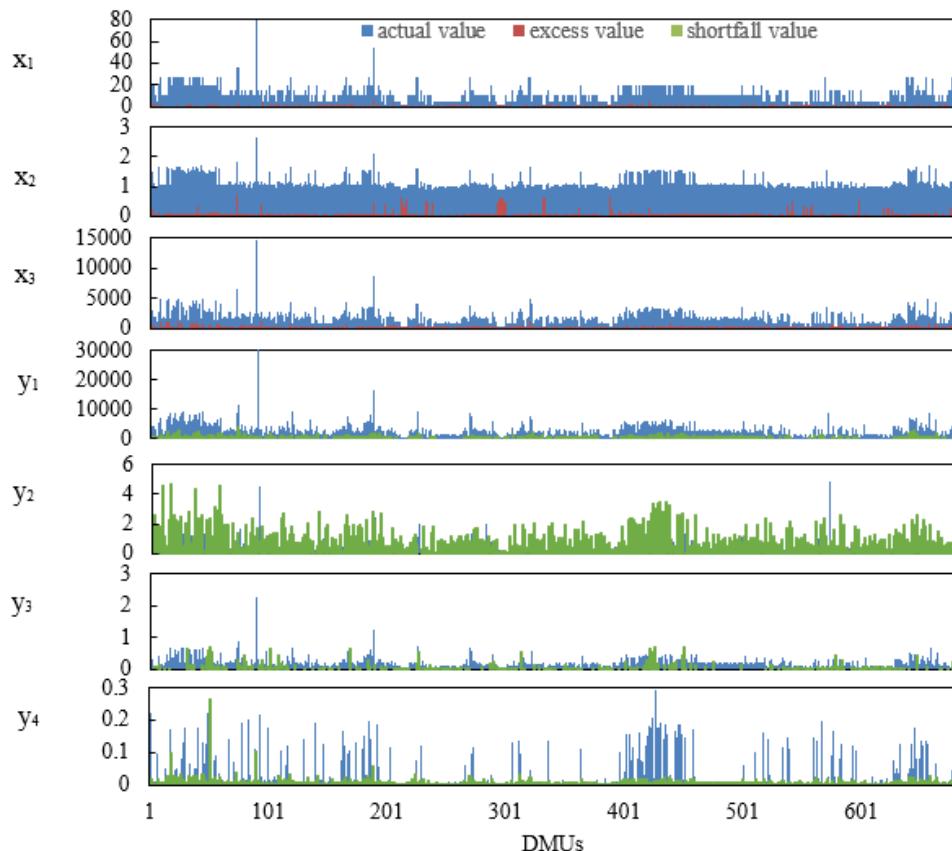


Fig. 4. Comparison of the inputs and outputs for the efficient and inefficient DMUs.

SBM model directly constructs slack variables in the objective function to take the slack of the inputs and the outputs into account. In other words, taking efficient samples as benchmark, it can quantify potential improvement of each item for inefficient DMUs to improve scores of inefficient facilities. The results were shown in Fig. 5 and Table 2. The level of output shortfall in 654 inefficient treatment facilities was serious. For these samples, the treatment capacity (y_1) had the greatest improvement potential, which could improve about 92.45% (1.61×10^5 ton/year) under the current input level. Moreover, the potential improvement for the removal of COD, $\text{NH}_3\text{-N}$ and TP was 45.49% (357 ton/year), 91.97% (11 ton/year) and 25.33% (20 ton/year) respectively. Under the current output level, the capital cost, the operating cost and electricity consumption could be respectively reduced by 1.74% (130×10^4 CNY), 4.67% (35×10^4 CNY/year) and 8.60% (1.06×10^5 kWh/year). There was almost no input excess. Therefore, the manager of the plants should focus on solving problems of low load operation and poor removal of pollutants.



265

266

Fig. 5. Potential improvement of each item for every DMU.

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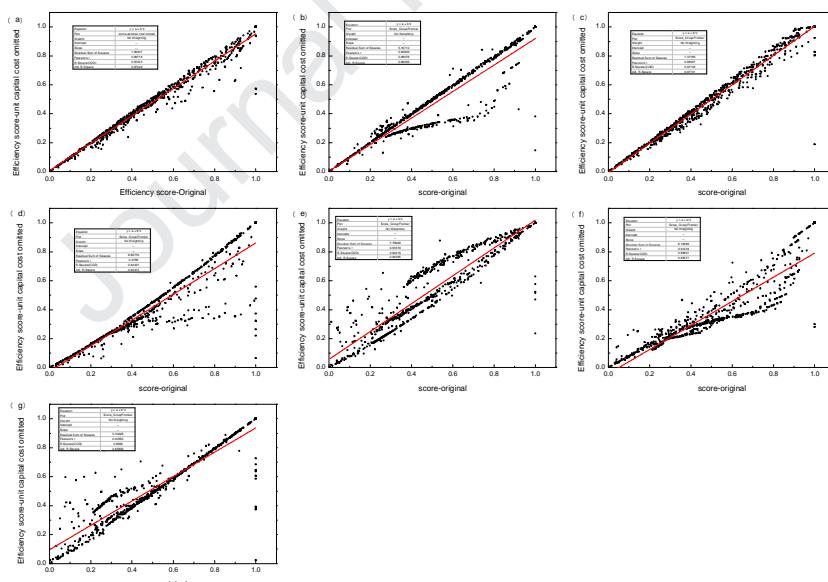
Table 2 The mean improvement potential of 681 DMUs.

	capital cost (10^4 CNY)	operating cost (10^4 CNY /year)	electricity consumption (10^4 kWh/year)	treatment capacity (10^4 ton/year)	COD removed (ton/year)	NH ₃ -N removed (ton/year)	TP removed (ton/year)
origin value	7483	744	117	0.214	0.654	0.139	0.027
target value	7613	779	127	0.197	0.298	0.128	0.007
slack movement	-130	-35	-10	0.017	0.357	0.011	0.020
potential for improvement	-1.74%	-4.67%	-8.60%	92.45%	45.49%	91.97%	25.33%

268 **3.3 Sensitivity analysis of inputs and outputs**

269 The efficiency scores of DMUs are influenced directly by the change of inputs
 270 and outputs because each vector introduced uncertainty into DEA model (Castellet
 271 and Molinos-Senante, 2016). Changing the input or output of the DMUs to observe
 272 the changes in efficiency is the main sensitivity analysis method (Hu et al., 2019).
 273 SBM model, as a non-parameter model, the efficiency score has no specific
 274 quantitative relations with the number of inputs and outputs (Guo et al., 2017). Thus,

omitting one input or one output variable once time to examine degree of change in efficiency score is an effective approach for sensitivity analysis. Fitting the scatters to calculate slope and coefficient of correlation (R^2) of proportional function. Then, the sensitivity of the variables can be identified by the gap between 1 and slope of the function (Hu et al., 2019). The greater the gap, the higher the sensitivity. Fig. 6 and Table 3 showed the result of sensitivity analysis of seven variables. Omitting the variable (y_3), the highest value of $|1-\text{slope}|$ (0.167) occurred, indicating the removal of $\text{NH}_3\text{-N}$ (f) was the most sensitive factor. Other significant factors include TP removed (g), treatment capacity (d) and operating cost (b). The electricity consumption was the least sensitive factor mainly because of the small difference in power consumption of treatment facilities at the same scale. Overall, the outputs were more sensitive than the inputs. Therefore, improving the removal rate of nitrogen and phosphorus and increasing the treatment capacity are the key to the efficient operation of rural sewage treatment facilities.



289

Fig. 6. Sensitivity analysis for capital cost (a), operating cost (b), electricity consumption (c), treatment capacity (d), COD removed (e), $\text{NH}_3\text{-N}$ removed (f) and TP removed (g).

291

293

Table 3 Sensitivity analysis results and variable sensitivity rankings.

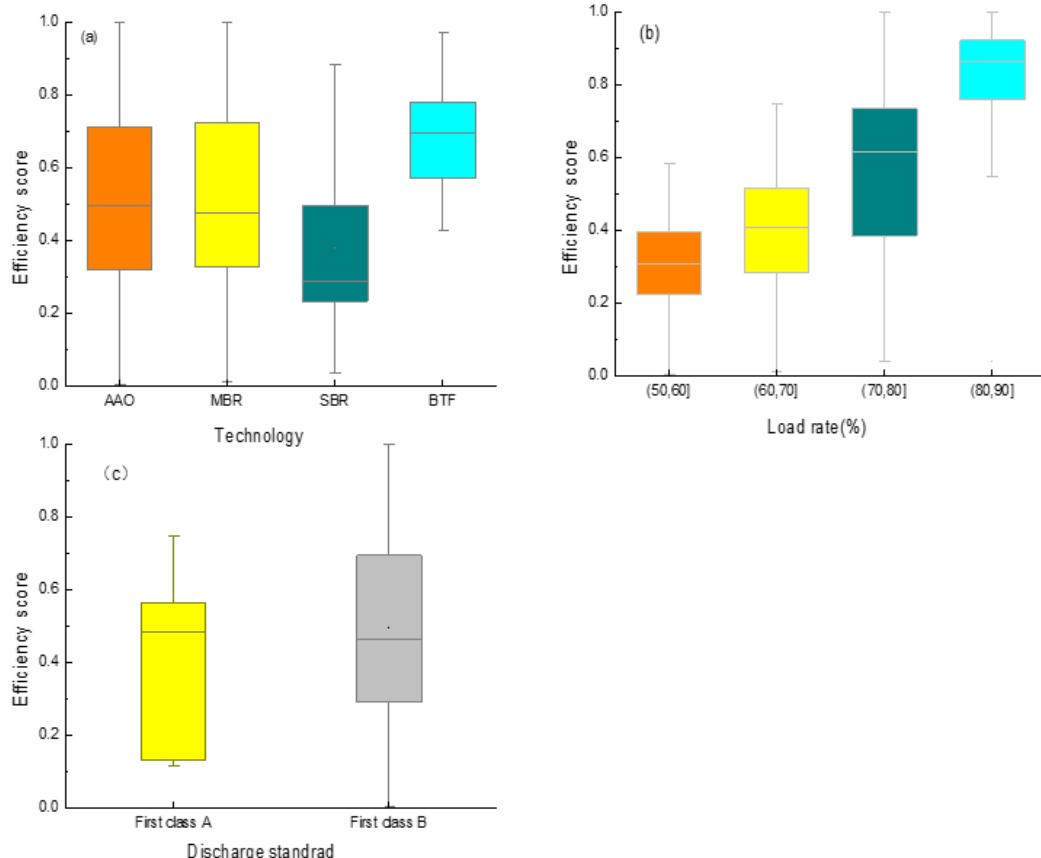
Ranking	Variables	Slope	$ 1\text{-slope} $	R^2	Classification
1	operating cost (10^4 CNY /year)	0.916	0.084	0.881	
2	capital cost (10^4 CNY)	0.946	0.054	0.975	Input
3	electricity consumption (10^4 kWh/year)	1.012	0.012	0.971	
4	$\text{NH}_3\text{-N}$ removed (t/year)	0.833	0.167	0.836	
5	TP removed (t/year)	0.842	0.158	0.857	Output
6	treatment capacity (10^4 t/year)	0.888	0.112	0.844	
7	COD removed (t/year)	0.962	0.039	0.844	

294

3.4 Implicit explanatory factors

295

DMUs were grouped according to three selected explanatory factors. The
296 characteristics of efficiency scores were shown in Fig. 7.



297

298

Fig. 7. Box charts of the explanatory factors.

299

Due to non-normal distribution of analyzed samples, the Kruskal–Wallis
300 non-parametric test, as the most suited way, had been taken to verify significant

301 differences among different groups in this study (Kruskal and Wallis, 1952; Sueyoshi
 302 and Aoki, 2001). The statistical significance (p) value is equal or less than 0.05
 303 meaning the explanatory factor significantly impact efficiency scores of samples.
 304 Otherwise, the explanatory factor has no significant impact on efficiency score of
 305 samples. Table 4 displayed detailed results.

306 **Table 4** Kruskal–Wallis test statistics for explanatory factors.

explanatory factors	total DMUs	mean	std.dev.	P-value	Chi-sq.
Technology				0	38.251
AAO	334	0.519	0.257		
MBR	222	0.515	0.259		
SBR	116	0.382	0.235		
BF	9	0.681	0.162		
Load rate				0	258.706
(50,60]	157	0.316	0.160		
(60,70]	218	0.404	0.178		
(70,80]	187	0.564	0.237		
(80,90]	119	0.799	0.216		
Discharge standard				0.589	0.293
First class A	6	0.422	0.250		
First class B	675	0.497	0.259		

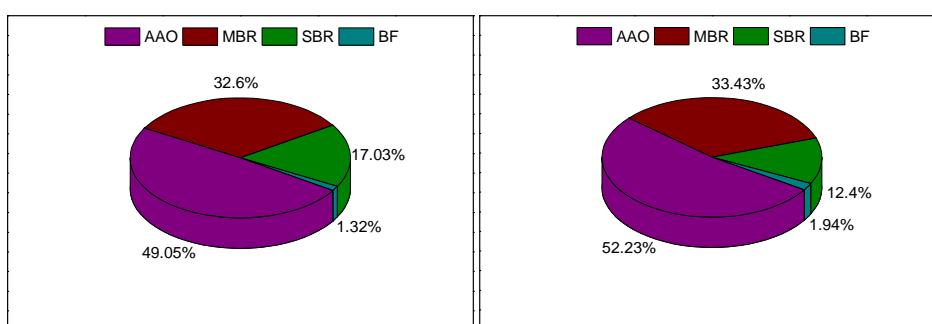
307 **3.4.1 Technology**

308 The sewage treatment technology means removing pollutants in wastewater
 309 through physical, chemical and biological processes, directly influencing the removal
 310 of pollutants. The process is generally divided into three levels: primary, secondary
 311 and tertiary treatment (Jin et al., 2014). Based on the classification of secondary
 312 treatment (Rodriguez-Garcia et al., 2011), the 681 samples were divided into four
 313 categories: (i) anaerobic-anoxic-oxic process (AAO), (ii) membrane bio-reactor
 314 process (MBR), (iii) sequencing batch reactor process (SBR) and (iv) Bio-trickling
 315 Filter (BTF).

316 According to the K-W test results (Table 4), the difference in performance of
 317 DMUs across the categories of technology was significant ($p < 0.05$). Hence,
 318 selecting efficient and economical technology can improve the performance of
 319 facilities. The boxplot for the four technologies efficiency scores were shown in Fig. 7

320 (a). The average score of AAO, MBR, SBR and BTF was 0.519, 0.515, 0.382 and
 321 0.681, respectively. SBR and MBR had lower scores mainly resulting from their low
 322 efficiency in removing contaminants. In addition, SBR and MBR required aerators to
 323 provide oxygen source, which increased operation costs and electricity consumption.
 324 This conclusion was consistent with previous views (Tolkou and Zouboulis, 2016).
 325 BTF had the highest score. When operating cost and energy consumption were similar,
 326 BTF had an advantage in pollutant removal, especially for the removal rate of COD
 327 (75%) and NH₃-N (94%). Therefore, BTF was suitable for underdeveloped rural areas
 328 effectively dealing with small-scale domestic sewage to improve rural water
 329 environment. This result agreed with the conclusion of Yang (2011).

330 The percentages of the number and total treatment capacity of facilities adopting
 331 different technologies in WUXI city were shown in Fig. 8. At present, 556 sewage
 332 treatment facilities had adopted AAO and MBR and the total treatment capacity was
 333 1.61×10^6 T/A. Only 9 facilities adopted the BTF, accounting for 1.94% of the total
 334 treatment capacity. Assuming that all facilities adopt BTF, when the treatment
 335 capacity and effect are the same, the average annual operating cost and power
 336 consumption of each facility will be reduced by 4,400 CNY and 49.54 kWh,
 337 respectively, and the capital cost will also be reduced by 6,400 CNY. Therefore, it is
 338 necessary for local government to upgrade of rural domestic sewage treatment
 339 facilities and to promote appropriate technology (BTF).



346 (a) Number of facilities (b) Total treatment capacity

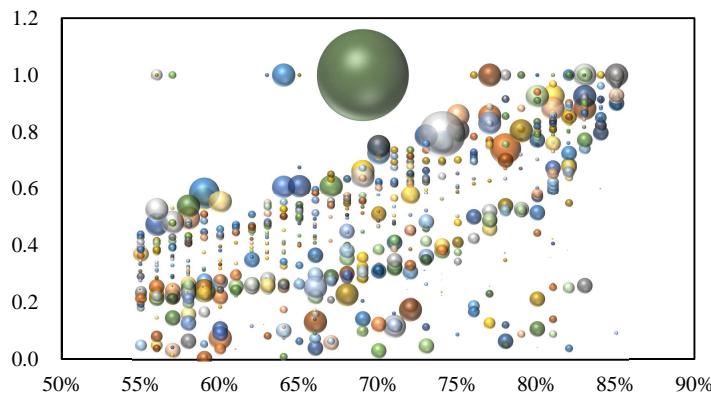
347 **Fig. 8.** Percentage of different treatment technologies.

348 3.4.2 Load rate

349 A common problem in treatment facilities of rural areas is that the design

350 treatment capacity is significantly higher than the actual treatment water volume,
351 resulting in the idleness of the facilities (Li and Xu, 2015; Yang et al., 2016). In other
352 words, the operating condition of facilities can be affected by not only the design
353 capacity but also actual capacity. Thus, this paper selected load rate as the second
354 implicit factor. DMUs had been divided into four groups based on load rate: (i)
355 50%-60%, (ii) 60%-70%, (iii) 70%-80% and (iv) 80%-90%. As shown in Table 4 and
356 Fig. 7 (b), the impact of the capacity load rate was significant ($p<0.05$). The average
357 efficiency scores of four groups was 0.316, 0.404, 0.564 and 0.799, respectively. The
358 performance efficiency of DMUs with a high load rate operate relatively better than
359 that of those with a low load rate. Our result was consistent with the finding of Hu et
360 al. (2019). As shown in Fig. 9, the load rates of facilities were all less than 100% also
361 confirmed that phenomenon of idle facilities mentioned above. Therefore, it is
362 essential to design the scale of treatment facilities reasonably to ensure the high load
363 operation of the facilities.

364 There were also a few DMUs that do not obey this rule: despite the relatively
365 larger scale and higher load rate, the scores of them were very unsatisfactory. This
366 phenomenon had also appeared in M.Molinos-Senante's study (2013). For example,
367 DMU 31 processed 5694 tons sewage in 2017, with a load rate of 78%, but efficiency
368 score of this plant was only 0.06. Studies showed that the component and
369 concentration of influent influence sewage treatment performance (Dong et al., 2017;
370 Hu et al., 2019). These abnormal inefficient DMUs had low concentration of pollutant
371 inflows, resulting in a poor removal of pollutants. Serious shortfalls of outputs were
372 considered to be the main explanation for the phenomenon. Besides, the relatively
373 higher energy consumption and operation costs also were reasons for low score.
374 Therefore, increasing the concentration of influent by a certain pretreatment while
375 taking the reduction of inputs and the increase of the capacity load rate into account
376 can be a good way to improve the performance of treatment facilities.



377

378 **Fig. 9.** Efficiency scores of DMUs in WUXI. Bubble size represents the actual capacity of the
 379 facilities, and every color represents one facility.

380 3.4.3 Discharge standard

381 The discharge standards directly affect the construction, operation and
 382 management of rural domestic sewage treatment facilities. According to “Discharge
 383 Standard of Pollutants for Municipal Wastewater Treatment Plant (GB18918-2002)”
 384 currently implemented in Wuxi rural areas, the samples were divided into two
 385 categories: (i) the first class A, (ii) the first class B. As shown in Fig. 7 (d), with the
 386 discharge standard more stringent, the efficiency score of samples became lower. The
 387 average score decreased from 0.497 to 0.422. At present, the effluent quality of 681
 388 rural sewage treatment facilities all met the first class B standard and 6 (0.89%)
 389 facilities met the first class A. Compared with the DMUs that met class B standard,
 390 the DMUs meeting the class A can increase the removal of COD, NH₃-N and TP by
 391 0.292, 0.150, and 0.012 tons equally each year, but the operating cost and electricity
 392 consumption will equally increase by 5000 CNY and 2587 kWh, respectively. The
 393 result of the K-W test showed that discharge standards had no significant effect on
 394 performance scores of DMUs. Therefore, upgrading the standard seems not an ideal
 395 measure to improve performance scores of rural sewage treatment facilities.
 396 Considering the effluent water quality was good, the tail water reuse should be the
 397 focus of the local government, which will not only improve the reuse rate of water
 398 resources, but also greatly reduce the cost of rural water environmental pollution
 399 treatment.

400 **4 Conclusion**

401 With the number and capacity of rural sewage treatment facilities increasing, a
 402 comprehensive, quantitative and objective evaluation of them is becoming urgent.
 403 DEA is considered to be an effective performance evaluation tool to solve this
 404 problem. In this paper, 681 rural sewage treatment facilities were evaluated by
 405 SBM-DEA model based on group-frontier from multiple dimensions including
 406 economy, environment and society. The main results are as follows: (1) the average
 407 efficiency score of samples was 0.496, of which only 27 facilities were operating
 408 effectively; (2) compared with efficient DMUs, the inefficient DMUs had significant
 409 shortfalls in the outputs, especially in treatment capacity and NH₃-N removal,
 410 respectively with the improvement potential of 92.45% and 91.97%; (3) the removal
 411 of nitrogen and phosphorus and treatment capacity are the sensitive factors to the
 412 efficiencies of rural sewage treatment facilities; (4) technology and capacity load rate
 413 had significant impacts on the performance of facilities.

414 Based on the results above, the targeted recommendations presented as follows
 415 to improve the performance of rural sewage treatment infrastructures in China: (1)
 416 upgrade and optimize treatment technologies: applying technologies which achieve
 417 the trade-off between pollutant removal and cost inputs, such as BTF process; (2)
 418 adjust operating conditions: increasing the operating load to avoid facilities idleness
 419 and increasing the concentration of influent by pretreatment; (3) encourage reuse of
 420 reclaimed water: reusing reclaimed water in various ways to achieve environment
 421 benefits and reduce the cost of rural water pollution treatment.

422 The SBM model selected in this paper identifies the efficient DMUs as the best
 423 practices, calculating slack improvement value of inputs and outputs to maximize the
 424 efficiencies of the inefficient facilities. It can help government and managers of water
 425 companies to evaluate the operation performance of a large number of sewage
 426 treatment facilities and realize the effective supervision and management of local
 427 facilities. On the other hand, this method obtains the relative efficiency of the
 428 evaluation object, its absolute environmental impact being unknown yet. Besides, this
 429 article has not given the quantitative suggestion of improving the performance score.

430 Thus, further research can combine DEA with quantitative analysis methods such as
431 life cycle assessment or cost-effectiveness analysis to evaluate efficiency of facilities
432 more accurately and provide quantitative improvement measures.

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439 **Appendix A. Supplementary data**

440 Analytical data related to this article can be found at online version and the initial
441 data that support the finding of this study are available from the corresponding author
442 upon request.

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- 568

Highlights

- Efficiency scores of 681 rural sewage treatment facilities were assessed by SBM-DEA model based on group-frontier.
- The improvement potential for samples was about 50.4%.
- 27 treatment facilities were regarded as best practices.
- Explicit factors affecting the performance of treatment facilities were discussed.
- Suggestions to improve efficiency of facilities in rural areas of China were proposed.

Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

