

Journal of Environmental Chemical Engineering

Effect of Fenton pre-oxidation on the physicochemical properties of Sludge-based biochar and its adsorption mechanisms for ammonia nitrogen removal --Manuscript Draft--

Manuscript Number:	JECE-D-23-04956R1
Article Type:	Research Paper
Keywords:	Fenton pre-oxidation; sludge-based char; ammonia nitrogen adsorption ; Surface composite modification; Structural enhancement
Corresponding Author:	Jun Cao Hohai University CHINA
First Author:	Jun Cao
Order of Authors:	Jun Cao Ruochen Wang Hualun Zhu Songsshan Cao Zhipeng Duan
Abstract:	Fenton pre-oxidation, as a promising technology for sludge deep dehydration, has been used intensively; however, the outputs (i.e., low-water-content sludge cakes) need to be further disposed for a full line sludge resource utilization. Carbonization presents a promising approach for disposing of the sludge cake within the process. Here, the physicochemical properties of sludge-based char obtained after Fenton pre-oxidation were studied. Both Fenton pre-oxidized sludge-based char (FSC) and raw sludge-based char (RSC) were produced within the temperature range of 400 °C to 600 °C, and the adsorption performance of the prepared chars, exemplified by their ability to adsorb ammonia nitrogen (NH₄⁺), was tested under various conditions. FSC exhibited superior char yield and adsorption performance compared to RSC, with chemical adsorption being the primary controlling factor. The optimum conditions for FSC were observed at 400 °C, resulting in a char yield of 72.2%, and an adsorption capacity of 221.7 mg/g. Notably, FSC demonstrated an increase in C-C, C-N, O=C-O, C-O and C=C functional groups in comparison to RSC, which suggests that pre-oxidation increases organic matter decomposition and chemical reactivity of the sludge. The introduction of Si-O-Si, Si-O-C and Fe-O-H functional groups in the FSC facilitated ion exchange with NH₄⁺ in an aqueous solution. The key findings indicate that Fenton pre-oxidation not only enhances the adsorption performance of FSC but also reduces the required carbonization temperature for excellent-char preparation. These results demonstrated the potential to integrate sludge deep dewatering, including pre-oxidation, with char preparation.

Response to reviewers

Effect of Fenton pre-oxidation on the physicochemical properties of Sludge-based biochar and its adsorption mechanisms for ammonia nitrogen removal
(Ref. JECE-D-23-04956)

Dear Editor Prof. Apostolos Giannis,

The authors gratefully acknowledge the valuable comments made by the reviewers. We have studied their comments carefully and have made correction which we hope meet with their approval. The changes in the revised manuscript are highlighted in blue according to the referee's comments.

Reviewer 1:

Issue 1:

In my opinion it is an interesting investigation and in general terms it is well done, but there are certain aspects that may be subject to discussion.

Page 15, end of section 3.1.2. From "Therefore, in order to analyze..." to "...fixed at 60 minutes." has already been explained above. This paragraph can be deleted.

Response 1:

Thank you for the pointing out the omission. According to the comment of the reviewer, we have deleted this part as following,

“Above all, the carbonization temperature plays a more important role in char formation and yield than residence time, since a residence time of 60 min is sufficient to achieve optimal char yield while minimizing energy consumption.” (Line 254 to 256)

Issue 2:

- The kinetic study I think has not been carried out properly. It is surprising that adsorption occurs by one mechanism and at the same time by a different mechanism. The case of an intermediate mechanism in which the value of R^2 would be far from 1 for both models is not proposed. The authors propose that both kinetic mechanisms are true, which is incongruent.

There is an explanation for this fact and it is a mistake that the authors have probably (as I can see from Fig 6B) made. There is one set of data that fits both models in any adsorption phenomenon studied: those where equilibrium has been reached. And most of the data presented in figure 6B are such data. But the study of kinetic models is interesting because it tells us how we get to equilibrium, not what happens at equilibrium. Theoretical models have to be applied to experimental data prior to equilibrium. I think you should repeat the experiment with shorter times.

- Continuing with this theme, in figures 6B and 6C there is hardly any difference

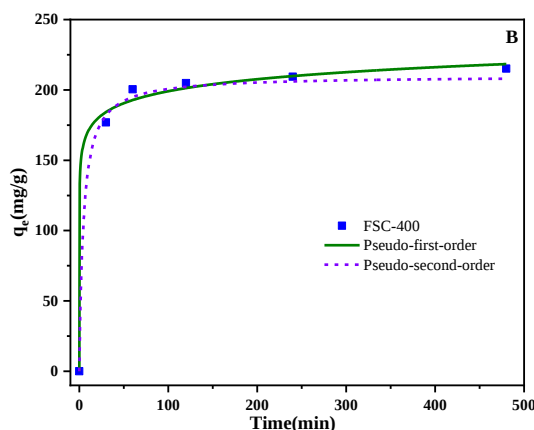
between the lines of the two models proposed in each figure. Although we live in a world that tends more and more towards visual communication, I believe that sometimes information is clearer if it is presented in the form of numerical tables. I therefore suggest that tables S2 and S3 be included in the main text.

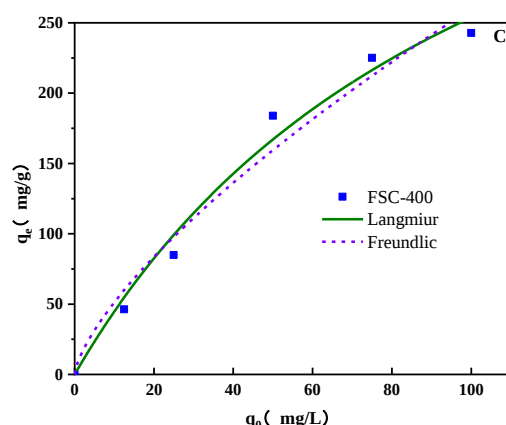
Response 2:

Thank you for your comment. Regarding issue 2 and 3, both of which are related to the kinetic study, we answered the questions together. We have carefully reviewed the reviewer's comments and made the necessary revisions to address the concerns raised. After thorough consideration and consulting relevant books and literatures, we acknowledge that the kinetics equations are generally derived under specific conditions within the equilibrium zone. According to the reviewer's suggestion, we conducted additional experiments with shorter times, and based on the obtained data, we have redrawn Figure 6B and 6C, and incorporated Table 4 and 5 in the main text. The results were more clearly presented. (Line 397 to 423)

“The results of the dynamics model’s equations and adsorption isotherms of FSC are presented in Fig. 6 B and 6C, as well as Table 4 and 5. The kinetic analysis of ammonia nitrogen adsorption on FSC indicates that it followed pseudo-second-order, as evidenced by the high values of R^2 (greater than 0.9964) in Fig. 6B.”

“The Langmuir model provided a better fit to the experimental data compared to the Freundlich model, as shown in Table 5, indicating that the adsorption process follows homogeneous monolayer adsorption.”





B. Ammonia nitrogen isotherms on FSC-400 and dynamics fitting models; C. Ammonia nitrogen isotherms on FSC-400 and adsorption fitting models

Table 4 Parameters of adsorption dynamic model

Model	K	q_e	R^2
pseude-first-order	1.23 ± 0.674	204.95 ± 6.72	0.9688
pseude-second-order	0.001 ± 1.745	209.44 ± 5.00	0.9964

Table 5 Parameters of adsorption isotherm model

Model	K	Q/n	R^2
Langmuir	$0.0082 \pm 8.7E-5$	56.35 ± 0.39	0.9959
Freundlich	10.09 ± 4.54	1.42 ± 0.21	0.9885

We believe that these modifications significantly improve the accuracy and presentation of the kinetic study, thereby strengthening the overall quality of the manuscript.

Issue 3:

- Page 28, lines 444 to 447. The authors propose that the carbon dioxide that can be generated is primarily responsible for changes in porosity. I believe this statement should be revised for the following reasons:

- Activation with carbon dioxide is usually carried out at higher temperatures than those used in this research, as it is a molecule with limited reactivity.
- During carbonization, it is possible that part of the solid structure may be destroyed by temperature. This includes pore walls and usually results in a wider porosity.
- N_2 adsorption at low pressures is higher in the RSC series, so presumably the treatment decreases the number and volume of micropores contrary to the authors' claims.

Response 3:

We thank the reviewer for raising valuable points regarding the proposed role of

carbon dioxide in the changes in porosity. Upon careful consideration and consulting relevant literature, we acknowledge that carbon dioxide is more likely to be generated at a slight level at 400 °C, as the activation with carbon dioxide typically occurs at higher temperatures due to its limited reactivity. Based on the suggestion from the reviewer, we have provided the following explanation: The Fenton pre-oxidation process led to the degradation of the biopolymers and the conversion of sludge flocs into small fragments [26]. This transformation has significant benefits for pore development during carbonization, as it enhanced decarboxylation reactions and caused partial destruction of the solid structure, resulting in a wider porosity. Additionally, Fenton oxidation promoted the transition from micropores to mesopores (Fig.5). (Line 453 to 459)

By considering these factors, we agree that the role of carbon dioxide in the observed changes in porosity should be revised, and we appreciate the reviewer's valuable insights that have helped improve the accuracy and clarity of our explanation.

Reviewer 2:

Issue 1:

The novelty or specific contribution to knowledge should be clearly emphasized in abstract as well as the text.

Accordingly, the authors should expand further on the literature review. If this is not the first time this technology has been used, what exactly have the authors done differently?

Response 1:

According to the suggestion of reviewer, we have revised the Abstract and the relevant part in the text to emphasize the novelty and specific contribution of our study.

In the Abstract, we now clearly state that our study demonstrates the novelty contribution of using Fenton pre-oxidation as a component for the carbonization process. “The key findings indicate that Fenton pre-oxidation not only enhances the adsorption performance of FSC but also reduces the required carbonization temperature for excellent-char preparation. These results demonstrated the potential to integrate sludge deep dewatering, including pre-oxidation, with char preparation.” (Line 39 to 43)

Furthermore, as per the reviewer's comment, we have enhanced the literature review section (Introduction) by incorporating relevant references to previous studies: Lin et al. and Li et al. found that Fenton conditioning greatly enhanced the adsorption capacity by affecting the porosity of the sludge [26, 27]. Belete et al. reported that the Fenton reagents increased the acidic group's concentration and the BET surface area, but without changing the morphology [28]. However, our study contributes to the existing body of knowledge by exploring the relatively limited discussions on the

impact of Fenton pre-oxidation on the physicochemical properties of char surfaces and its subsequent effects on the adsorption mechanism. (Line 88 to 95)

We also have enhanced the conclusion section: These findings highlight the effectiveness of Fenton pre-oxidation on sludge-based char, and present it as a promising technology by integrating the dehydration and carbonization process with lower energy consumption. (Line 511 to 514)

Issue 2:

How were the optimum conditions stated in the abstract and elsewhere in the body arrived at? One could not find any specific objective and/or the mathematical model used for this? Or, on what basis?

Response 2:

Thank you for your inquiry. The determination of the optimum conditions, specifically the carbonization temperature of 400°C, was based on the following considerations:

1. Highest adsorption capacity: Experimental results indicated that the adsorption capacity of the Fenton sludge-based biochar (FSC) was the highest at a carbonization temperature of 400°C, reaching 221.7 mg/g. This value was compared to higher temperatures, where the adsorption capacity was lower. The significant increase in adsorption capacity at 400°C can be attributed to the enhanced performance of FSC due to Fenton pre-oxidation.
2. Char yield optimization: Additionally, the char yield achieved at 400°C was the highest at 72.2%. This means that a substantial amount of biochar is obtained at this temperature, which translates to lower energy consumption and higher returns. The higher char yield at 400°C further supports the selection of this temperature as the optimum condition.
3. Integration of dehydration and carbonization: Another essential aspect considered is the integration of the dehydration and carbonization process. Fenton pre-oxidation not only enhances the adsorption performance of FSC but also reduces the required carbonization temperature for the preparation of excellent-char. This integration results in lower overall energy consumption, making the technology more promising and efficient.

These points collectively supported the selection of 400°C as the optimum carbonization temperature. It ensured the highest adsorption capacity, optimal char yield, and energy-efficient integration of the dehydration and carbonization processes. We believe that these factors contribute to the significance and validity of the chosen conditions.

Issue 3:

Was the FS prepared in this research?

Response 3:

Yes, in this research, the FS was indeed prepared. We followed the specific

preparation procedures and used the optimal condition that was obtained in our previous study. We have revised the section 2.1 Materials and provided additional information:

“The RS was obtained from the secondary sedimentation tank, with a moisture content of 82.5%. The FS was prepared following the procedures and the optimum conditions obtained in our previous study [7]. The mole ratio H_2O_2 : Fe^{2+} was 2. The content of catalyst $\text{FeSO}_4 \cdot 7\text{H}_2\text{O}$ was 6% (m/m), and the concentration of oxidants H_2O_2 was 30%. Under these specified conditions, the moisture content of FS was successfully reduced to 53.2%.” (Line 127 to 133)

Issue 4:

Line 26/27: "pseude-first-order (3) and pseude-second-order r (4) model" What are these "pseude" orders? Did the authors mean "pseudo"?

Response 4:

We apologize for the oversight. Based on the reviewer's comment, we have revised the word “pseude” to “pseudo”. We appreciate the reviewer's feedback and attention to detail. (Line 199)

Reviewer 3:

Issue 1:

I have reviewed the paper entitled 'Effect of Fenton pre-oxidation on the physicochemical properties of Sludge-based biochar and its adsorption mechanisms for ammonia nitrogen removal'. The paper presents relevant information. It is well designed and easy to follow. The applied methodology is adequate, and the literature cited is actualized. It deserves publication in Journal of Environmental Chemical Engineering after Major Revisions. The authors must consider the following comments:

- What is the significance of this study? The novelty of this work should be further highlighted. The authors should clarify the principal motivation behind this manuscript.

Response 1:

Thank you for the valuable comments, and thank you for recognizing the innovation and originality of this article. According to your suggestion, we have made the necessary revisions to further highlight the significance and novelty of this study.

In the Abstract, we now clearly state that our study demonstrates the novelty contribution of using Fenton pre-oxidation as a component for the carbonization process. “The key findings indicate that Fenton pre-oxidation not only enhances the adsorption performance of FSC but also reduces the required carbonization temperature for excellent-char preparation. These results demonstrated the potential to

integrate sludge deep dewatering, including pre-oxidation, with char preparation.” (Line 39 to 43)

Furthermore, as per the reviewer's comment, we have enhanced the literature review section (Introduction) by incorporating relevant references to previous studies: Lin et al. and Li et al. found that Fenton conditioning greatly enhanced the adsorption capacity by affecting the porosity of the sludge [26, 27]. Belete et al. reported that the Fenton reagents increased the acidic group's concentration and the BET surface area, but without changing the morphology [28]. However, our study contributes to the existing body of knowledge by exploring the relatively limited discussions on the impact of Fenton pre-oxidation on the physicochemical properties of char surfaces and its subsequent effects on the adsorption mechanism. (Line 88 to 95)

We also have enhanced the conclusion section: These findings highlight the effectiveness of Fenton pre-oxidation on sludge-based char, and present it as a promising technology by integrating the dehydration and carbonization process with lower energy consumption. (Line 511 to 514)

Issue 2:

For keywords, please revise Line 46 and include more innovative keywords.

Response 2:

As per the reviewer's comment, we have revised the keywords that with more innovative words.

“Fenton pre-oxidation; sludge-based char; ammonia nitrogen adsorption; Surface composite modification; Structural enhancement” (Line 45 to 46)

Issue 3:

For Line 160, why the author used the word 'quality' in this sentence. Better to be removed and replaced only by 'mass'.

Response 3:

Thank you for pointing out the omission. According to the comment of the reviewer, we have revised the word “quality” to “mass”. (Line 166 to 167)

Issue 4:

Avoid repeating same words many times through the manuscript. Define the abbreviation once then just use the abbreviations.

Response 4:

Based on the reviewer's comment, we have made the necessary revision to avoid words repeated many times and only use the abbreviations after defined. We appreciate the reviewer's attention to detail and their valuable feedback.

Issue 5:

For supplementary materials, the title page with authors and affiliations should be

included.

Response 5:

Thank you for pointing out the omission. According to the comment of the reviewer, we have revised the title page, and now the authors and affiliations are included in the supplementary materials.

Issue 6:

The authors should appropriately insert related references especially at the methods section. Please revise accordingly.

Response 6:

As per the reviewer's comment, we have added the necessary references to the manuscript (Section 2.4)

“For the adsorption kinetic study, a NH_4^+ solution with 50 mg/L initial concentration was applied. The behavior of the adsorption with the function of time was fitted using the pseudo-first-order (3) and pseudo-second-order (4) model [28],”

“The data were fitted to Langmuir and Freundlich model [29], the equations were as follows” (Line 199 to 200, Line 206)

Issue 7:

A cost-estimation analysis should be included in this manuscript to show the potential application of this work.

Response 7:

Thank you for the suggestion. We agree that a comprehensive economic analysis is essential to demonstrate the potential application and feasibility of the proposed technology. In our previous work, we have studied the promotion of Fenton pre-oxidation on secondary advanced dewatering of sludge on pilot scale. And we have a cost-estimation analysis of the pilot test. The total cost considered reagent cost and electricity bill primarily, resulting in a range of 98.5-113.7 CNY/t · wb [1]. In this work, we intended to prove that Fenton pre-oxidation is not only a promising technology for sludge deep dehydration, but also can synchronously improve the structure and surface properties of sludge-base char without additional consumption of materials and energy on lab scale. The results showed that pre-oxidation not only reduces the energy consumption associated with deep dehydration, but also lowers the energy requirement of carbonization, thereby significantly reducing overall energy consumption. However, we acknowledge that a complete economic analysis has not been conducted in this particular study since we primarily worked on a fixed bed at the laboratory scale. To address this, we appreciate your suggestion and plan to conduct experiments on a pilot scale to thoroughly analyze the cost estimation of integrating the dehydration and carbonization processes. By doing so, we can provide a more comprehensive understanding of the economic feasibility and practical application of this technology in larger-scale scenarios. We sincerely thank you for your insightful feedback, and we are committed to enhancing the overall quality and

impact of our manuscript.

[1] N. Lin, W. Zhu, X. Fan, C. Wang, C. Chen, H. Zhang, L. Chen, S. Wu, Y. Cui, Key Factor On Improving Secondary Advanced Dewatering Performance of Municipal Dewatered Sludge: Selective Oxidative Decomposition of Polysaccharides, *Chemosphere*, 249 (2020) 126108. <https://doi.org/https://doi.org/10.1016/j.chemosphere.2020.126108>

Issue 8:

Some statistics are needed for this work such as including a principal component analysis considering the impact of the different parameters on the potential ammonia adsorption using chars. This would validate the obtained results.

Response 8:

Thank you for your suggestion regarding the inclusion of a principal component analysis to validate the obtained results. We agree that statistical analysis, such as principal component analysis, can provide valuable insights into the impact of different parameters on the potential ammonia adsorption using chars. In this study, we have indeed conducted some experiments on the potential ammonia with different parameters, such as different carbonization temperatures and pH of aqueous solutions. The results were shown is the following figures. The main focus of this paper is on the effect of Fenton pre-oxidation on the performance of sludge-based biochar, so the data are mostly presented around this point. To address this, we appreciate your feedback and will ensure that the principal component analysis and further statistical analysis are included in our future work. By conducting a more in-depth analysis, we aim to reveal the complex improving mechanism proposed in this paper and gain a better understanding of the relationship between process parameters and char adsorption capacity.

We thank you for your valuable input, which has contributed to the refinement of our research approach. Your feedback is highly appreciated, and we look forward to incorporating the principal component analysis and conducting more comprehensive studies in the future

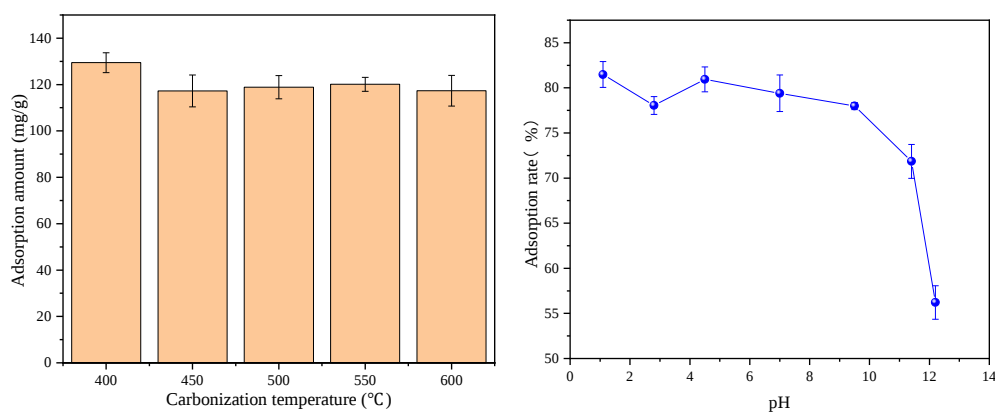
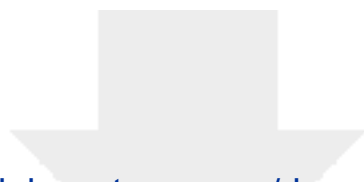


Fig.1 Effect of carbonization temperature on adsorption characteristics Fig. 2 The effect of pH on adsorption characteristic

Highlights:

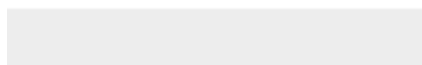
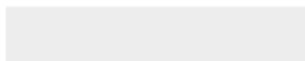
1. Fenton pre-oxidation reduced the required carbonization temperature for FSC
2. Pre-oxidation modified the structure of FSC leading to greater adsorption capacity
3. The best adsorption capacity (221.7 mg/g) of FSC was achieved at 400°C
4. The study provides the possibility of integrating dehydration and carbonization



[Click here to access/download](#)

Supplementary Material

Revised Supplementary documents.docx



CRedit authorship contribution statement

Jun Cao: Conceptualization, Investigation, Data curation, original draft. **Ruochen**

Wang: Investigation, Data curation. **Hualun Zhu:** Review & Editing. **Songshan Cao:**

Methodology, Review & Editing. **Zhipeng Duan:** Review & Editing.

Declaration of interests

☒The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

☐The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

**Effect of Fenton pre-oxidation on the physicochemical properties
of Sludge-based biochar and its adsorption mechanisms for
ammonia nitrogen removal**

Jun Cao^{a,b,*}, Ruochen Wang^b, Hualun Zhu^c, Songshan Cao^{c,d}, Zhipeng Duan^e

^aNational Engineering Research Center of Water Resources Efficient Utilization and
Engineering Safety, Hohai University, Nanjing 211111, China

^bCenter for Taihu Basin, Institute of Water Science and Technology, Hohai University,
Nanjing 211111, China

^cDepartment of Chemical Engineering, University College London, London WC1E 7JE,
United Kingdom

^dKey Laboratory of Energy Thermal Conversion and Control of Ministry of Education, School
of Energy and Environment, Southeast University, Nanjing, 210096, China.

^e College of Environment, Hohai University, Nanjing 211111, China

* Corresponding author: Jun Cao

National Engineering Research Center of Water Resources Efficient Utilization and
Engineering Safety, Hohai University, Nanjing 211111, China. Tel.: +86-15850598642;

E-mail address: caojunxm@126.com.

21 **Abstract:**

22 Fenton pre-oxidation, as a promising technology for sludge deep dehydration,
23 has been used intensively; however, the outputs (i.e., low-water-content sludge
24 cakes) need to be further disposed for a full line sludge resource utilization.
25 Carbonization presents a promising approach for disposing of the sludge cake
26 within the process. Here, the physicochemical properties of sludge-based char
27 obtained after Fenton pre-oxidation were studied. Both Fenton pre-oxidized
28 sludge-based char (FSC) and raw sludge-based char (RSC) were produced within
29 the temperature range of 400 °C to 600 °C, and the adsorption performance of the
30 prepared chars, exemplified by their ability to adsorb ammonia nitrogen (NH_4^+),
31 was tested under various conditions. FSC exhibited superior char yield and
32 adsorption performance compared to RSC, with chemical adsorption being the
33 primary controlling factor. The optimum conditions for FSC were observed at
34 400 °C, resulting in a char yield of 72.2%, and an adsorption capacity of 221.7
35 mg/g. Notably, FSC demonstrated an increase in C-C, C-N, O=C-O, C-O and
36 C=C functional groups in comparison to RSC, which suggests that pre-oxidation
37 increases organic matter decomposition and chemical reactivity of the sludge. The
38 introduction of Si-O-Si, Si-O-C and Fe-O-H functional groups in the FSC
39 facilitated ion exchange with NH_4^+ in an aqueous solution. The key findings

40 indicate that Fenton pre-oxidation not only enhances the adsorption performance
41 of FSC but also reduces the required carbonization temperature for excellent-char
42 preparation. These results demonstrated the potential to integrate sludge deep
43 dewatering, including pre-oxidation, with char preparation.

44 **Keywords:** Fenton pre-oxidation; sludge-based char; ammonia nitrogen
45 adsorption; Surface composite modification; Structural enhancement

46

47 **1 Introduction**

48 The challenge of achieving comprehensive treatment of sludge is becoming
49 increasingly prevalent worldwide. With the growing collection and treatment of
50 domestic sewage in China, the country's sewage treatment capacity has
51 significantly increased [1]. Consequently, the disposal of sewage sludge has also
52 experienced a rapid rise, reaching approximately 60 million tons in 2020 (80%
53 water content). There is an urgent need to develop a technology that allows for
54 sludge recycling while minimizing its volume and associated environmental
55 impact. However, the high moisture content of sludge, which can reach up to
56 99.5wt% [2], currently poses a major obstacle disposal method due to considerable
57 energy consumption and operational costs associated with the sequent thermal
58 conversion technologies required for further treatment of sludge. The Fenton or

59 Fenton-like pre-oxidation method is widely recognized as a crucial step in the
60 pretreatment process for achieving deep water removal [3]. This method is
61 performed prior to subsequent thermal disposal and disintegration [4, 5]. In our
62 previous study, we modified this technology and reduced the water content from
63 82.5% to 53.2% under optimal conditions (mole ratio H_2O_2 : Fe^{2+} =2) [6,8]. This
64 modification effectively reduced the volume and elevated the low calorific value
65 of the sludge. However, it resulted in the production of a significant amount of
66 Fe-rich sludge cake, which presents challenges for proper disposal [9]. The
67 obtained sludge cake typically retains a significant amount of organic matter,
68 making it a key research focus to develop sustainable treatment methods in line
69 with the goals of “emission peak” and “carbon neutralization”.

70 Sludge carbonization is widely regarded as a promising method to accomplish
71 this objective [10,12]. Carbonization can greatly reduce the weight and volume by
72 decomposing and evaporating water, while simultaneously facilitating energy
73 conservation and the acquisition of byproducts. Sludge-based char, due to its large
74 specific surface area, well-developed microstructure, and abundant oxygen
75 functional groups and aromatic compounds on the surface, exhibits characteristics
76 of both an energy material and an adsorbent. Currently, the char finds significant
77 applications in various fields, including fuel energy, environmental protection,

78 materials and agriculture [3, 13,16]. An important utilization of char involves its
79 ability to adsorb and immobilize heavy metal ions [17], as well as nitrogen,
80 phosphorus and organics in water purification, hereby acting as a cost-effective
81 alternative to commercial adsorbents [18]. The potential for widespread
82 application and expansion of the means depends on whether the char prepared
83 from pre-oxidized sludge has a high utilization value. Studies have demonstrated
84 that the Fenton and Fenton-like char synthesized by carbonization exhibits
85 superior performance compared to that of RS, making them efficient adsorbents
86 for pollutant removal [19]. Until now, studies have investigated the effects of
87 temperature, residence time, preparation materials and additives on the structural
88 characteristics of chars [20,25]. Lin et al. and Li et al. found that Fenton
89 conditioning greatly enhanced the adsorption capacity by affecting the porosity of
90 the sludge [26, 27]. Belete et al. reported that the Fenton reagents increased the
91 acidic group's concentration and the BET surface area, but without changing the
92 morphology [28]. However, our study contributes to the existing body of
93 knowledge by exploring the relatively limited discussions on the impact of Fenton
94 pre-oxidation on the physicochemical properties of char surfaces and its
95 subsequent effects on the adsorption mechanism.

96 In aquatic environment, excessive emission of ammonia nitrogen is a

97 significant contributor to eutrophication, posing a threat to water resources
98 worldwide [29]. This issue has garnered global attention and has become a hot
99 topic of concern. Consequently, the discharge standards of ammonia nitrogen have
100 become increasingly stringent both domestically and internationally, and the
101 efficient removal of ammonia nitrogen from water has become a crucial objective
102 in safeguarding aquatic environment. The biochar derived from rice straw, wheat
103 straw, coconut shell, and other materials has been identified as a simple and
104 low-cost adsorbent for ammonia nitrogen removal from water [30]. This is due to
105 its large specific surface area, strong ion exchange ability, and the presence of a
106 negative charge on the surface. Previous studies have shown that there are more
107 oxygenated functional groups formed, and the newly Fe (III) and Fe (II) groups
108 introduced on the Fenton and Fenton-like char. These groups may enhance the ion
109 exchange and provide an increased number of negative charges during the
110 adsorption process [9, 19]. Considering these findings, the removal of ammonia
111 nitrogen using Fenton pre-oxidized sludge-based char has been identified as a
112 potential application direction due to the favorable characteristics of the char for
113 this purpose.

114 Above all, in order to evaluate the mechanism of pre-oxidation on the
115 adsorption properties of sludge-based char, both Fenton pre-oxidized sludge (FS)

116 and the raw sludge (RS) were used to produce chars under different carbonization
117 temperatures and residence times. The influences of Fenton pre-oxidation on char
118 yield, element composition, specific surface area, microstructure, and functional
119 groups were analyzed. Furthermore, the adsorption performance of ammonia
120 nitrogen on the char and the corresponding adsorption mechanism were also
121 studied.

122 **2 Materials and methods**

123 **2.1 Materials**

124 Municipal Sludge samples were collected from Xushe Wastewater Treatment
125 Plant located in Yixing city, China. The plant has a sewage treatment capacity of
126 10,000 tons per day. The plant employs the A²O+filter processing system for
127 municipal wastewater treatment. The RS was obtained from the secondary
128 sedimentation tank, with a moisture content of 82.5%. The FS was prepared
129 following the procedures and the optimum conditions obtained in our previous
130 study [7]. The mole ratio H₂O₂: Fe²⁺ was 2. The content of catalyst FeSO₄·7H₂O
131 was 6% (m/m), and the concentration of oxidants H₂O₂ was 30%. Under these
132 specified conditions, the moisture content of FS was successfully reduced to
133 53.2%. The higher heating values (HHVs) of sludge rose with the addition of
134 oxidants, increased from 0.06 to 7.97 MJ/kg. FeSO₄·7H₂O and H₂O₂ (30%) were

both purchased from Sinopharm Chemical Reagent Co., Ltd. The samples were then crushed and screened without pre-drying before all experiments. The basic property parameters of sample RS and FS were listed in Table 1.

Table 1 Basic property parameters of sludge

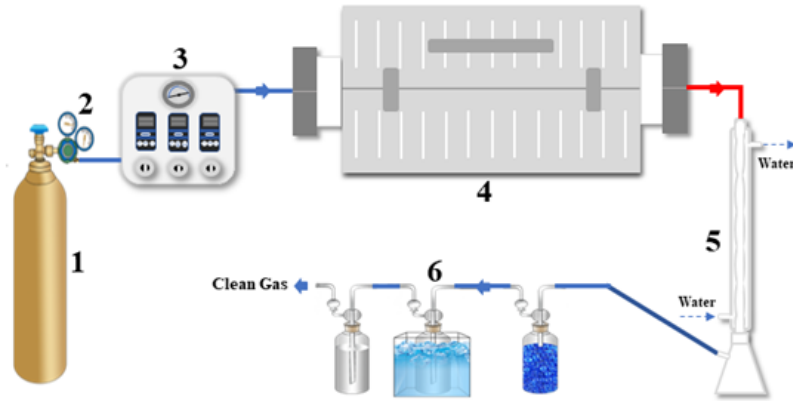
Sample	Proximate Analysis (wt%, daf)			Elemental Analysis (%)					pH	Q_{ar}
	A_{ar}	V_{ar}	FC_{ar}	C	H	O	N	S	-	$MJ \cdot kg^{-1}$
RS	45.81	50.27	6.92	25.35	4.89	16.99	2.86	0.99	6.78	0.06
FS	42.34	51.35	3.31	23.56	4.39	18.44	3.45	1.84	4.95	7.97

2.2 Materials preparation

The sludge-based char was prepared using a system illustrated in Fig.1. An N_2 flow, which controlled by a mass flow controller, was used to maintain an inert atmosphere. The fixed bed furnace employed a quartz tube with a length of 800 mm and an inside diameter of 40 mm, heated externally by a resistance furnace controlled by a temperature programmed controller. One thermocouple was placed inside the tube to monitor the real-time temperature of the reaction zone, while the other thermocouple was set outside the tube to monitor the temperature of the reactor. The gas produced during the experiment was purified by passing it through a contented condensing device and a tail gas treatment system.

The preparation of the char proceeded as follows: First, the sludge was placed in the middle of the tube. To remove all oxygen out of the tube, N_2 was pumped-in

152 with a flow rate of 200 ml/min for 30 min after leak checking. Then the reactor
 153 was heated at a rate of 10 °C/min, starting from room temperature to the target
 154 temperatures (400, 450, 500, 550 and 600 °C) and maintained for a certain time
 155 (30, 60, 90 and 120 min). The reactor was turned off and allowed to cool down
 156 until the furnace temperature dropped below 150 °C. Subsequently, the samples
 157 were removed from the reactor, weighted and collected. The resulting products
 158 were denoted as RSC-T and FSC-T, where T represents the target temperature. The
 159 sludge-based char yield was calculated according to the Formula 1:



160
 161 1. N₂ gas cylinder; 2. Pressure reduce regulator; 3. Mass flowmeter and controller; 4.
 162 Fixed bed with electric furnace; 5. Condensing device; 6. Tail gas treatment
 163 Fig. 1 Sludge char preparation experimental schematic diagram
 164

$$\eta = \frac{m_0 - m_1}{m_0} \times 100\% \quad (1)$$

166 where η is char yield, %; m_0 is original sludge mass, g; m_1 is mass of
 167 obtained sludge-based char, g.

168 **2.3 Characterization and Analysis**

169 An element analyzer (EA, EURO EA3000, Italy) was applied to determine the
170 contents of C, H, O, N and S in sludge and char obtained at different temperatures.
171 An acidimeter (pH, NY/T 1377.2007, China) was used to analyze the pH level and
172 conductivity of the samples. Fourier infrared analysis (FTIR, Model Vextor 22,
173 Germany) was conducted from 4000 to 400 cm^{-1} and each spectrum was a spectral
174 average of at least four scans. X-ray photoelectron microscopy (XPS, Thermo
175 Scientific ESCALAB250Xi, USA) analysis was performed and the binding energy
176 was corrected by contaminated char (C1s, 284.6eV). The CasaXPS program was
177 used to fit and divide peaks. The N_2 isotherm adsorption profile was obtained
178 using an automated gas sorption analyzer (ASAP2020, Micomeritics, USA) at
179 77K. The specific surface areas (S_{BET}) of samples were calculated according to the
180 Brunauer, Emmett, Teller method (BET) from the adsorption isotherms. The pore
181 size (P_A) distribution of samples before and after calcination was calculated from
182 desorption branch using the Barrett, Joyner and Halenda (BJH) method. Scanning
183 electron microscope (SEM, S4800, Japan) was used to analyze the surface
184 morphology of chars.

185 **2.4 Ammonia nitrogen adsorption**

186 The adsorption experiments were conducted using a set of 50 mL centrifuge

187 tube containing 0.1 g of sludge char and 50 mL of NH_4^+ with an initial
 188 concentration of 50 mg/L in an aqueous solution. The tubes were placed on a
 189 shaker at 20 °C and 120 r/min. The samples were then filtered and the residual
 190 concentration of NH_4^+ was analyzed using 420 nm of Nessler reagent
 191 spectrophotometry. The adsorption capacity of NH_4^+ was calculated as follows:

$$q_e = \frac{(C_0 - C_e) \times V}{m_1} \quad (2)$$

193 Where, q_e represents the adsorption amount per unit mass of sludge char at
 194 adsorption equilibrium, mg/g. C_0 and C_e represent the initial and equilibrium
 195 concentrations of NH_4^+ in solution respectively, mg/L. V represents the volume of
 196 solution, mL.

197 For the adsorption kinetic study, a NH_4^+ solution with 50 mg/L initial
 198 concentration was applied. The behavior of the adsorption with the function of
 199 time was fitted using the pseudo-first-order (3) and pseudo-second-order (4) model
 200 [31], as followed:

$$q_t = q_e (1 - e^{-K_1 t}) \quad (3)$$

$$\frac{t}{q_t} = \frac{1}{K_2 q_e^2} + \frac{1}{q_e} t \quad (4)$$

203 Where, t represents reaction time, min. q_t represents the sorption capacity at
 204 t time, mg/g. K_1 and K_2 represent rate constants of the two models, respectively,

205 g/(mg·min).

206 The data were fitted to Langmuir and Freundlich model [32], the equations
207 were as follows:

$$208 \quad q_e = \frac{q_m K_L C_e}{1 + K_L C_e} \quad (5)$$

209 Linearized and obtained,

$$210 \quad \frac{C_e}{q_e} = \frac{1}{K_L q_m} + \frac{C_e}{q_m} \quad (6)$$

$$211 \quad q_e = K_F C_e^{\frac{1}{n}} \quad (7)$$

212 Linearized and obtained,

$$213 \quad \ln q_e = \ln K_F + \frac{1}{n} \ln C_e \quad (8)$$

214 Where q_m represents saturated adsorption capacity, mg/g. K_L represents
215 Langmuir adsorption constant. K_F represents the Langmuir adsorption constant. C_e
216 represents the equilibrium mass concentration of ammonia nitrogen, mg/L. n
217 represents the empirical constant.

218 **3 Results and discussion**

219 **3.1 Carbonization**

220 **3.1.1 Effect of carbonization temperature**

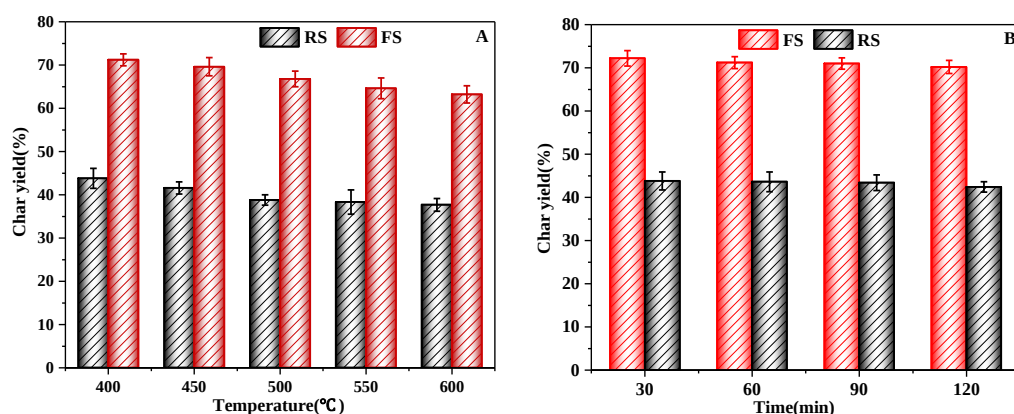
221 Fig. 2A showed the char production rate of RS and FS when the
222 carbonization temperature rose from 400 to 600 °C, the sample mass was 5 g and

the residence time was 60 min. It can be observed that the char yield of FSC was higher than that of RSC at each temperature point, with the char production rate of FSC ranging from 62.4% to 72.2%, while that of RSC was 36.67% to 43.8%. Both FS and RS achieved the maximum char yield at 400 °C, which were 72.2% and 43.8% respectively. With the increase of carbonization temperature, the yield of both FSC and RSC decreased. However, as the temperature further increased up to 500°C, the downward trend of sludge char yield became smaller and eventually stabilized at 550 °C. This can be attributed to the decomposition of volatile components in the sludge, which is mostly completed by 500 °C. At this point, further increase in temperature has a lesser effect on the char yield as the remaining components are more stable. The difference in char yield between FS and RS can be attributed to their disparate moisture content [33]. The RS had a moisture content of 82.5%, while the FS had a reduced moisture content of 53.2%. The lower moisture content in FS might lead to higher char yield, as moisture is vaporized during the carbonization process, resulting in a greater mass loss in RS compared to FS.

3.1.2 Effect of residence time

In the study conducted at 400 °C, the effect of residence time on energy consumption and char yield was investigated. The experiments were performed

242 with a sample weight of 5 g and residence times of 30, 60, 90, and 120 min. Fig.
 243 2B shows that as the residence time extended from 30 min to 120 min, the
 244 decrease in char yield was 2% for FS and 1.4% for RS, respectively. This indicates
 245 that the decomposition of organic matter in the sludge had essentially completed
 246 within a residence time of 60 min. Further extending the residence time did not
 247 significantly affect the char yield, suggesting that the carbonization process had
 248 reached a relatively stable state in terms of char formation.



249 Fig. 2 A. Effect of carbonization temperature on char yield: The mass of sample was 5 g, the
 250 residence time was 60 min, the temperature was 400, 450, 500, 550, 600°C. B. Effect of
 251 residence time on char yield: The mass of sample was 5g, the temperature was 400°C and the
 252 residence time was 30, 60, 90 and 120 min.
 253

254 Above all, the carbonization temperature plays a more important role in char
 255 formation and yield than residence time, since a residence time of 60 min is
 256 sufficient to achieve optimal char yield while minimizing energy consumption.

257 3.2 Effect of Fenton pre-oxidation

258 3.2.1 Effect of Fenton pre-oxidation on elemental composition

259 The element composition and ash content of the char derived from different
260 carbonization temperatures, ranging from normal atmospheric temperature to
261 various carbonization temperatures, were carried out. The results were shown in
262 Table 1 and Table 2. According to Table 1, the contents of ash, volatiles and S of
263 cakes pretreated with Fenton agent were higher compared to the RS, which can be
264 attributed to the addition of inorganic oxidants. In Table 2, it can be observed that
265 the carbonization temperature increased, the contents of C, H, O and N, as well as
266 the ratios of H/C, decreased for both FSC and RSC. This reduction indicated that
267 the organic matter in the sludge was more thoroughly decomposed and released in
268 the form of H_2O , CO_2 , CO , CH_4 , H_2 and NO_x with increasing temperature [4].
269 Thus, the content of carboxyl, hydroxyl, protein and other functional groups on the
270 surface of char progressively decreased, indicating an increased degree of
271 carbonization. As an important parameter to characterize the carbonization level,
272 the decrease of H/C ratio further confirmed the enhancement of the aromaticity
273 and stability of the char. The S content in the char showed slight growth regardless
274 of the temperature, due to the addition of oxidants [34]. The increase of S in the
275 char can be attributed to the addition of $FeSO_4 \cdot 7H_2O$, which exceeded the
276 decomposition of S during the process. The O/C ratio was boosted from 0.93 for
277 RSC-400 to 1.33 for FSC-400, and the changes of the O/C ratio could correlate

278 with the changes of chemical functional groups including oxygen, which
 279 suggested that a surface oxidation reaction occurred during Fenton pre-oxidation,
 280 and there could be a low content of carboxylic groups on the sample surfaces [35].

281 Table 2 Elemental analysis and calorific value of sludge-based char

Carbonization Temperature/ °C	Elemental analysis/%							Ash/%
T	C	H	O	N	S	H/C	O/C	-
RSC-400	17.10	1.93	15.90	2.53	0.67	0.11	0.93	61.87
RSC-450	13.15	1.37	14.17	2.01	0.97	0.10	1.08	68.33
RSC-500	12.46	1.05	12.08	1.93	0.93	0.08	0.97	71.55
RSC-550	11.31	0.85	11.43	1.66	0.84	0.08	1.01	73.91
RSC-600	11.42	0.63	10.46	1.49	0.73	0.06	0.92	75.27
FSC-400	13.41	1.70	17.90	2.24	1.53	0.13	1.33	63.12
FSC-450	12.37	1.52	17.57	2.03	1.58	0.12	1.42	64.83
FSC-500	12.14	1.31	16.26	1.84	1.65	0.11	1.34	66.7
FSC-550	11.67	1.13	15.08	1.71	1.67	0.10	1.29	68.56
FSC-600	11.72	0.97	14.24	1.59	1.55	0.08	1.22	69.44

282

283 3.2.2 Effect of Fenton pre-oxidation on surface functional groups of char

284 The FTIR spectra of RSC-T and FSC-T were provided in Fig. 3. An
 285 adsorption peak observed between 2300~2420 cm^{-1} may be attributed to the S-H
 286 and C-N stretching vibrations [36]. The adsorption peak around 2305~2320 cm^{-1}
 287 are due to the -COO antisymmetric stretching vibration in carboxyl group or the
 288 stretching vibration of C=O in lactone group [37]. For RS, the peak intensity was
 289 the strongest at 400 °C. With the temperature increasing, the peak gradually
 290 disappeared, but reappeared when the temperature reached 600 °C. The peak
 291 observed in FSC was weak when the temperature rose from 400 to 450 °C.

292 However, the peak became more prominent at 500 °C, before weakening and
293 disappearing at 600 °C. This suggested that the oxidation process caused a
294 rearrangement of the functional groups in the sludge, resulting in changes in the
295 observed peak intensity. The observed trend is consistent with the changes in the N
296 ratio in both types of sludge. The region near 1550-1600 cm⁻¹, which only occurred
297 in FSC-T, corresponded to the stretching vibrations of carbonyl structures of C=O
298 and alkene of C=C bonds. This indicates that the pretreated sludge facilitates the
299 generation of the small molecule organics and indirectly increasing C=O and C=C
300 [38, 39]. It can be found that the surface of the sludge-based char contains various
301 substances or functional groups such as hydroxyl, phenolic hydroxyl, carboxyl,
302 amino, imino, lactone, anhydride, ether, ester and other substances or functional
303 groups [40], revealing that the surface of FSC was enriched with
304 oxygen-containing functional groups, particularly acids and aldehydes, after the
305 pre-oxidation. The stretching vibration adsorption peaks at 1040 cm⁻¹ and 976 ~
306 1038 cm⁻¹ can be endorsed to Si-O-Si or Si-O-C and Fe-O-H, respectively. The
307 peaks observed in FSC-T were evidently higher than those of RSC-T, which
308 showed that the addition of reagent (FeSO₄·7H₂O) introduced new minerals. The
309 adsorption peaks observed at 619 cm⁻¹ corresponded to Si-O-Fe bonds [28, 41].
310 The stretching vibration adsorption peaks at 680-580 cm⁻¹ corresponded to S=O

and O=S=O, which are characteristic of SO_4^{2-} . The presence of these peaks suggests that sulfuric acid groups were introduced onto the surface of char [42, 43], which is consistent with the increase in S element content as shown in Table 2.

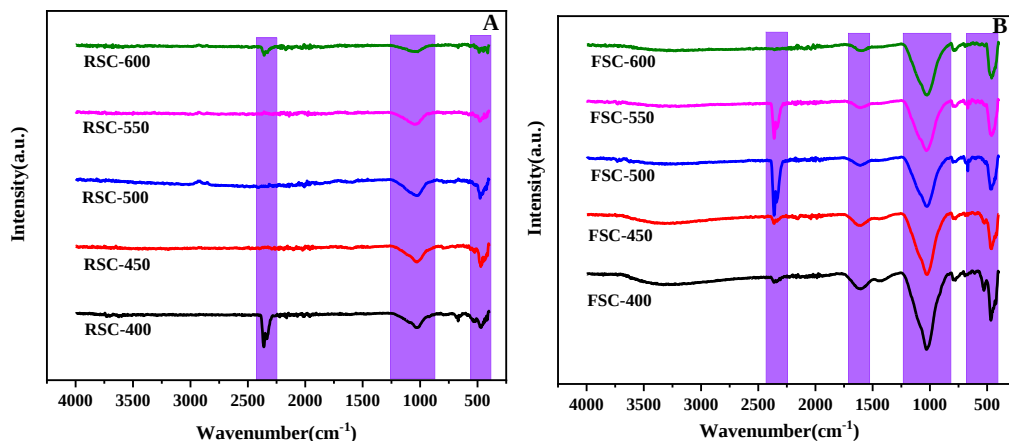


Fig. 3 A. FTIR spectra of RSC-400, RSC-450, RSC-500, RSC-550 and RSC-600; B. FTIR spectra of FSC-400, FSC-450, FSC-500, FSC-550 and FSC-600.

The effect of Fenton pre-oxidation on the carbon-containing groups on the surface of char obtained at 400 °C was investigated using XPS, as depicted in Fig. 4. The proportions of surface element content were analyzed and are presented in Table S1. The C1s peak of RSC-400 and FSC-400 was broken down into six peaks by Casa XPS software, representing typical sp^2 C=C (284.2 eV), sp^3 C-C\C-H (284.6 eV), C-O (285.4-286.0 eV), C-N (protein-N, 285.4eV) and O=C-O (288.2-290.2e) [44] as shown in Fig. 4. There was no similar variation pattern in the C content of the surface of RSC-T and FSC-T (Table S1). However, the C content on the surface of FSC-400 decreased compared to that of RSC-400. This

might be due to the increase of O content resulting from the addition of the oxygen-containing reagent [45], this observation is consistent with the findings presented in Table 2 and Fig. 3. Fig. 4 also show the increase of oxygen containing acidic functional groups of O=C-O and C-O. The C-N content on char of FSC-400 was higher compared to that of RSC-400, which was consistent with previous research results. [46]

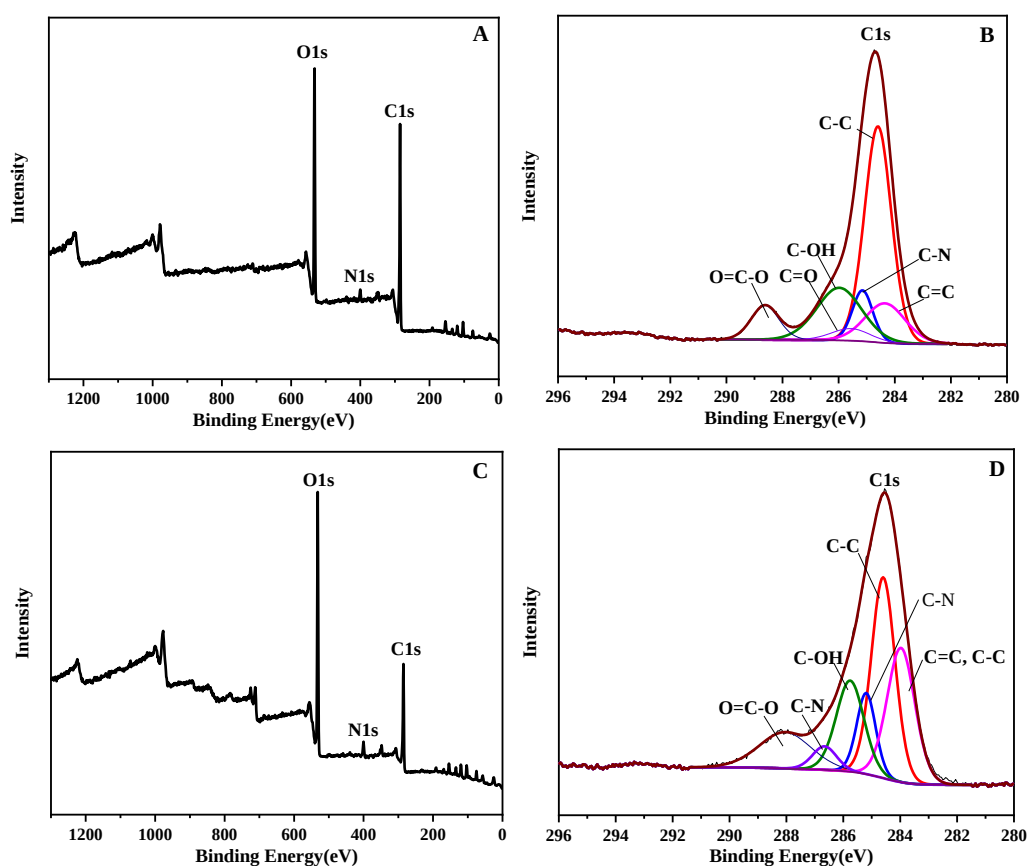


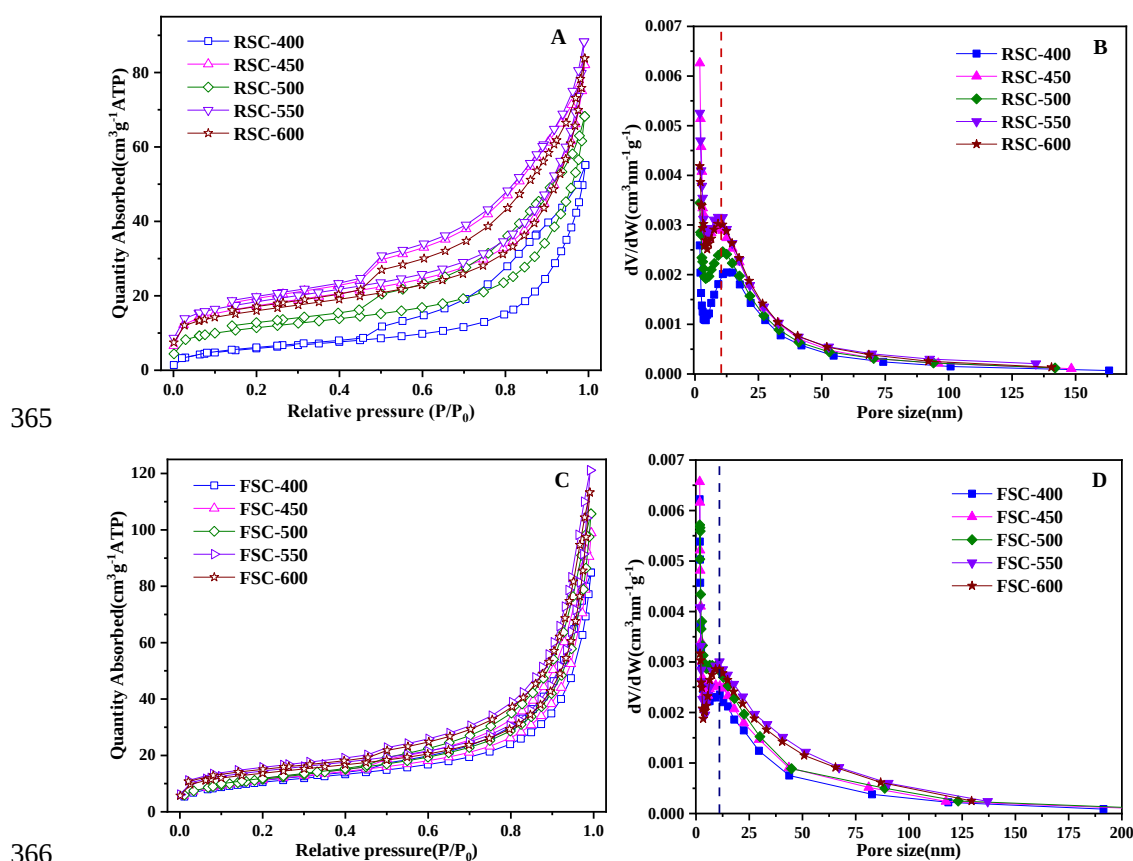
Fig. 4 A. XPS survey of RSC-400; B. Fitting spectra of C1s of RSC-400; C. XPS survey of FSC-400; D. Fitting spectra of C1s of FSC-400.

3.2.3 Effect of Fenton pre-oxidation on morphology characteristics of char

Fig. 5 provides the nitrogen isothermal adsorption profiles and pore size

distributions of RSC-T and FSC-T, while Table 3 summarizes their characteristics of porosity. The micrographs of RSC-400 and FSC-400 are also displayed. According to the classification method of IUPAC, the isothermal adsorption-desorption curves of both sludge-based chars fall into class II, indicating that the adsorption of sludge-based chars was a free single multi-layer reversible adsorption process, and the chars were predominantly composed of microporous and mesoporous structures. From Fig. 5A and C, an obvious enhancement in N₂ adsorption in the relative pressure range $0 < P/P_0 < 0.99$ with carbonization temperature increasing from 400 to 550 °C for both chars, which is mainly attributed to the formation of the great amount of mesopores [37]. This observation is further supported by Fig. 5B and D as well as Table 3, even Fig. 5E and F. The isothermal adsorption-desorption curves revealed that the adsorption capacity of FSC-T was superior to that of RSC-T, with values of 121.13 cm³/g and 88.30 cm³/g, respectively. The amounts of mesopores and macropores in FSC-T were higher than those of RSC-T. As shown in Fig. 5E and F, the char obtained after pretreatment exhibited more pores compared to the FSC at 400 °C, with a higher concentration of metallic oxide on the surface, as also observed in Fig. 3. Using the BET method, the specific surface area of each char was obtained. The variations of S_{BET} of both types of chars were found to be consistent with the

observed adsorption performance in Fig. 5A and C. This correlation was further supported by the changes of the total pore volume. It is worth noting that when the temperature further went up to 600 °C, the adsorption capacity remained almost unchanged, despite the relatively lower S_{BET} . This suggests that other factors, such as pore structure and surface chemistry, may play a significant role in the adsorption behavior of the chars at higher temperatures.



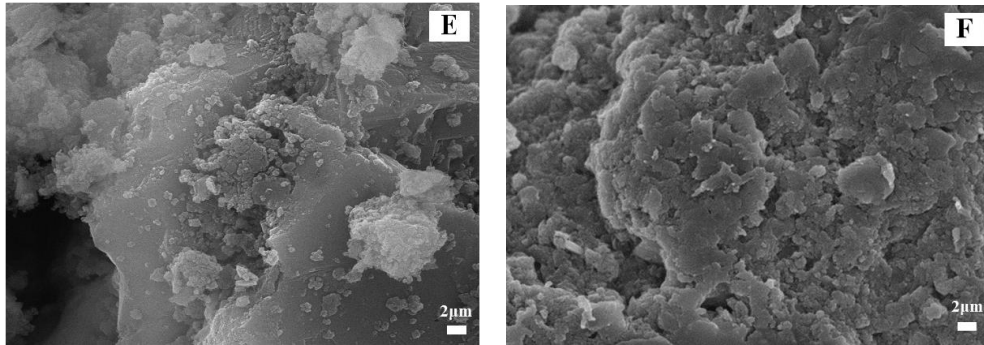


Fig. 5 A. Nitrogen isothermal adsorption profiles; B. pore size distribution of RSC-400, RSC-450, RSC-500, RSC-550 and RSC-600; C. Nitrogen isothermal adsorption profiles; D. pore size distribution of FSC-400, FSC-450, FSC-500, FSC-550 and FSC-600; E. the SEM micrograph for char of RSC-400; F. the SEM micrograph for char of FSC-400.

Table 3 Textural properties of sludge-based char: BET surface area, total pore volume and average pore size

Sample	S_{BET} (m^2/g)	V_t (cm^3/g)	P_A (nm)
RSC-400	21.12	0.085	16.40
RSC-450	40.11	0.12	15.19
RSC-500	40.85	0.10	16.19
RSC-550	48.61	0.12	15.68
RSC-600	46.08	0.12	19.70
FSC-400	37.61	0.13	14.44
FSC-450	60.41	0.15	12.39
FSC-500	61.02	0.16	14.38
FSC-550	63.61	0.18	13.17
FSC-600	55.40	0.17	13.72

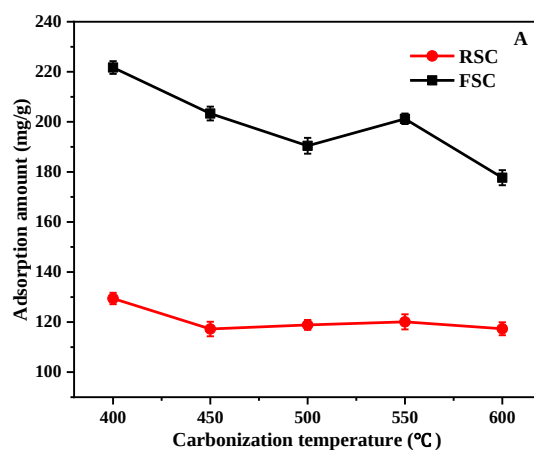
3.3 Ammonia nitrogen adsorption on chars

To compare the adsorption performance of ammonia nitrogen on chars, RSC-T and FSC-T were used as adsorbents, and the dynamics models and adsorption isotherms of ammonia nitrogen adsorption on sludge-based chars were studied (Fig 6). The results indicated that the adsorption capacity of FSC-T was

381 significantly higher than that of RSC-T. Among the tested chars, the highest
382 adsorption capacity of chars was obtained for FSC-400, reaching approximately
383 221.7mg/g. As the temperature increased, the adsorption amount of FSC-400
384 initially continuously declined until 500 °C, and then showed a slight increasing
385 trend at 550 °C. In contrast, the adsorption amounts of RSC remained relatively
386 stable across different temperatures, with a slight decline observed from 400 to
387 450 °C. This difference in behavior might be attributed to the fact that FSC-T had a
388 higher specific surface area and more oxygenated functional groups compared to
389 RSC-T at the same temperature, which makes FSC-T more favorable for the
390 adsorption of ammonia nitrogen. The presence of newly formed Fe (III) and Fe (II)
391 groups could also react with NH_4^+ in an aqueous solution, enhancing the ion
392 exchange process. Fig. 6a further illustrates that FSC-400 had the lowest specific
393 surface area among the FSC-T samples, while it exhibited the highest abundance
394 of oxygenated functional groups (Fig. 3, 4 and Table 2). This observation supports
395 the conclusion that the adsorption process is primarily governed by chemical
396 reactions rather than physical adsorption mechanisms.

397 The results of the dynamics model's equations and adsorption isotherms of
398 FSC are presented in Fig. 6 B and C, as well as Table 4 and 5. The kinetic analysis
399 of ammonia nitrogen adsorption on FSC indicates that it followed

400 pseudo-second-order, as evidenced by the high values of R^2 (greater than 0.9964)
 401 in Fig. 6B. The adsorption for ammonia nitrogen was rapid during the initial 80
 402 min after which it slowed down and eventually reached a plateau with the increase
 403 of reaction time. Fig. 6C describes the adsorption isotherms of ammonia nitrogen
 404 adsorption on FSC-400. The adsorption isotherms were obtained for the ammonia
 405 nitrogen initial concentration ranging from 12.5 mg/L to 100 mg/L. The Langmuir
 406 model provided a better fit to the experimental data compared to the Freundlich
 407 model, as shown in Table 5, indicating that the adsorption process follows
 408 homogeneous monolayer adsorption. In addition, the parameter $1/n=0.7$ ($0 < 0.7 <$
 409 1) obtained from the Langmuir model indicated that the FSC was more effective in
 410 adsorbing ammonia nitrogen.



411

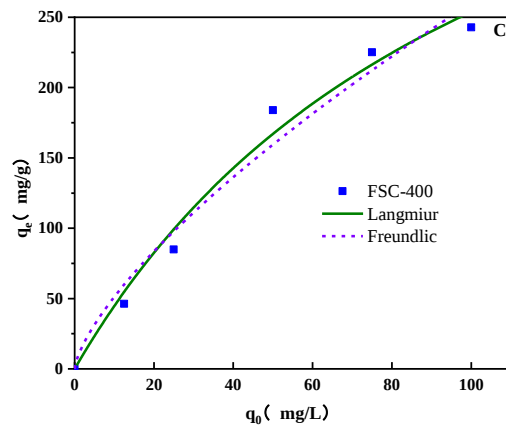
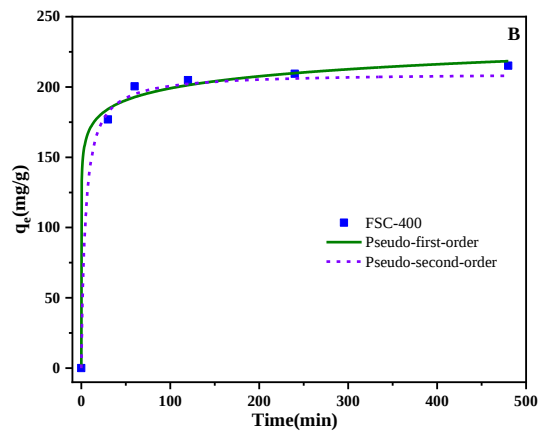


Fig.6 A. Ammonia nitrogen adsorption on sludge-based chars; B. Ammonia nitrogen isotherms on FSC-400 and dynamics fitting models; C. Ammonia nitrogen isotherms on FSC-400 and adsorption fitting models

Table 4 Parameters of adsorption dynamic model

Model	K	q_e	R^2
pseude-first-order	1.23 ± 0.674	204.95 ± 6.72	0.9688
pseude-second-order	0.001 ± 1.745	209.44 ± 5.00	0.9964

Table 5 Parameters of adsorption isotherm model

Model	K	Q/n	R^2
Langmuir	$0.0082 \pm 8.7E-5$	56.35 ± 0.39	0.9959
Freundlich	10.09 ± 4.54	1.42 ± 0.21	0.9885

426 **3.4 The mechanism of pre-oxidation on the adsorption properties of** 427 **sludge-based char**

428 The impact of the pre-oxidation process on FSC in this study can be
429 summarized in four key points: Firstly, the pre-oxidation promoted the
430 cross-linking of internal structures. The bonding conditions indicated an increase
431 in the formation of C-C and C-N bonds. The results might be due to that the
432 oxidation process causing further degradation of proteins, cellulose, and other
433 substances contained in sludge. This degradation leads to the breakdown of
434 long-chain substances, the addition of oxygen atoms, and the generation of a
435 higher concentration of oxygen-containing groups. The oxygen-containing groups
436 had higher reactivity than the macromolecules before oxidation, and new
437 carbon-carbon single bonds and carbon-nitrogen double bonds were formed during
438 the carbonization process along with the dehydration polymerization reaction [11].
439 Additionally, the introduction of Si-O-Si, Si-O-C and Fe-O-H occurred in FSC.
440 These newly formed bonds were more stable and formed faster than those in RSC,
441 and modified the reactions between the char and the ammonia nitrogen, as
442 depicted in Fig. 7. The observed differences in bonding, compared to RSC, suggest
443 a tendency towards cross-linking within the sludge, which can further enhance the
444 adsorption properties of FSC for ammonia nitrogen.

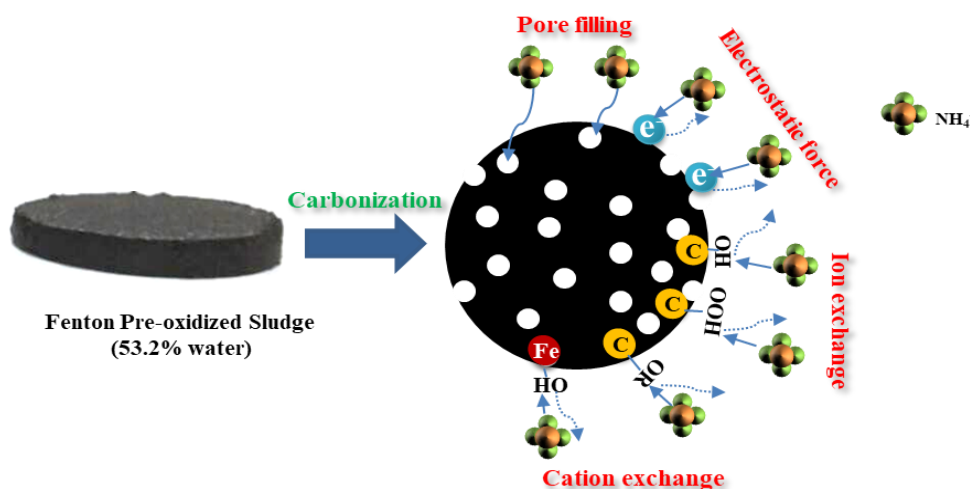
445 Secondly, a large number of oxygen-containing groups were generated and
446 concentrated on the surface of the char. The occurrence of the oxidation process
447 promoted the production of oxygen-containing groups on the surface of the sludge,
448 such as carboxyl and ester groups. As a result, it greatly enhanced the negative
449 charge of the FSC surface, thereby greatly enhancing its ability to attract ammonia
450 ions and improving its adsorption capacity for these ions [47].

451 Thirdly, the pore structure of sludge-char was also altered. The addition of
452 oxygen-containing groups during Fenton oxidation could result in increased
453 decarboxylation reactions during sludge carbonization. The Fenton pre-oxidation
454 process led to the degradation of the biopolymers and the conversion of sludge
455 flocs into small fragments [26]. This transformation has significant benefits for
456 pore development during carbonization, as it enhanced decarboxylation reactions
457 and caused partial destruction of the solid structure, resulting in a wider porosity.
458 Additionally, Fenton oxidation promoted the transition from micropores to
459 mesopores (Fig.5). As a result, the total pore volume of the char increased, and the
460 distribution was more concentrated on mesopores [48]. These changes can
461 enhance the capacity of FSC by improving pore filling, ion and cation exchange.

462 Finally, the influence of carbonization temperature on char quality was
463 mitigated after pre-oxidation. Fenton oxidation promoted the homogenization of

464 sludge and increased the presence of oxygen-containing groups with higher
465 chemical reactivity. This may reduce the activation energy required during the
466 carbonization reaction, resulting in a reduced reaction temperature for
467 carbonization. An advanced char with relatively complete carbonization could
468 already be obtained at 400 °C.

469 In summary, pre-oxidation facilitates the formation of cross-linked carbon
470 bonds during carbonization, and the chemical properties of FSC are more
471 advanced than those of RSC under the same reaction conditions. The oxidation
472 process generates an abundance of oxygen-containing groups on the surface,
473 which increases the reaction activity of the sludge and greatly reduced the
474 activation energy required for the reaction. This also reduces the dependence of
475 char properties on carbonization temperature. The Fenton reaction can promote a
476 more uniform distribution of pores in the char.



477

Fig.7 The adsorption mechanisms of FSC for ammonia nitrogen

3.5 Effect of Fenton pre-oxidation on sludge dewatering and sludge-based char resource utilization

Until now, the utilization of sludge in the form of char has encountered several problems: The high moisture content of sludge requires extra energy consumption because of the high energy cost of water evaporation before and during the reaction [49]; The structure of sludge-based char is usually relatively poor and often requires a high carbonization temperature or the uses of additional catalyst [47]. However, this study shows that: 1. Pre-oxidation can synchronously improve the structure and surface properties of sludge without additional consumption of materials and energy. 2. Pre-oxidation not only reduces the energy consumption associated with deep dehydration, but also lowers the energy requirement of carbonization. This approach allows for significant improvement in the char performance without major modification to the original system. 3. Integrating the dehydration and carbonization process enables better control over the quality of the final product, and adjusts the process in the dehydration and carbonization section as required to impart more unique properties to the char.

4 Conclusions

The Fenton pre-oxidation process resulted in significant surface and structural

497 modifications of the FSC, leading to enhanced adsorption capacity for ammonia
498 nitrogen. Compared to the RSC at the same temperature, FSC exhibited a higher
499 abundance of functional groups on its surface. The addition of oxidants introduced
500 new functional groups to FSC, which could react with NH_4^+ in an aqueous solution.
501 The specific surface area and pore size distribution of FSC became larger and
502 more uniform, facilitating improved diffusion and providing more active sites for
503 adsorption, which is beneficial for FSC to enhance its adsorption capacity by
504 promoting chemisorption and increasing ion interaction with ammonium ions. The
505 adsorption capacity of ammonia nitrogen of FSC was 1.69 times higher than that
506 of RSC, indicating the significant improvement achieved through pre-treatment.
507 The kinetic of ammonia nitrogen adsorption on FSC followed
508 pseudo-second-order, suggesting the involvement of both physical adsorption and
509 chemical adsorption processes. The Langmuir model provided a better fit to the
510 adsorption isotherm data compared to the Freundlich model, indicating that the
511 process was characterized by homogeneous monolayer adsorption. These findings
512 highlight the effectiveness of Fenton pre-oxidation on sludge-based char, and
513 present it as a promising technology by integrating the dehydration and
514 carbonization process with lower energy consumption.

515

516 **CRedit authorship contribution statement**

517 **Jun Cao:** Conceptualization, Investigation, Data curation, original draft. **Ruochen**
518 **Wang:** Investigation, Data curation. **Hualun Zhu:** Review & Editing. **Songshan**
519 **Cao:** Methodology, Review & Editing. **Zhipeng Duan:** Review & Editing.

520

521 **Acknowledgements**

522 This work was supported by National Natural Science Foundation of China
523 (NO.52076067), Natural Science Foundation of Jiangsu Province
524 (NO.BK20201319), National Key R&D Program of China (NO.
525 2022YFC3202705), and the China Postdoctoral Science Foundation
526 (NO.2020M681472).

527

528 **References**

- 529 [1]G. Bertanza, P. Baroni, M. Canato, Ranking Sewage Sludge Management Strategies by
530 Means of Decision Support Systems: A Case Study, Resources, Conservation and
531 Recycling, 110 (2016) 1-15. <https://doi.org/10.1016/j.resconrec.2016.03.011>.
532 [2]V. K. Tyagi, S. Lo, Sludge: A Waste Or Renewable Source for Energy and Resources
533 Recovery? Renewable and Sustainable Energy Reviews, 25 (2013) 708-728.
534 <https://doi.org/10.1016/j.rser.2013.05.029>.
535 [3]L. Wang, L. Xie, X. Feng, X. Li, H. Ma, G. Han, D. Yang, J. Zhou, Optimization Process
536 of Porous Carbon Derived From Coconut Shell: A Novel Preparation Condition and
537 Investigation On Pore Surface Roughness, Diam. Relat. Mat., 123 (2022) 108854.
538 <https://doi.org/10.1016/j.diamond.2022.108854>.
539 [4]H. Liu, L. Yi, Q. Zhang, H. Hu, G. Lu, A. Li, H. Yao, Co-Production of Clean Syngas and
540 Ash Adsorbent During Sewage Sludge Gasification: Synergistic Effect of Fenton

541 Peroxidation and Cao Conditioning, *Appl. Energy*, 179 (2016) 1062-1068.
542 <https://doi.org/10.1016/j.apenergy.2016.07.063>.

543 [5]S. Yildiz, A. Cömert, Fenton Process Effect On Sludge Disintegration, *Int. J. Environ.*
544 *Health Res.*, 30 (2020) 89-104. <https://doi.org/10.1080/09603123.2019.1576162>.

545 [6]L. Chen, W. Zhu, N. Lin, B. Mu, X. Fan, C. Wang, H. Chen, J. Zhong, Mechanism of
546 Separation and Removal of Water From Dewatered Sludge Using L-Dme to Dissolve
547 Hydrophilic Organic Matter, *Chemosphere*, 246 (2020) 125648.
548 <https://doi.org/https://doi.org/10.1016/j.chemosphere.2019.125648>.

549 [7]N. Lin, W. Zhu, X. Fan, C. Wang, C. Chen, H. Zhang, L. Chen, S. Wu, Y. Cui, Key Factor
550 On Improving Secondary Advanced Dewatering Performance of Municipal Dewatered
551 Sludge: Selective Oxidative Decomposition of Polysaccharides, *Chemosphere*, 249
552 (2020) 126108. <https://doi.org/https://doi.org/10.1016/j.chemosphere.2020.126108>.

553 [8]N. Lin, W. Zhu, X. You, X. Wang, J. Zhong, J. Sun, T. Sun, J. Cao, L. Li, Verification of
554 the Mechanism and Effect of Secondary Advanced Dewatering Promoted by Selective
555 Oxidative Decomposition: On Pilot Scale, *Journal of Environmental Chemical*
556 *Engineering*, 10 (2022) 107217. <https://doi.org/10.1016/j.jece.2022.107217>.

557 [9]S. Tao, J. Yang, H. Hou, S. Liang, K. Xiao, J. Qiu, J. Hu, B. Liu, W. Yu, H. Deng,
558 Enhanced Sludge Dewatering Via Homogeneous and Heterogeneous Fenton Reactions
559 Initiated by Fe-Rich Biochar Derived From Sludge, *Chem. Eng. J.*, 372 (2019) 966-977.
560 <https://doi.org/10.1016/j.cej.2019.05.002>.

561 [10]A. D. Grossman, Y. Yang, U. Yogev, D. C. Camarena, G. Oron, R. Bernstein, Effect of
562 Ultrafiltration Membrane Material On Fouling Dynamics in a Submerged Anaerobic
563 Membrane Bioreactor Treating Domestic Wastewater, *Environmental science water*
564 *research & technology*, 5 (2019) 1145-1156. <https://doi.org/10.1039/c9ew00205g>.

565 [11]Y. Z. Belete, E. Ziemann, A. Gross, R. Bernstein, Facile Activation of Sludge-Based
566 Hydrochar by Fenton Oxidation for Ammonium Adsorption in Aqueous Media,
567 *Chemosphere (Oxford)*, 273 (2021) 128526.
568 <https://doi.org/10.1016/j.chemosphere.2020.128526>.

569 [12]R. Wang, W. Zhu, S. Zhao, J. Cao, Hydrothermal Oxidation-Precipitation Method for
570 Recovering Phosphorus From Dewatered Sludge and the Mechanisms Involved, *Sep.*
571 *Purif. Technol.*, 298 (2022) 121580. <https://doi.org/10.1016/j.seppur.2022.121580>.

572 [13]L. Gu, Y. Wang, N. Zhu, D. Zhang, S. Huang, H. Yuan, Z. Lou, M. Wang, Preparation of
573 Sewage Sludge Based Activated Carbon by Using Fenton'S Reagent and their Use in
574 2-Naphthol Adsorption, *Bioresour. Technol.*, 146 (2013) 779-784.
575 <https://doi.org/10.1016/j.biortech.2013.07.147>.

576 [14]L. Wang, J. Wu, H. Ma, G. Han, D. Yang, Y. Chen, J. Zhou, H₃PO₄ -Assisted Synthesis
577 of Apricot Shell Lignin-Based Activated Carbon for Capacitors: Understanding the Pore
578 Structure/Electrochemical Performance Relationship, *Energy Fuels*, 35 (2021)

8303-8312. <https://doi.org/10.1021/acs.energyfuels.1c00169>.

[15] S. Tao, S. Liang, X. Wu, H. Hou, W. Yu, K. Xiao, B. Liu, S. Yuan, J. Hu, J. Yang, Enhanced Silicon Bioavailability of Biochar Derived From Sludge Conditioned with Fenton's Reagent and Lime, *Sci. Total Environ.*, 806 (2022) 150941. <https://doi.org/10.1016/j.scitotenv.2021.150941>.

[16] L. Wang, X. Feng, X. Li, H. Ma, J. Wu, Y. Chen, J. Zhou, Valorization of Lignin: Application of Lignin-Derived Activated Carbon in Capacitors and Investigation of its Textural Properties and Electrochemical Performance, *Diam. Relat. Mat.*, 122 (2022) 108791. <https://doi.org/10.1016/j.diamond.2021.108791>.

[17] Z. Hu, L. Zhang, L. Zhong, Y. Zhou, J. Xue, Y. Li, Preparation of an Antibacterial Chitosan-Coated Biochar-Nanosilver Composite for Drinking Water Purification, *Carbohydr. Polym.*, 219 (2019) 290-297. <https://doi.org/10.1016/j.carbpol.2019.05.017>.

[18] S. Liang, S. Shi, H. Zhang, J. Qiu, W. Yu, M. Li, Q. Gan, W. Yu, K. Xiao, B. Liu, J. Hu, H. Hou, J. Yang, One-Pot Solvothermal Synthesis of Magnetic Biochar From Waste Biomass: Formation Mechanism and Efficient Adsorption of Cr(VI) in an Aqueous Solution, *Sci. Total Environ.*, 695 (2019) 133886. <https://doi.org/10.1016/j.scitotenv.2019.133886>.

[19] J. Song, J. Yang, S. Liang, Y. Shi, W. Yu, C. Li, X. Xu, J. Xiao, R. Guan, N. Ye, X. Wu, H. Hou, J. Hu, B. Xiao, Red Mud Enhanced Hydrogen Production From Pyrolysis of Deep-Dewatered Sludge Cakes Conditioned with Fenton's Reagent and Red Mud, *Int. J. Hydrog. Energy*, 41 (2016) 16762-16771. <https://doi.org/10.1016/j.ijhydene.2016.06.217>.

[20] A. Magdziarz, S. Werle, Analysis of the Combustion and Pyrolysis of Dried Sewage Sludge by Tga and Ms, *Waste management (Elmsford)*, 34 (2014) 174-179. <https://doi.org/10.1016/j.wasman.2013.10.033>.

[21] L. Nowicki, S. Ledakowicz, Comprehensive Characterization of Thermal Decomposition of Sewage Sludge by Tg–Ms, *J. Anal. Appl. Pyrolysis*, 110 (2014) 220-228. <https://doi.org/https://doi.org/10.1016/j.jaap.2014.09.004>.

[22] F. Wei, J. Cao, X. Zhao, J. Ren, B. Gu, X. Wei, Formation of Aromatics and Removal of Nitrogen in Catalytic Fast Pyrolysis of Sewage Sludge: A Study of Sewage Sludge and Model Amino Acids, *Fuel*, 218 (2018) 148-154. <https://doi.org/https://doi.org/10.1016/j.fuel.2018.01.025>.

[23] S. Aravind, P. S. Kumar, N. S. Kumar, N. Siddarth, Conversion of Green Algal Biomass Into Bioenergy by Pyrolysis. a Review, *Environ. Chem. Lett.*, 18 (2020) 829-849. <https://doi.org/10.1007/s10311-020-00990-2>.

[24] Y. Li, F. Chang, B. Huang, Y. Song, H. Zhao, K. Wang, Activated Carbon Preparation From Pyrolysis Char of Sewage Sludge and its Adsorption Performance for Organic Compounds in Sewage, *Fuel*, 266 (2020) 117053.

617 <https://doi.org/10.1016/j.fuel.2020.117053>.

618 [25]X. Dai, C. Zhang, L. Zhang, R. Zhang, G. Chen, W. Hu, Thoughts On the Development
619 Direction of Sludge Treatment an Resource Recovery Under the Background of Carbon
620 Neutrality, *Water&Wastewater Engineering*, 57 (2021) 1-5.

621 [26]L. Li, J. Ai, W. Zhang, S. Peng, T. Dong, Y. Deng, Y. Cui, D. Wang, Relationship
622 Between the Physicochemical Properties of Sludge-Based Carbons and the Adsorption
623 Capacity of Dissolved Organic Matter in Advanced Wastewater Treatment: Effects of
624 Chemical Conditioning, *Chemosphere*, 243 (2020) 125333.
625 <https://doi.org/10.1016/j.chemosphere.2019.125333>.

626 [27]L. Gu, N. Zhu, H. Guo, S. Huang, Z. Lou, H. Yuan, Adsorption and Fenton-Like
627 Degradation of Naphthalene Dye Intermediate On Sewage Sludge Derived Porous
628 Carbon, *J. Hazard. Mater.*, 246-247 (2013) 145-153.
629 <https://doi.org/10.1016/j.jhazmat.2012.12.012>.

630 [28]E. Doelsch, A. Masion, J. Rose, W. E. E. Stone, J. Y. Bottero, P. M. Bertsch, Chemistry
631 and Structure of Colloids Obtained by Hydrolysis of Fe(III) in the Presence of SiO₂
632 Ligands, *Colloids and Surfaces A: Physicochemical and Engineering Aspects*, 217
633 (2003) 121-128. [https://doi.org/10.1016/S0927-7757\(02\)00566-6](https://doi.org/10.1016/S0927-7757(02)00566-6).

634 [29]F. Zhao, X. Zhan, H. Xu, G. Zhu, W. Zou, M. Zhu, L. Kang, Y. Guo, X. Zhao, Z. Wang,
635 W. Tang, New Insights Into Eutrophication Management: Importance of Temperature
636 and Water Residence Time, *J. Environ. Sci.*, 111 (2022) 229-239.
637 <https://doi.org/10.1016/j.jes.2021.02.033>.

638 [30]S. Jellali, L. El-Bassi, Y. Charabi, M. Usman, B. Khiari, M. Al-Wardy, M. Jeguirim,
639 Recent Advancements On Biochars Enrichment with Ammonium and Nitrates From
640 Wastewaters: A Critical Review On Benefits for Environment and Agriculture, *J.*
641 *Environ. Manage.*, 305 (2022) 114368. <https://doi.org/10.1016/j.jenvman.2021.114368>.

642 [31]E. A. Ustinov, Effect of Crystallization and Surface Potential On the Nitrogen Adsorption
643 Isotherm On Graphite: A Refined Monte Carlo Simulation, *Carbon*, 100 (2016) 52-63.
644 <https://doi.org/10.1016/j.carbon.2015.12.099>.

645 [32]H. Chung, W. Kim, J. Park, J. Cho, T. Jeong, P. Park, Application of Langmuir and
646 Freundlich Isotherms to Predict Adsorbate Removal Efficiency Or Required Amount of
647 Adsorbent, *J. Ind. Eng. Chem.*, 28 (2015) 241-246.
648 <https://doi.org/10.1016/j.jiec.2015.02.021>.

649 [33]M. M. Hasan, X. Hu, R. Gunawan, C. Li, Pyrolysis of Large Mallee Wood Particles:
650 Temperature Gradients within a Pyrolysing Particle and Effects of Moisture Content,
651 *Fuel Process. Technol.*, 158 (2017) 163-171.
652 <https://doi.org/10.1016/j.fuproc.2016.12.018>.

653 [34]T. Wang, X. Liu, D. Wang, Z. Gong, B. Si, Y. Zhai, Persulfate Assisted Hydrothermal

Processing of Spirulina for Enhanced Deoxidation Carbonization, *Bioresour. Technol.*,
 322 (2021) 124543. <https://doi.org/10.1016/j.biortech.2020.124543>.

[35] K. Wu, W. Ying, Z. Shi, H. Yang, Z. Zheng, J. Zhang, J. Yang, Fenton Reaction-Oxidized
 Bamboo Lignin Surface and Structural Modification to Reduce Nonproductive Cellulase
 Binding and Improve Enzyme Digestion of Cellulose, *ACS Sustain. Chem. Eng.*, 6
 (2018) 3853-3861. <https://doi.org/10.1021/acssuschemeng.7b04191>.

[36] J. Alvarez, M. Amutio, G. Lopez, I. Barbarias, J. Bilbao, M. Olazar, Sewage Sludge
 Valorization by Flash Pyrolysis in a Conical Spouted Bed Reactor, *Chem. Eng. J.*, 273
 (2015) 173-183. <https://doi.org/https://doi.org/10.1016/j.cej.2015.03.047>.

[37] D. Li, J. Yang, Y. Zhao, H. Yuan, Y. Chen, Ultra-Highly Porous Carbon From Wasted
 Soybean Residue with Tailored Porosity and Doped Structure as Renewable
 Multi-Purpose Absorbent for Efficient CO₂, Toluene and Water Vapor Capture, *J. Clean.
 Prod.*, 337 (2022) 130283. <https://doi.org/10.1016/j.jclepro.2021.130283>.

[38] S. Buytoz, M. M. Yildirim, H. Eren, Microstructural and Microhardness Characteristics of
 Gas Tungsten Arc Synthesized Fe–Cr–C Coating On Aisi 4340, *Mater. Lett.*, 59 (2005)
 607-614. <https://doi.org/10.1016/j.matlet.2004.08.038>.

[39] R. R. N. Marques, F. Stüber, K. M. Smith, A. Fabregat, C. Bengoa, J. Font, A. Fortuny, S.
 Pullket, G. D. Fowler, N. J. D. Graham, Sewage Sludge Based Catalysts for Catalytic
 Wet Air Oxidation of Phenol: Preparation, Characterisation and Catalytic Performance,
Applied Catalysis B: Environmental, 101 (2011) 306-316.
<https://doi.org/10.1016/j.apcatb.2010.09.033>.

[40] G. Q. LU, Preparation and Evaluation of Adsorbents From Waste Carbonaceous Materials
 for SO_x and NO_x Removal, *Environmental Progress*, 15 (1996) 12-18.
<https://doi.org/10.1002/ep.670150114>.

[41] S. Yuan, X. Dai, Facile Synthesis of Sewage Sludge-Derived Mesoporous Material as an
 Efficient and Stable Heterogeneous Catalyst for Photo-Fenton Reaction, *Applied
 Catalysis B: Environmental*, 154-155 (2014) 252-258.
<https://doi.org/10.1016/j.apcatb.2014.02.031>.

[42] M. Roman, W. T. Winter, Effect of Sulfate Groups From Sulfuric Acid Hydrolysis On the
 Thermal Degradation Behavior of Bacterial Cellulose, *Biomacromolecules* (2004)
 1671-1677. <https://doi.org/10.1021/bm034519>.

[43] G. Li, B. Shen, F. Lu, The Mechanism of Sulfur Component in Pyrolyzed Char From
 Waste Tire On the Elemental Mercury Removal, *Chem. Eng. J.*, 273 (2015) 446-454.
<https://doi.org/10.1016/j.cej.2015.03.040>.

[44] M. A. Abdoul, M. S. Islam, Y. Chen, L. Weng, J. Li, J. Ma, Y. Li, Enhanced Adsorption
 of Polystyrene Nanoplastics (Psnps) Onto Oxidized Corn Cob Biochar with High
 Pyrolysis Temperature, *Sci. Total Environ.*, 784 (2021) 147115.
<https://doi.org/10.1016/j.scitotenv.2021.147115>.

- 692 [45]E. Neyens, J. Baeyens, A Review of Classic Fenton'S Peroxidation as an Advanced
693 Oxidation Technique, 2003. [https://doi.org/10.1016/S0304-3894\(02\)00282-0](https://doi.org/10.1016/S0304-3894(02)00282-0).
- 694 [46]C. Gong, J. Jiang, D. Li, Ultrasound Coupled with Fenton Oxidation Pre-Treatment of
695 Sludge to Release Organic Carbon, Nitrogen and Phosphorus, *Sci. Total Environ.*, 532
696 (2015) 495-500. <https://doi.org/10.1016/j.scitotenv.2015.05.131>.
- 697 [47]S. Xie, G. Yu, C. Li, J. Li, G. Wang, S. Dai, Y. Wang, Treatment of High-Ash Industrial
698 Sludge for Producing Improved Char with Low Heavy Metal Toxicity, *J. Anal. Appl.*
699 *Pyrolysis*, 150 (2020) 104866. <https://doi.org/10.1016/j.jaap.2020.104866>.
- 700 [48]J. Shi, H. Cui, J. Xu, N. Yan, C. Zhang, S. You, Fabrication of Nitrogen Doped and
701 Hierarchically Porous Carbon Flowers for Co₂ Adsorption, *J. CO₂ Util.*, 51 (2021)
702 101617. <https://doi.org/10.1016/j.jcou.2021.101617>.
- 703 [49]X. Xu, R. Tu, Y. Sun, Y. Wu, E. Jiang, J. Zhen, The Influence of Combined Pretreatment
704 with Surfactant/Ultrasonic and Hydrothermal Carbonization On Fuel Properties,
705 *Pyrolysis and Combustion Behavior of Corn Stalk*, *Bioresour. Technol.*, 271 (2019)
706 427-438. <https://doi.org/10.1016/j.biortech.2018.09.066>.

707

1 **Effect of Fenton pre-oxidation on the physicochemical properties**
2 **of Sludge-based biochar and its adsorption mechanisms for**
3 **ammonia nitrogen removal**

4 Jun Cao^{a,b,*}, Ruochen Wang^b, Hualun Zhu^c, Songshan Cao^{c,d}, Zhipeng Duan^e

5 ^aNational Engineering Research Center of Water Resources Efficient Utilization and
6 Engineering Safety, Hohai University, Nanjing 211111, China

7 ^bCenter for Taihu Basin, Institute of Water Science and Technology, Hohai University,
8 Nanjing 211111, China

9 ^cDepartment of Chemical Engineering, University College London, London WC1E 7JE,
10 United Kingdom

11 ^dKey Laboratory of Energy Thermal Conversion and Control of Ministry of Education, School
12 of Energy and Environment, Southeast University, Nanjing, 210096, China.

13 ^e College of Environment, Hohai University, Nanjing 211111, China

14

15 * Corresponding author: Jun Cao

16 National Engineering Research Center of Water Resources Efficient Utilization and
17 Engineering Safety, Hohai University, Nanjing 211111, China. Tel.: +86-15850598642;

18 E-mail address: caojunxm@126.com.

19

20

21 **Abstract:**

22 Fenton pre-oxidation, as a promising technology for sludge deep dehydration,
23 has been used intensively; however, the outputs (i.e., low-water-content sludge
24 cakes) need to be further disposed for a full line sludge resource utilization.
25 Carbonization presents a promising approach for disposing of the sludge cake
26 within the process. Here, the physicochemical properties of sludge-based char
27 obtained after Fenton pre-oxidation were studied. Both Fenton pre-oxidized
28 sludge-based char (FSC) and raw sludge-based char (RSC) were produced within
29 the temperature range of 400 °C to 600 °C, and the adsorption performance of the
30 prepared chars, exemplified by their ability to adsorb ammonia nitrogen (NH_4^+),
31 was tested under various conditions. FSC exhibited superior char yield and
32 adsorption performance compared to RSC, with chemical adsorption being the
33 primary controlling factor. The optimum conditions for FSC were observed at
34 400 °C, resulting in a char yield of 72.2%, and an adsorption capacity of 221.7
35 mg/g. Notably, FSC demonstrated an increase in C-C, C-N, O=C-O, C-O and C=C
36 functional groups in comparison to RSC, which suggests that pre-oxidation
37 increases organic matter decomposition and chemical reactivity of the sludge. The
38 introduction of Si-O-Si, Si-O-C and Fe-O-H functional groups in the FSC
39 facilitated ion exchange with NH_4^+ in an aqueous solution. [The key findings](#)

40 indicate that Fenton pre-oxidation not only enhances the adsorption performance
41 of FSC but also reduces the required carbonization temperature for excellent-char
42 preparation. These results demonstrated the potential to integrate sludge deep
43 dewatering, including pre-oxidation, with char preparation.

44 **Keywords:** Fenton pre-oxidation; sludge-based char; ammonia nitrogen
45 adsorption; Surface composite modification; Structural enhancement

46

47 **1 Introduction**

48 The challenge of achieving comprehensive treatment of sludge is becoming
49 increasingly prevalent worldwide. With the growing collection and treatment of
50 domestic sewage in China, the country's sewage treatment capacity has
51 significantly increased [1]. Consequently, the disposal of sewage sludge has also
52 experienced a rapid rise, reaching approximately 60 million tons in 2020 (80%
53 water content). There is an urgent need to develop a technology that allows for
54 sludge recycling while minimizing its volume and associated environmental
55 impact. However, the high moisture content of sludge, which can reach up to
56 99.5wt% [2], currently poses a major obstacle disposal method due to considerable
57 energy consumption and operational costs associated with the sequent thermal
58 conversion technologies required for further treatment of sludge. The Fenton or

59 Fenton-like pre-oxidation method is widely recognized as a crucial step in the
60 pretreatment process for achieving deep water removal [3]. This method is
61 performed prior to subsequent thermal disposal and disintegration [4, 5]. In our
62 previous study, we modified this technology and reduced the water content from
63 82.5% to 53.2% under optimal conditions (mole ratio H_2O_2 : Fe^{2+} =2) [6,8]. This
64 modification effectively reduced the volume and elevated the low calorific value
65 of the sludge. However, it resulted in the production of a significant amount of
66 Fe-rich sludge cake, which presents challenges for proper disposal [9]. The
67 obtained sludge cake typically retains a significant amount of organic matter,
68 making it a key research focus to develop sustainable treatment methods in line
69 with the goals of “emission peak” and “carbon neutralization”.

70 Sludge carbonization is widely regarded as a promising method to accomplish
71 this objective [10,12]. Carbonization can greatly reduce the weight and volume by
72 decomposing and evaporating water, while simultaneously facilitating energy
73 conservation and the acquisition of byproducts. Sludge-based char, due to its large
74 specific surface area, well-developed microstructure, and abundant oxygen
75 functional groups and aromatic compounds on the surface, exhibits characteristics
76 of both an energy material and an adsorbent. Currently, the char finds significant
77 applications in various fields, including fuel energy, environmental protection,

78 materials and agriculture [3, 13,16]. An important utilization of char involves its
79 ability to adsorb and immobilize heavy metal ions [17], as well as nitrogen,
80 phosphorus and organics in water purification, hereby acting as a cost-effective
81 alternative to commercial adsorbents [18]. The potential for widespread
82 application and expansion of the means depends on whether the char prepared
83 from pre-oxidized sludge has a high utilization value. Studies have demonstrated
84 that the Fenton and Fenton-like char synthesized by carbonization exhibits
85 superior performance compared to that of RS, making them efficient adsorbents
86 for pollutant removal [19]. Until now, studies have investigated the effects of
87 temperature, residence time, preparation materials and additives on the structural
88 characteristics of chars [20,25]. Lin et al. and Li et al. found that Fenton
89 conditioning greatly enhanced the adsorption capacity by affecting the porosity of
90 the sludge [26, 27]. Belete et al. reported that the Fenton reagents increased the
91 acidic group's concentration and the BET surface area, but without changing the
92 morphology [28]. However, our study contributes to the existing body of
93 knowledge by exploring the relatively limited discussions on the impact of Fenton
94 pre-oxidation on the physicochemical properties of char surfaces and its
95 subsequent effects on the adsorption mechanism.

96 In aquatic environment, excessive emission of ammonia nitrogen is a

97 significant contributor to eutrophication, posing a threat to water resources
98 worldwide [29]. This issue has garnered global attention and has become a hot
99 topic of concern. Consequently, the discharge standards of ammonia nitrogen have
100 become increasingly stringent both domestically and internationally, and the
101 efficient removal of ammonia nitrogen from water has become a crucial objective
102 in safeguarding aquatic environment. The biochar derived from rice straw, wheat
103 straw, coconut shell, and other materials has been identified as a simple and
104 low-cost adsorbent for ammonia nitrogen removal from water [30]. This is due to
105 its large specific surface area, strong ion exchange ability, and the presence of a
106 negative charge on the surface. Previous studies have shown that there are more
107 oxygenated functional groups formed, and the newly Fe (III) and Fe (II) groups
108 introduced on the Fenton and Fenton-like char. These groups may enhance the ion
109 exchange and provide an increased number of negative charges during the
110 adsorption process [9, 19]. Considering these findings, the removal of ammonia
111 nitrogen using Fenton pre-oxidized sludge-based char has been identified as a
112 potential application direction due to the favorable characteristics of the char for
113 this purpose.

114 Above all, in order to evaluate the mechanism of pre-oxidation on the
115 adsorption properties of sludge-based char, both Fenton pre-oxidized sludge (FS)

116 and the raw sludge (RS) were used to produce chars under different carbonization
117 temperatures and residence times. The influences of Fenton pre-oxidation on char
118 yield, element composition, specific surface area, microstructure, and functional
119 groups were analyzed. Furthermore, the adsorption performance of ammonia
120 nitrogen on the char and the corresponding adsorption mechanism were also
121 studied.

122 **2 Materials and methods**

123 **2.1 Materials**

124 Municipal Sludge samples were collected from Xushe Wastewater Treatment
125 Plant located in Yixing city, China. The plant has a sewage treatment capacity of
126 10,000 tons per day. The plant employs the A²O+filter processing system for
127 municipal wastewater treatment. The RS was obtained from the secondary
128 sedimentation tank, with a moisture content of 82.5%. The FS was prepared
129 following the procedures and the optimum conditions obtained in our previous
130 study [7]. The mole ratio H₂O₂: Fe²⁺ was 2. The content of catalyst FeSO₄·7H₂O
131 was 6% (m/m), and the concentration of oxidants H₂O₂ was 30%. Under these
132 specified conditions, the moisture content of FS was successfully reduced to
133 53.2%. The higher heating values (HHVs) of sludge rose with the addition of
134 oxidants, increased from 0.06 to 7.97 MJ/kg. FeSO₄·7H₂O and H₂O₂ (30%) were

both purchased from Sinopharm Chemical Reagent Co., Ltd. The samples were then crushed and screened without pre-drying before all experiments. The basic property parameters of sample RS and FS were listed in Table 1.

Table 1 Basic property parameters of sludge

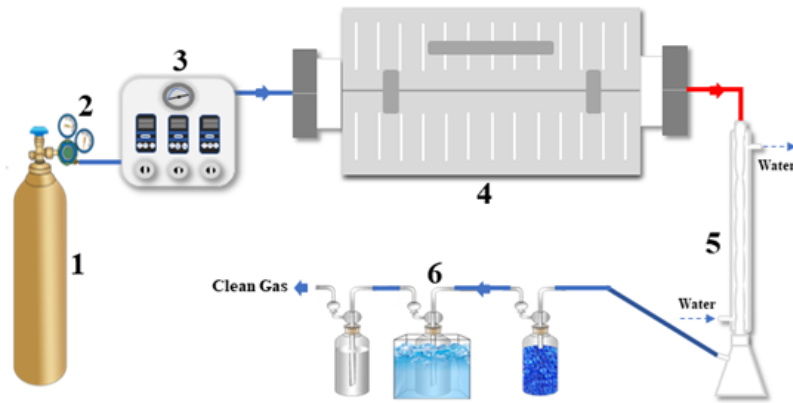
Sample	Proximate Analysis (wt%, daf)			Elemental Analysis (%)					pH	Q_{ar}
	A_{ar}	V_{ar}	FC_{ar}	C	H	O	N	S	-	$MJ \cdot kg^{-1}$
RS	45.81	50.27	6.92	25.35	4.89	16.99	2.86	0.99	6.78	0.06
FS	42.34	51.35	3.31	23.56	4.39	18.44	3.45	1.84	4.95	7.97

2.2 Materials preparation

The sludge-based char was prepared using a system illustrated in Fig.1. An N_2 flow, which controlled by a mass flow controller, was used to maintain an inert atmosphere. The fixed bed furnace employed a quartz tube with a length of 800 mm and an inside diameter of 40 mm, heated externally by a resistance furnace controlled by a temperature programmed controller. One thermocouple was placed inside the tube to monitor the real-time temperature of the reaction zone, while the other thermocouple was set outside the tube to monitor the temperature of the reactor. The gas produced during the experiment was purified by passing it through a contented condensing device and a tail gas treatment system.

The preparation of the char proceeded as follows: First, the sludge was placed in the middle of the tube. To remove all oxygen out of the tube, N_2 was pumped-in

152 with a flow rate of 200 ml/min for 30 min after leak checking. Then the reactor
 153 was heated at a rate of 10 °C/min, starting from room temperature to the target
 154 temperatures (400, 450, 500, 550 and 600 °C) and maintained for a certain time
 155 (30, 60, 90 and 120 min). The reactor was turned off and allowed to cool down
 156 until the furnace temperature dropped below 150 °C. Subsequently, the samples
 157 were removed from the reactor, weighted and collected. The resulting products
 158 were denoted as RSC-T and FSC-T, where T represents the target temperature. The
 159 sludge-based char yield was calculated according to the Formula 1:



160
 161 1. N₂ gas cylinder; 2. Pressure reduce regulator; 3. Mass flowmeter and controller; 4.
 162 Fixed bed with electric furnace; 5. Condensing device; 6. Tail gas treatment
 163 Fig. 1 Sludge char preparation experimental schematic diagram
 164

$$\eta = \frac{m_0 - m_1}{m_0} \times 100\% \quad (1)$$

166 where η is char yield, %; m_0 is original sludge mass, g; m_1 is mass of
 167 obtained sludge-based char, g.

168 **2.3 Characterization and Analysis**

169 An element analyzer (EA, EURO EA3000, Italy) was applied to determine the
170 contents of C, H, O, N and S in sludge and char obtained at different temperatures.
171 An acidimeter (pH, NY/T 1377.2007, China) was used to analyze the pH level and
172 conductivity of the samples. Fourier infrared analysis (FTIR, Model Vextor 22,
173 Germany) was conducted from 4000 to 400 cm⁻¹ and each spectrum was a spectral
174 average of at least four scans. X-ray photoelectron microscopy (XPS, Thermo
175 Scientific ESCALAB250Xi, USA) analysis was performed and the binding energy
176 was corrected by contaminated char (C1s, 284.6eV). The CasaXPS program was
177 used to fit and divide peaks. The N₂ isotherm adsorption profile was obtained
178 using an automated gas sorption analyzer (ASAP2020, Micomeritics, USA) at
179 77K. The specific surface areas (S_{BET}) of samples were calculated according to the
180 Brunauer, Emmett, Teller method (BET) from the adsorption isotherms. The pore
181 size (P_A) distribution of samples before and after calcination was calculated from
182 desorption branch using the Barrett, Joyner and Halenda (BJH) method. Scanning
183 electron microscope (SEM, S4800, Japan) was used to analyze the surface
184 morphology of chars.

185 **2.4 Ammonia nitrogen adsorption**

186 The adsorption experiments were conducted using a set of 50 mL centrifuge

187 tube containing 0.1 g of sludge char and 50 mL of NH_4^+ with an initial
 188 concentration of 50 mg/L in an aqueous solution. The tubes were placed on a
 189 shaker at 20 °C and 120 r/min. The samples were then filtered and the residual
 190 concentration of NH_4^+ was analyzed using 420 nm of Nessler reagent
 191 spectrophotometry. The adsorption capacity of NH_4^+ was calculated as follows:

$$q_e = \frac{(C_0 - C_e) \times V}{m_1} \quad (2)$$

193 Where, q_e represents the adsorption amount per unit mass of sludge char at
 194 adsorption equilibrium, mg/g. C_0 and C_e represent the initial and equilibrium
 195 concentrations of NH_4^+ in solution respectively, mg/L. V represents the volume of
 196 solution, mL.

197 For the adsorption kinetic study, a NH_4^+ solution with 50 mg/L initial
 198 concentration was applied. The behavior of the adsorption with the function of
 199 time was fitted using the [pseudo-first-order \(3\) and pseudo-second-order \(4\) model](#)
 200 [\[31\]](#), as followed:

$$q_t = q_e (1 - e^{-K_1 t}) \quad (3)$$

$$\frac{t}{q_t} = \frac{1}{K_2 q_e^2} + \frac{1}{q_e} t \quad (4)$$

203 Where, t represents reaction time, min. q_t represents the sorption capacity at
 204 t time, mg/g. K_1 and K_2 represent rate constants of the two models, respectively,

205 g/(mg·min).

206 The data were fitted to Langmuir and Freundlich model [32], the equations
207 were as follows:

$$208 \quad q_e = \frac{q_m K_L C_e}{1 + K_L C_e} \quad (5)$$

209 Linearized and obtained,

$$210 \quad \frac{C_e}{q_e} = \frac{1}{K_L q_m} + \frac{C_e}{q_m} \quad (6)$$

$$211 \quad q_e = K_F C_e^{\frac{1}{n}} \quad (7)$$

212 Linearized and obtained,

$$213 \quad \ln q_e = \ln K_F + \frac{1}{n} \ln C_e \quad (8)$$

214 Where q_m represents saturated adsorption capacity, mg/g. K_L represents
215 Langmuir adsorption constant. K_F represents the Langmuir adsorption constant. C_e
216 represents the equilibrium mass concentration of ammonia nitrogen, mg/L. n
217 represents the empirical constant.

218 3 Results and discussion

219 3.1 Carbonization

220 3.1.1 Effect of carbonization temperature

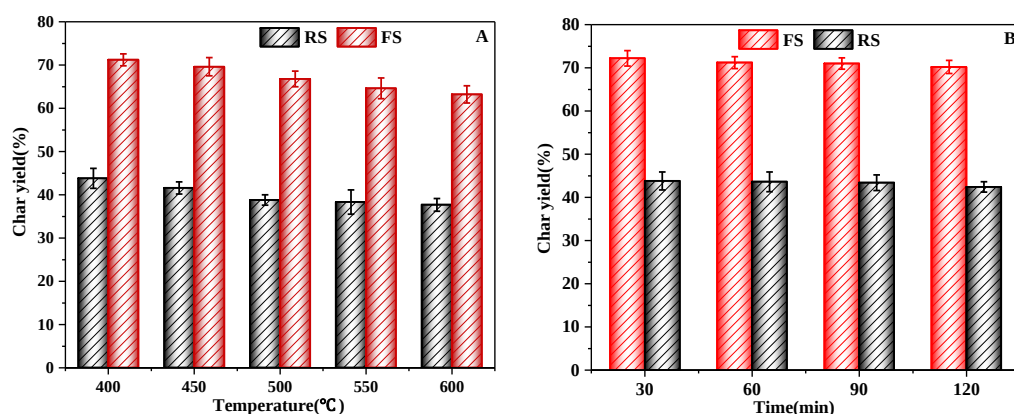
221 Fig. 2A showed the char production rate of RS and FS when the
222 carbonization temperature rose from 400 to 600 °C, the sample mass was 5 g and

the residence time was 60 min. It can be observed that the char yield of FSC was higher than that of RSC at each temperature point, with the char production rate of FSC ranging from 62.4% to 72.2%, while that of RSC was 36.67% to 43.8%. Both FS and RS achieved the maximum char yield at 400 °C, which were 72.2% and 43.8% respectively. With the increase of carbonization temperature, the yield of both FSC and RSC decreased. However, as the temperature further increased up to 500°C, the downward trend of sludge char yield became smaller and eventually stabilized at 550 °C. This can be attributed to the decomposition of volatile components in the sludge, which is mostly completed by 500 °C. At this point, further increase in temperature has a lesser effect on the char yield as the remaining components are more stable. The difference in char yield between FS and RS can be attributed to their disparate moisture content [33]. The RS had a moisture content of 82.5%, while the FS had a reduced moisture content of 53.2%. The lower moisture content in FS might lead to higher char yield, as moisture is vaporized during the carbonization process, resulting in a greater mass loss in RS compared to FS.

3.1.2 Effect of residence time

In the study conducted at 400 °C, the effect of residence time on energy consumption and char yield was investigated. The experiments were performed

242 with a sample weight of 5 g and residence times of 30, 60, 90, and 120 min. Fig.
 243 2B shows that as the residence time extended from 30 min to 120 min, the
 244 decrease in char yield was 2% for FS and 1.4% for RS, respectively. This indicates
 245 that the decomposition of organic matter in the sludge had essentially completed
 246 within a residence time of 60 min. Further extending the residence time did not
 247 significantly affect the char yield, suggesting that the carbonization process had
 248 reached a relatively stable state in terms of char formation.



249 Fig. 2 A. Effect of carbonization temperature on char yield: The mass of sample was 5 g, the
 250 residence time was 60 min, the temperature was 400, 450, 500, 550, 600°C. B. Effect of
 251 residence time on char yield: The mass of sample was 5g, the temperature was 400°C and the
 252 residence time was 30, 60, 90 and 120 min.
 253

254 Above all, the carbonization temperature plays a more important role in char
 255 formation and yield than residence time, since a residence time of 60 min is
 256 sufficient to achieve optimal char yield while minimizing energy consumption.

257 3.2 Effect of Fenton pre-oxidation

258 3.2.1 Effect of Fenton pre-oxidation on elemental composition

259 The element composition and ash content of the char derived from different
260 carbonization temperatures, ranging from normal atmospheric temperature to
261 various carbonization temperatures, were carried out. The results were shown in
262 Table 1 and Table 2. According to Table 1, the contents of ash, volatiles and S of
263 cakes pretreated with Fenton agent were higher compared to the RS, which can be
264 attributed to the addition of inorganic oxidants. In Table 2, it can be observed that
265 the carbonization temperature increased, the contents of C, H, O and N, as well as
266 the ratios of H/C, decreased for both FSC and RSC. This reduction indicated that
267 the organic matter in the sludge was more thoroughly decomposed and released in
268 the form of H₂O, CO₂, CO, CH₄, H₂ and NO_x with increasing temperature [4].
269 Thus, the content of carboxyl, hydroxyl, protein and other functional groups on the
270 surface of char progressively decreased, indicating an increased degree of
271 carbonization. As an important parameter to characterize the carbonization level,
272 the decrease of H/C ratio further confirmed the enhancement of the aromaticity
273 and stability of the char. The S content in the char showed slight growth regardless
274 of the temperature, due to the addition of oxidants [34]. The increase of S in the
275 char can be attributed to the addition of FeSO₄·7H₂O, which exceeded the
276 decomposition of S during the process. The O/C ratio was boosted from 0.93 for
277 RSC-400 to 1.33 for FSC-400, and the changes of the O/C ratio could correlate

278 with the changes of chemical functional groups including oxygen, which
 279 suggested that a surface oxidation reaction occurred during Fenton pre-oxidation,
 280 and there could be a low content of carboxylic groups on the sample surfaces [35].

281 Table 2 Elemental analysis and calorific value of sludge-based char

Carbonization Temperature/ °C	Elemental analysis/%							Ash/%
T	C	H	O	N	S	H/C	O/C	-
RSC-400	17.10	1.93	15.90	2.53	0.67	0.11	0.93	61.87
RSC-450	13.15	1.37	14.17	2.01	0.97	0.10	1.08	68.33
RSC-500	12.46	1.05	12.08	1.93	0.93	0.08	0.97	71.55
RSC-550	11.31	0.85	11.43	1.66	0.84	0.08	1.01	73.91
RSC-600	11.42	0.63	10.46	1.49	0.73	0.06	0.92	75.27
FSC-400	13.41	1.70	17.90	2.24	1.53	0.13	1.33	63.12
FSC-450	12.37	1.52	17.57	2.03	1.58	0.12	1.42	64.83
FSC-500	12.14	1.31	16.26	1.84	1.65	0.11	1.34	66.7
FSC-550	11.67	1.13	15.08	1.71	1.67	0.10	1.29	68.56
FSC-600	11.72	0.97	14.24	1.59	1.55	0.08	1.22	69.44

282

283 3.2.2 Effect of Fenton pre-oxidation on surface functional groups of char

284 The FTIR spectra of RSC-T and FSC-T were provided in Fig. 3. An
 285 adsorption peak observed between 2300~2420 cm^{-1} may be attributed to the S-H
 286 and C-N stretching vibrations [36]. The adsorption peak around 2305~2320 cm^{-1}
 287 are due to the -COO antisymmetric stretching vibration in carboxyl group or the
 288 stretching vibration of C=O in lactone group [37]. For RS, the peak intensity was
 289 the strongest at 400 °C. With the temperature increasing, the peak gradually
 290 disappeared, but reappeared when the temperature reached 600 °C. The peak
 291 observed in FSC was weak when the temperature rose from 400 to 450 °C.

292 However, the peak became more prominent at 500 °C, before weakening and
293 disappearing at 600 °C. This suggested that the oxidation process caused a
294 rearrangement of the functional groups in the sludge, resulting in changes in the
295 observed peak intensity. The observed trend is consistent with the changes in the N
296 ratio in both types of sludge. The region near 1550-1600 cm⁻¹, which only occurred
297 in FSC-T, corresponded to the stretching vibrations of carbonyl structures of C=O
298 and alkene of C=C bonds. This indicates that the pretreated sludge facilitates the
299 generation of the small molecule organics and indirectly increasing C=O and C=C
300 [38, 39]. It can be found that the surface of the sludge-based char contains various
301 substances or functional groups such as hydroxyl, phenolic hydroxyl, carboxyl,
302 amino, imino, lactone, anhydride, ether, ester and other substances or functional
303 groups [40], revealing that the surface of FSC was enriched with
304 oxygen-containing functional groups, particularly acids and aldehydes, after the
305 pre-oxidation. The stretching vibration adsorption peaks at 1040 cm⁻¹ and 976 ~
306 1038 cm⁻¹ can be endorsed to Si-O-Si or Si-O-C and Fe-O-H, respectively. The
307 peaks observed in FSC-T were evidently higher than those of RSC-T, which
308 showed that the addition of reagent (FeSO₄·7H₂O) introduced new minerals. The
309 adsorption peaks observed at 619 cm⁻¹ corresponded to Si-O-Fe bonds [28, 41].
310 The stretching vibration adsorption peaks at 680-580 cm⁻¹ corresponded to S=O

and O=S=O, which are characteristic of SO_4^{2-} . The presence of these peaks suggests that sulfuric acid groups were introduced onto the surface of char [42, 43], which is consistent with the increase in S element content as shown in Table 2.

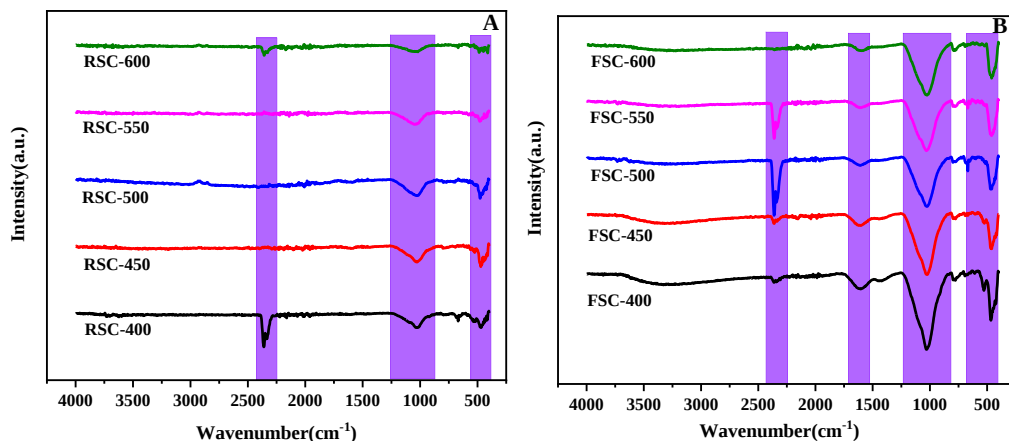


Fig. 3 A. FTIR spectra of RSC-400, RSC-450, RSC-500, RSC-550 and RSC-600; B. FTIR spectra of FSC-400, FSC-450, FSC-500, FSC-550 and FSC-600.

The effect of Fenton pre-oxidation on the carbon-containing groups on the surface of char obtained at 400 °C was investigated using XPS, as depicted in Fig. 4. The proportions of surface element content were analyzed and are presented in Table S1. The C1s peak of RSC-400 and FSC-400 was broken down into six peaks by Casa XPS software, representing typical sp^2 C=C (284.2 eV), sp^3 C-C\C-H (284.6 eV), C-O (285.4-286.0 eV), C-N (protein-N, 285.4eV) and O=C-O (288.2-290.2e) [44] as shown in Fig. 4. There was no similar variation pattern in the C content of the surface of RSC-T and FSC-T (Table S1). However, the C content on the surface of FSC-400 decreased compared to that of RSC-400. This

might be due to the increase of O content resulting from the addition of the oxygen-containing reagent [45], this observation is consistent with the findings presented in Table 2 and Fig. 3. Fig. 4 also show the increase of oxygen containing acidic functional groups of O=C-O and C-O. The C-N content on char of FSC-400 was higher compared to that of RSC-400, which was consistent with previous research results. [46]

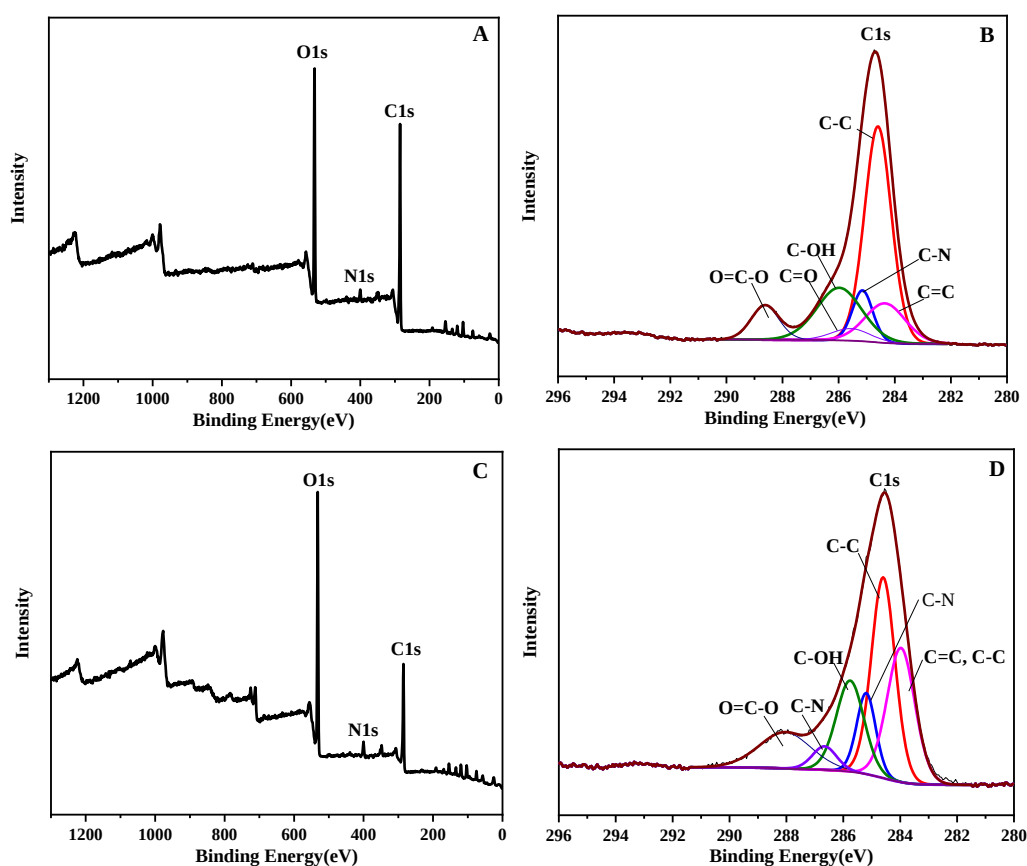


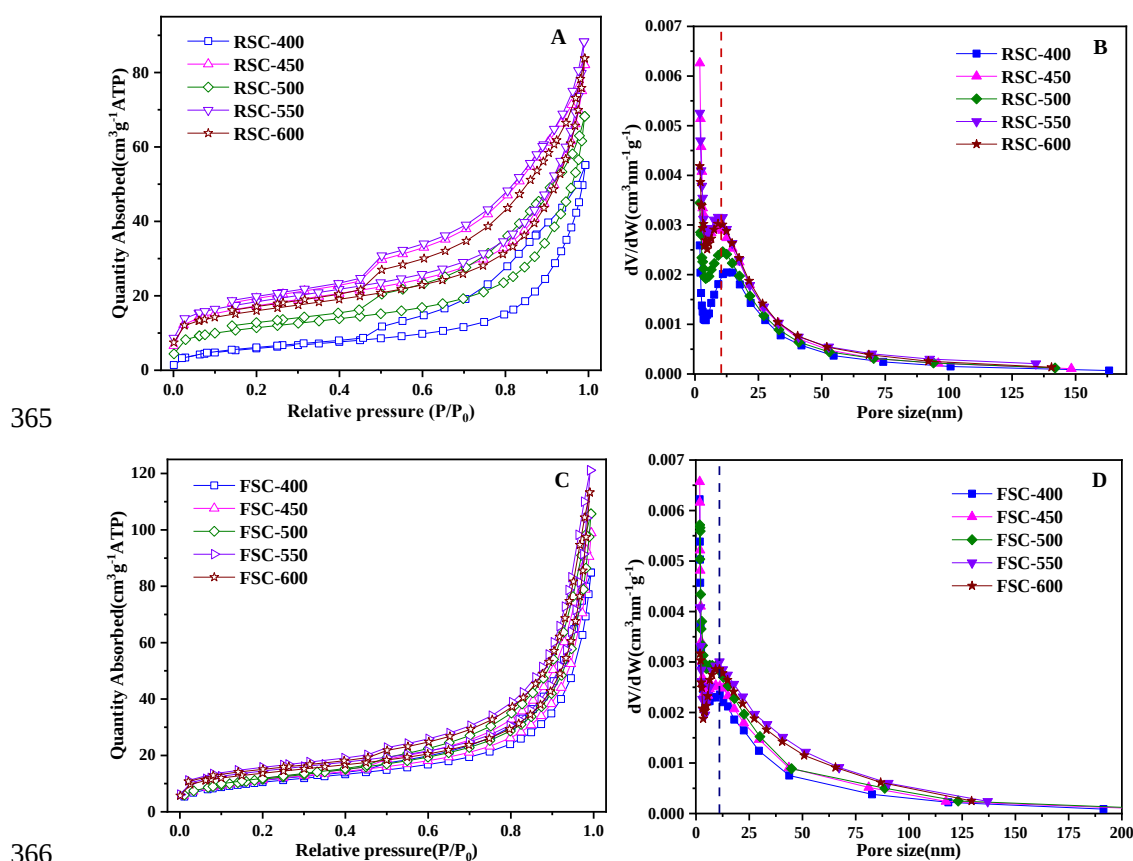
Fig. 4 A. XPS survey of RSC-400; B. Fitting spectra of C1s of RSC-400; C. XPS survey of FSC-400; D. Fitting spectra of C1s of FSC-400.

3.2.3 Effect of Fenton pre-oxidation on morphology characteristics of char

Fig. 5 provides the nitrogen isothermal adsorption profiles and pore size

distributions of RSC-T and FSC-T, while Table 3 summarizes their characteristics of porosity. The micrographs of RSC-400 and FSC-400 are also displayed. According to the classification method of IUPAC, the isothermal adsorption-desorption curves of both sludge-based chars fall into class II, indicating that the adsorption of sludge-based chars was a free single multi-layer reversible adsorption process, and the chars were predominantly composed of microporous and mesoporous structures. From Fig. 5A and C, an obvious enhancement in N₂ adsorption in the relative pressure range $0 < P/P_0 < 0.99$ with carbonization temperature increasing from 400 to 550 °C for both chars, which is mainly attributed to the formation of the great amount of mesopores [37]. This observation is further supported by Fig. 5B and D as well as Table 3, even Fig. 5E and F. The isothermal adsorption-desorption curves revealed that the adsorption capacity of FSC-T was superior to that of RSC-T, with values of 121.13 cm³/g and 88.30 cm³/g, respectively. The amounts of mesopores and macropores in FSC-T were higher than those of RSC-T. As shown in Fig. 5E and F, the char obtained after pretreatment exhibited more pores compared to the FSC at 400 °C, with a higher concentration of metallic oxide on the surface, as also observed in Fig. 3. Using the BET method, the specific surface area of each char was obtained. The variations of S_{BET} of both types of chars were found to be consistent with the

observed adsorption performance in Fig. 5A and C. This correlation was further supported by the changes of the total pore volume. It is worth noting that when the temperature further went up to 600 °C, the adsorption capacity remained almost unchanged, despite the relatively lower S_{BET} . This suggests that other factors, such as pore structure and surface chemistry, may play a significant role in the adsorption behavior of the chars at higher temperatures.



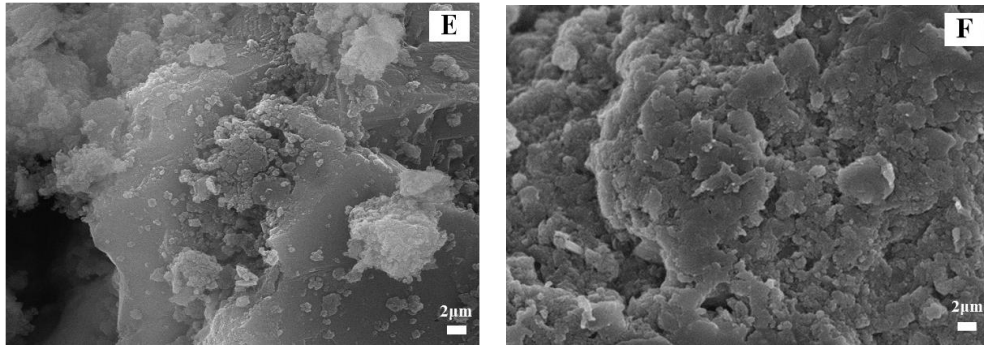


Fig. 5 A. Nitrogen isothermal adsorption profiles; B. pore size distribution of RSC-400, RSC-450, RSC-500, RSC-550 and RSC-600; C. Nitrogen isothermal adsorption profiles; D. pore size distribution of FSC-400, FSC-450, FSC-500, FSC-550 and FSC-600; E. the SEM micrograph for char of RSC-400; F. the SEM micrograph for char of FSC-400.

Table 3 Textural properties of sludge-based char: BET surface area, total pore volume and average pore size

Sample	S_{BET} (m^2/g)	V_t (cm^3/g)	P_A (nm)
RSC-400	21.12	0.085	16.40
RSC-450	40.11	0.12	15.19
RSC-500	40.85	0.10	16.19
RSC-550	48.61	0.12	15.68
RSC-600	46.08	0.12	19.70
FSC-400	37.61	0.13	14.44
FSC-450	60.41	0.15	12.39
FSC-500	61.02	0.16	14.38
FSC-550	63.61	0.18	13.17
FSC-600	55.40	0.17	13.72

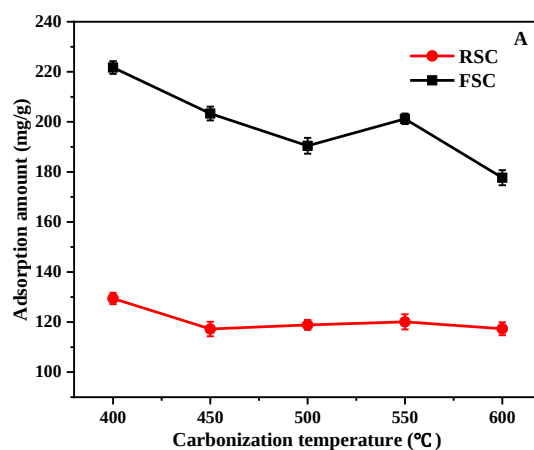
3.3 Ammonia nitrogen adsorption on chars

To compare the adsorption performance of ammonia nitrogen on chars, RSC-T and FSC-T were used as adsorbents, and the dynamics models and adsorption isotherms of ammonia nitrogen adsorption on sludge-based chars were studied (Fig 6). The results indicated that the adsorption capacity of FSC-T was

381 significantly higher than that of RSC-T. Among the tested chars, the highest
382 adsorption capacity of chars was obtained for FSC-400, reaching approximately
383 221.7mg/g. As the temperature increased, the adsorption amount of FSC-400
384 initially continuously declined until 500 °C, and then showed a slight increasing
385 trend at 550 °C. In contrast, the adsorption amounts of RSC remained relatively
386 stable across different temperatures, with a slight decline observed from 400 to
387 450 °C. This difference in behavior might be attributed to the fact that FSC-T had a
388 higher specific surface area and more oxygenated functional groups compared to
389 RSC-T at the same temperature, which makes FSC-T more favorable for the
390 adsorption of ammonia nitrogen. The presence of newly formed Fe (III) and Fe (II)
391 groups could also react with NH_4^+ in an aqueous solution, enhancing the ion
392 exchange process. Fig. 6a further illustrates that FSC-400 had the lowest specific
393 surface area among the FSC-T samples, while it exhibited the highest abundance
394 of oxygenated functional groups (Fig. 3, 4 and Table 2). This observation supports
395 the conclusion that the adsorption process is primarily governed by chemical
396 reactions rather than physical adsorption mechanisms.

397 The results of the dynamics model's equations and adsorption isotherms of
398 FSC are presented in Fig. 6 B and C, as well as Table 4 and 5. The kinetic analysis
399 of ammonia nitrogen adsorption on FSC indicates that it followed

400 pseudo-second-order, as evidenced by the high values of R^2 (greater than 0.9964)
 401 in Fig. 6B. The adsorption for ammonia nitrogen was rapid during the initial 80
 402 min after which it slowed down and eventually reached a plateau with the increase
 403 of reaction time. Fig. 6C describes the adsorption isotherms of ammonia nitrogen
 404 adsorption on FSC-400. The adsorption isotherms were obtained for the ammonia
 405 nitrogen initial concentration ranging from 12.5 mg/L to 100 mg/L. The Langmuir
 406 model provided a better fit to the experimental data compared to the Freundlich
 407 model, as shown in Table 5, indicating that the adsorption process follows
 408 homogeneous monolayer adsorption. In addition, the parameter $1/n=0.7$ ($0 < 0.7 <$
 409 1) obtained from the Langmuir model indicated that the FSC was more effective in
 410 adsorbing ammonia nitrogen.



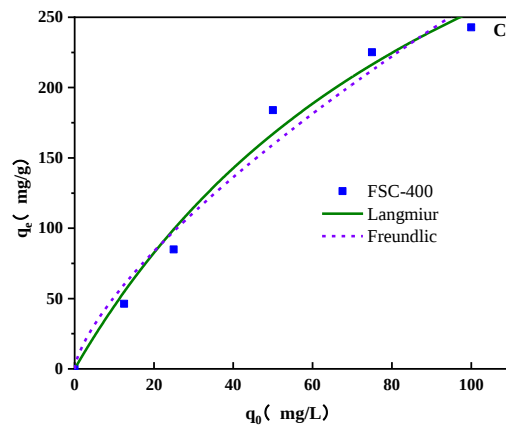
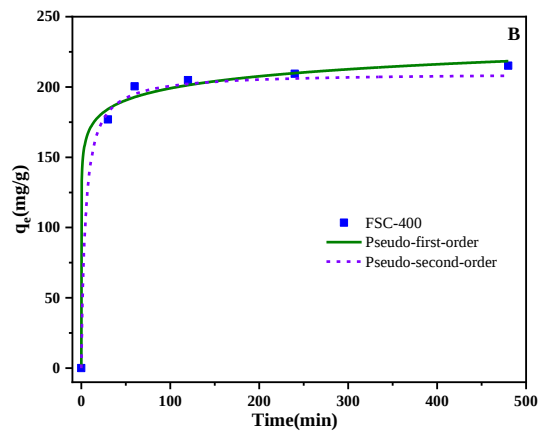


Fig.6 A. Ammonia nitrogen adsorption on sludge-based chars; B. Ammonia nitrogen isotherms on FSC-400 and dynamics fitting models; C. Ammonia nitrogen isotherms on FSC-400 and adsorption fitting models

Table 4 Parameters of adsorption dynamic model

Model	K	q_e	R^2
pseude-first-order	1.23 ± 0.674	204.95 ± 6.72	0.9688
pseude-second-order	0.001 ± 1.745	209.44 ± 5.00	0.9964

Table 5 Parameters of adsorption isotherm model

Model	K	Q/n	R^2
Langmuir	$0.0082 \pm 8.7E-5$	56.35 ± 0.39	0.9959
Freundlich	10.09 ± 4.54	1.42 ± 0.21	0.9885

426 **3.4 The mechanism of pre-oxidation on the adsorption properties of** 427 **sludge-based char**

428 The impact of the pre-oxidation process on FSC in this study can be
429 summarized in four key points: Firstly, the pre-oxidation promoted the
430 cross-linking of internal structures. The bonding conditions indicated an increase
431 in the formation of C-C and C-N bonds. The results might be due to that the
432 oxidation process causing further degradation of proteins, cellulose, and other
433 substances contained in sludge. This degradation leads to the breakdown of
434 long-chain substances, the addition of oxygen atoms, and the generation of a
435 higher concentration of oxygen-containing groups. The oxygen-containing groups
436 had higher reactivity than the macromolecules before oxidation, and new
437 carbon-carbon single bonds and carbon-nitrogen double bonds were formed during
438 the carbonization process along with the dehydration polymerization reaction [11].
439 Additionally, the introduction of Si-O-Si, Si-O-C and Fe-O-H occurred in FSC.
440 These newly formed bonds were more stable and formed faster than those in RSC,
441 and modified the reactions between the char and the ammonia nitrogen, as
442 depicted in Fig. 7. The observed differences in bonding, compared to RSC, suggest
443 a tendency towards cross-linking within the sludge, which can further enhance the
444 adsorption properties of FSC for ammonia nitrogen.

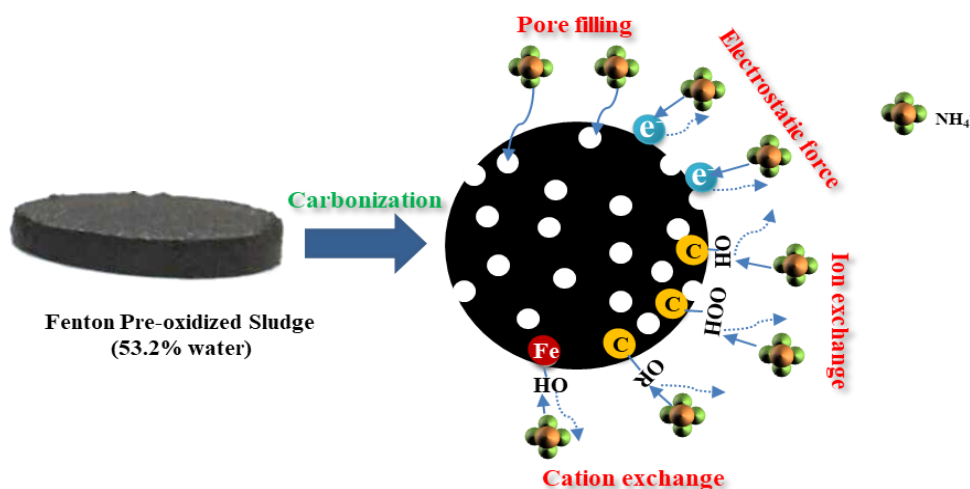
445 Secondly, a large number of oxygen-containing groups were generated and
446 concentrated on the surface of the char. The occurrence of the oxidation process
447 promoted the production of oxygen-containing groups on the surface of the sludge,
448 such as carboxyl and ester groups. As a result, it greatly enhanced the negative
449 charge of the FSC surface, thereby greatly enhancing its ability to attract ammonia
450 ions and improving its adsorption capacity for these ions [47].

451 Thirdly, the pore structure of sludge-char was also altered. The addition of
452 oxygen-containing groups during Fenton oxidation could result in increased
453 decarboxylation reactions during sludge carbonization. The Fenton pre-oxidation
454 process led to the degradation of the biopolymers and the conversion of sludge
455 flocs into small fragments [26]. This transformation has significant benefits for
456 pore development during carbonization, as it enhanced decarboxylation reactions
457 and caused partial destruction of the solid structure, resulting in a wider porosity.
458 Additionally, Fenton oxidation promoted the transition from micropores to
459 mesopores (Fig.5). As a result, the total pore volume of the char increased, and the
460 distribution was more concentrated on mesopores [48]. These changes can
461 enhance the capacity of FSC by improving pore filling, ion and cation exchange.

462 Finally, the influence of carbonization temperature on char quality was
463 mitigated after pre-oxidation. Fenton oxidation promoted the homogenization of

464 sludge and increased the presence of oxygen-containing groups with higher
465 chemical reactivity. This may reduce the activation energy required during the
466 carbonization reaction, resulting in a reduced reaction temperature for
467 carbonization. An advanced char with relatively complete carbonization could
468 already be obtained at 400 °C.

469 In summary, pre-oxidation facilitates the formation of cross-linked carbon
470 bonds during carbonization, and the chemical properties of FSC are more
471 advanced than those of RSC under the same reaction conditions. The oxidation
472 process generates an abundance of oxygen-containing groups on the surface,
473 which increases the reaction activity of the sludge and greatly reduced the
474 activation energy required for the reaction. This also reduces the dependence of
475 char properties on carbonization temperature. The Fenton reaction can promote a
476 more uniform distribution of pores in the char.



477

478 Fig.7 The adsorption mechanisms of FSC for ammonia nitrogen

479 **3.5 Effect of Fenton pre-oxidation on sludge dewatering and sludge-based**
480 **char resource utilization**

481 Until now, the utilization of sludge in the form of char has encountered
482 several problems: The high moisture content of sludge requires extra energy
483 consumption because of the high energy cost of water evaporation before and
484 during the reaction [49]; The structure of sludge-based char is usually relatively
485 poor and often requires a high carbonization temperature or the uses of additional
486 catalyst [47]. However, this study shows that: 1. Pre-oxidation can synchronously
487 improve the structure and surface properties of sludge without additional
488 consumption of materials and energy. 2. Pre-oxidation not only reduces the energy
489 consumption associated with deep dehydration, but also lowers the energy
490 requirement of carbonization. This approach allows for significant improvement in
491 the char performance without major modification to the original system. 3.
492 Integrating the dehydration and carbonization process enables better control over
493 the quality of the final product, and adjusts the process in the dehydration and
494 carbonization section as required to impart more unique properties to the char.

495 **4 Conclusions**

496 The Fenton pre-oxidation process resulted in significant surface and structural

497 modifications of the FSC, leading to enhanced adsorption capacity for ammonia
498 nitrogen. Compared to the RSC at the same temperature, FSC exhibited a higher
499 abundance of functional groups on its surface. The addition of oxidants introduced
500 new functional groups to FSC, which could react with NH_4^+ in an aqueous solution.
501 The specific surface area and pore size distribution of FSC became larger and
502 more uniform, facilitating improved diffusion and providing more active sites for
503 adsorption, which is beneficial for FSC to enhance its adsorption capacity by
504 promoting chemisorption and increasing ion interaction with ammonium ions. The
505 adsorption capacity of ammonia nitrogen of FSC was 1.69 times higher than that
506 of RSC, indicating the significant improvement achieved through pre-treatment.
507 The kinetic of ammonia nitrogen adsorption on FSC followed
508 pseudo-second-order, suggesting the involvement of both physical adsorption and
509 chemical adsorption processes. The Langmuir model provided a better fit to the
510 adsorption isotherm data compared to the Freundlich model, indicating that the
511 process was characterized by homogeneous monolayer adsorption. These findings
512 highlight the effectiveness of Fenton pre-oxidation on sludge-based char, and
513 present it as a promising technology by integrating the dehydration and
514 carbonization process with lower energy consumption.

515

516 **CRedit authorship contribution statement**

517 **Jun Cao:** Conceptualization, Investigation, Data curation, original draft. **Ruochen**
518 **Wang:** Investigation, Data curation. **Hualun Zhu:** Review & Editing. **Songshan**
519 **Cao:** Methodology, Review & Editing. **Zhipeng Duan:** Review & Editing.

520

521 **Acknowledgements**

522 This work was supported by National Natural Science Foundation of China
523 (NO.52076067), Natural Science Foundation of Jiangsu Province
524 (NO.BK20201319), National Key R&D Program of China (NO.
525 2022YFC3202705), and the China Postdoctoral Science Foundation
526 (NO.2020M681472).

527

528 **References**

- 529 [1]G. Bertanza, P. Baroni, M. Canato, Ranking Sewage Sludge Management Strategies by
530 Means of Decision Support Systems: A Case Study, Resources, Conservation and
531 Recycling, 110 (2016) 1-15. <https://doi.org/10.1016/j.resconrec.2016.03.011>.
532 [2]V. K. Tyagi, S. Lo, Sludge: A Waste Or Renewable Source for Energy and Resources
533 Recovery? Renewable and Sustainable Energy Reviews, 25 (2013) 708-728.
534 <https://doi.org/10.1016/j.rser.2013.05.029>.
535 [3]L. Wang, L. Xie, X. Feng, X. Li, H. Ma, G. Han, D. Yang, J. Zhou, Optimization Process
536 of Porous Carbon Derived From Coconut Shell: A Novel Preparation Condition and
537 Investigation On Pore Surface Roughness, Diam. Relat. Mat., 123 (2022) 108854.
538 <https://doi.org/10.1016/j.diamond.2022.108854>.
539 [4]H. Liu, L. Yi, Q. Zhang, H. Hu, G. Lu, A. Li, H. Yao, Co-Production of Clean Syngas and
540 Ash Adsorbent During Sewage Sludge Gasification: Synergistic Effect of Fenton

541 Peroxidation and Cao Conditioning, *Appl. Energy*, 179 (2016) 1062-1068.
542 <https://doi.org/10.1016/j.apenergy.2016.07.063>.

543 [5]S. Yildiz, A. Cömert, Fenton Process Effect On Sludge Disintegration, *Int. J. Environ.*
544 *Health Res.*, 30 (2020) 89-104. <https://doi.org/10.1080/09603123.2019.1576162>.

545 [6]L. Chen, W. Zhu, N. Lin, B. Mu, X. Fan, C. Wang, H. Chen, J. Zhong, Mechanism of
546 Separation and Removal of Water From Dewatered Sludge Using L-Dme to Dissolve
547 Hydrophilic Organic Matter, *Chemosphere*, 246 (2020) 125648.
548 <https://doi.org/https://doi.org/10.1016/j.chemosphere.2019.125648>.

549 [7]N. Lin, W. Zhu, X. Fan, C. Wang, C. Chen, H. Zhang, L. Chen, S. Wu, Y. Cui, Key Factor
550 On Improving Secondary Advanced Dewatering Performance of Municipal Dewatered
551 Sludge: Selective Oxidative Decomposition of Polysaccharides, *Chemosphere*, 249
552 (2020) 126108. <https://doi.org/https://doi.org/10.1016/j.chemosphere.2020.126108>.

553 [8]N. Lin, W. Zhu, X. You, X. Wang, J. Zhong, J. Sun, T. Sun, J. Cao, L. Li, Verification of
554 the Mechanism and Effect of Secondary Advanced Dewatering Promoted by Selective
555 Oxidative Decomposition: On Pilot Scale, *Journal of Environmental Chemical*
556 *Engineering*, 10 (2022) 107217. <https://doi.org/10.1016/j.jece.2022.107217>.

557 [9]S. Tao, J. Yang, H. Hou, S. Liang, K. Xiao, J. Qiu, J. Hu, B. Liu, W. Yu, H. Deng,
558 Enhanced Sludge Dewatering Via Homogeneous and Heterogeneous Fenton Reactions
559 Initiated by Fe-Rich Biochar Derived From Sludge, *Chem. Eng. J.*, 372 (2019) 966-977.
560 <https://doi.org/10.1016/j.cej.2019.05.002>.

561 [10]A. D. Grossman, Y. Yang, U. Yogev, D. C. Camarena, G. Oron, R. Bernstein, Effect of
562 Ultrafiltration Membrane Material On Fouling Dynamics in a Submerged Anaerobic
563 Membrane Bioreactor Treating Domestic Wastewater, *Environmental science water*
564 *research & technology*, 5 (2019) 1145-1156. <https://doi.org/10.1039/c9ew00205g>.

565 [11]Y. Z. Belete, E. Ziemann, A. Gross, R. Bernstein, Facile Activation of Sludge-Based
566 Hydrochar by Fenton Oxidation for Ammonium Adsorption in Aqueous Media,
567 *Chemosphere (Oxford)*, 273 (2021) 128526.
568 <https://doi.org/10.1016/j.chemosphere.2020.128526>.

569 [12]R. Wang, W. Zhu, S. Zhao, J. Cao, Hydrothermal Oxidation-Precipitation Method for
570 Recovering Phosphorus From Dewatered Sludge and the Mechanisms Involved, *Sep.*
571 *Purif. Technol.*, 298 (2022) 121580. <https://doi.org/10.1016/j.seppur.2022.121580>.

572 [13]L. Gu, Y. Wang, N. Zhu, D. Zhang, S. Huang, H. Yuan, Z. Lou, M. Wang, Preparation of
573 Sewage Sludge Based Activated Carbon by Using Fenton'S Reagent and their Use in
574 2-Naphthol Adsorption, *Bioresour. Technol.*, 146 (2013) 779-784.
575 <https://doi.org/10.1016/j.biortech.2013.07.147>.

576 [14]L. Wang, J. Wu, H. Ma, G. Han, D. Yang, Y. Chen, J. Zhou, H₃PO₄ -Assisted Synthesis
577 of Apricot Shell Lignin-Based Activated Carbon for Capacitors: Understanding the Pore
578 Structure/Electrochemical Performance Relationship, *Energy Fuels*, 35 (2021)

8303-8312. <https://doi.org/10.1021/acs.energyfuels.1c00169>.

[15] S. Tao, S. Liang, X. Wu, H. Hou, W. Yu, K. Xiao, B. Liu, S. Yuan, J. Hu, J. Yang, Enhanced Silicon Bioavailability of Biochar Derived From Sludge Conditioned with Fenton's Reagent and Lime, *Sci. Total Environ.*, 806 (2022) 150941. <https://doi.org/10.1016/j.scitotenv.2021.150941>.

[16] L. Wang, X. Feng, X. Li, H. Ma, J. Wu, Y. Chen, J. Zhou, Valorization of Lignin: Application of Lignin-Derived Activated Carbon in Capacitors and Investigation of its Textural Properties and Electrochemical Performance, *Diam. Relat. Mat.*, 122 (2022) 108791. <https://doi.org/10.1016/j.diamond.2021.108791>.

[17] Z. Hu, L. Zhang, L. Zhong, Y. Zhou, J. Xue, Y. Li, Preparation of an Antibacterial Chitosan-Coated Biochar-Nanosilver Composite for Drinking Water Purification, *Carbohydr. Polym.*, 219 (2019) 290-297. <https://doi.org/10.1016/j.carbpol.2019.05.017>.

[18] S. Liang, S. Shi, H. Zhang, J. Qiu, W. Yu, M. Li, Q. Gan, W. Yu, K. Xiao, B. Liu, J. Hu, H. Hou, J. Yang, One-Pot Solvothermal Synthesis of Magnetic Biochar From Waste Biomass: Formation Mechanism and Efficient Adsorption of Cr(VI) in an Aqueous Solution, *Sci. Total Environ.*, 695 (2019) 133886. <https://doi.org/10.1016/j.scitotenv.2019.133886>.

[19] J. Song, J. Yang, S. Liang, Y. Shi, W. Yu, C. Li, X. Xu, J. Xiao, R. Guan, N. Ye, X. Wu, H. Hou, J. Hu, B. Xiao, Red Mud Enhanced Hydrogen Production From Pyrolysis of Deep-Dewatered Sludge Cakes Conditioned with Fenton's Reagent and Red Mud, *Int. J. Hydrog. Energy*, 41 (2016) 16762-16771. <https://doi.org/10.1016/j.ijhydene.2016.06.217>.

[20] A. Magdziarz, S. Werle, Analysis of the Combustion and Pyrolysis of Dried Sewage Sludge by Tga and Ms, *Waste management (Elmsford)*, 34 (2014) 174-179. <https://doi.org/10.1016/j.wasman.2013.10.033>.

[21] L. Nowicki, S. Ledakowicz, Comprehensive Characterization of Thermal Decomposition of Sewage Sludge by Tg–Ms, *J. Anal. Appl. Pyrolysis*, 110 (2014) 220-228. <https://doi.org/https://doi.org/10.1016/j.jaap.2014.09.004>.

[22] F. Wei, J. Cao, X. Zhao, J. Ren, B. Gu, X. Wei, Formation of Aromatics and Removal of Nitrogen in Catalytic Fast Pyrolysis of Sewage Sludge: A Study of Sewage Sludge and Model Amino Acids, *Fuel*, 218 (2018) 148-154. <https://doi.org/https://doi.org/10.1016/j.fuel.2018.01.025>.

[23] S. Aravind, P. S. Kumar, N. S. Kumar, N. Siddarth, Conversion of Green Algal Biomass Into Bioenergy by Pyrolysis. a Review, *Environ. Chem. Lett.*, 18 (2020) 829-849. <https://doi.org/10.1007/s10311-020-00990-2>.

[24] Y. Li, F. Chang, B. Huang, Y. Song, H. Zhao, K. Wang, Activated Carbon Preparation From Pyrolysis Char of Sewage Sludge and its Adsorption Performance for Organic Compounds in Sewage, *Fuel*, 266 (2020) 117053.

- 617 <https://doi.org/10.1016/j.fuel.2020.117053>.
- 618 [25]X. Dai, C. Zhang, L. Zhang, R. Zhang, G. Chen, W. Hu, Thoughts On the Development
619 Direction of Sludge Treatment an Resource Recovery Under the Background of Carbon
620 Neutrality, *Water&Wastewater Engineering*, 57 (2021) 1-5.
- 621 [26]L. Li, J. Ai, W. Zhang, S. Peng, T. Dong, Y. Deng, Y. Cui, D. Wang, Relationship
622 Between the Physicochemical Properties of Sludge-Based Carbons and the Adsorption
623 Capacity of Dissolved Organic Matter in Advanced Wastewater Treatment: Effects of
624 Chemical Conditioning, *Chemosphere*, 243 (2020) 125333.
625 <https://doi.org/10.1016/j.chemosphere.2019.125333>.
- 626 [27]L. Gu, N. Zhu, H. Guo, S. Huang, Z. Lou, H. Yuan, Adsorption and Fenton-Like
627 Degradation of Naphthalene Dye Intermediate On Sewage Sludge Derived Porous
628 Carbon, *J. Hazard. Mater.*, 246-247 (2013) 145-153.
629 <https://doi.org/10.1016/j.jhazmat.2012.12.012>.
- 630 [28]E. Doelsch, A. Masion, J. Rose, W. E. E. Stone, J. Y. Bottero, P. M. Bertsch, Chemistry
631 and Structure of Colloids Obtained by Hydrolysis of Fe(III) in the Presence of SiO₂
632 Ligands, *Colloids and Surfaces A: Physicochemical and Engineering Aspects*, 217
633 (2003) 121-128. [https://doi.org/10.1016/S0927-7757\(02\)00566-6](https://doi.org/10.1016/S0927-7757(02)00566-6).
- 634 [29]F. Zhao, X. Zhan, H. Xu, G. Zhu, W. Zou, M. Zhu, L. Kang, Y. Guo, X. Zhao, Z. Wang,
635 W. Tang, New Insights Into Eutrophication Management: Importance of Temperature
636 and Water Residence Time, *J. Environ. Sci.*, 111 (2022) 229-239.
637 <https://doi.org/10.1016/j.jes.2021.02.033>.
- 638 [30]S. Jellali, L. El-Bassi, Y. Charabi, M. Usman, B. Khiari, M. Al-Wardy, M. Jeguirim,
639 Recent Advancements On Biochars Enrichment with Ammonium and Nitrates From
640 Wastewaters: A Critical Review On Benefits for Environment and Agriculture, *J.*
641 *Environ. Manage.*, 305 (2022) 114368. <https://doi.org/10.1016/j.jenvman.2021.114368>.
- 642 [31]E. A. Ustinov, Effect of Crystallization and Surface Potential On the Nitrogen Adsorption
643 Isotherm On Graphite: A Refined Monte Carlo Simulation, *Carbon*, 100 (2016) 52-63.
644 <https://doi.org/10.1016/j.carbon.2015.12.099>.
- 645 [32]H. Chung, W. Kim, J. Park, J. Cho, T. Jeong, P. Park, Application of Langmuir and
646 Freundlich Isotherms to Predict Adsorbate Removal Efficiency Or Required Amount of
647 Adsorbent, *J. Ind. Eng. Chem.*, 28 (2015) 241-246.
648 <https://doi.org/10.1016/j.jiec.2015.02.021>.
- 649 [33]M. M. Hasan, X. Hu, R. Gunawan, C. Li, Pyrolysis of Large Mallee Wood Particles:
650 Temperature Gradients within a Pyrolysing Particle and Effects of Moisture Content,
651 *Fuel Process. Technol.*, 158 (2017) 163-171.
652 <https://doi.org/10.1016/j.fuproc.2016.12.018>.
- 653 [34]T. Wang, X. Liu, D. Wang, Z. Gong, B. Si, Y. Zhai, Persulfate Assisted Hydrothermal

654 Processing of Spirulina for Enhanced Deoxidation Carbonization, *Bioresour. Technol.*,
655 322 (2021) 124543. <https://doi.org/10.1016/j.biortech.2020.124543>.

656 [35]K. Wu, W. Ying, Z. Shi, H. Yang, Z. Zheng, J. Zhang, J. Yang, Fenton Reaction-Oxidized
657 Bamboo Lignin Surface and Structural Modification to Reduce Nonproductive Cellulase
658 Binding and Improve Enzyme Digestion of Cellulose, *ACS Sustain. Chem. Eng.*, 6
659 (2018) 3853-3861. <https://doi.org/10.1021/acssuschemeng.7b04191>.

660 [36]J. Alvarez, M. Amutio, G. Lopez, I. Barbarias, J. Bilbao, M. Olazar, Sewage Sludge
661 Valorization by Flash Pyrolysis in a Conical Spouted Bed Reactor, *Chem. Eng. J.*, 273
662 (2015) 173-183. <https://doi.org/https://doi.org/10.1016/j.cej.2015.03.047>.

663 [37]D. Li, J. Yang, Y. Zhao, H. Yuan, Y. Chen, Ultra-Highly Porous Carbon From Wasted
664 Soybean Residue with Tailored Porosity and Doped Structure as Renewable
665 Multi-Purpose Absorbent for Efficient CO₂, Toluene and Water Vapor Capture, *J. Clean.*
666 *Prod.*, 337 (2022) 130283. <https://doi.org/10.1016/j.jclepro.2021.130283>.

667 [38]S. Buytoz, M. M. Yildirim, H. Eren, Microstructural and Microhardness Characteristics of
668 Gas Tungsten Arc Synthesized Fe–Cr–C Coating On Aisi 4340, *Mater. Lett.*, 59 (2005)
669 607-614. <https://doi.org/10.1016/j.matlet.2004.08.038>.

670 [39]R. R. N. Marques, F. Stüber, K. M. Smith, A. Fabregat, C. Bengoa, J. Font, A. Fortuny, S.
671 Pullket, G. D. Fowler, N. J. D. Graham, Sewage Sludge Based Catalysts for Catalytic
672 Wet Air Oxidation of Phenol: Preparation, Characterisation and Catalytic Performance,
673 *Applied Catalysis B: Environmental*, 101 (2011) 306-316.
674 <https://doi.org/10.1016/j.apcatb.2010.09.033>.

675 [40]G. Q. LU, Preparation and Evaluation of Adsorbents From Waste Carbonaceous Materials
676 for SO_x and NO_x Removal, *Environmental Progress*, 15 (1996) 12-18.
677 <https://doi.org/10.1002/ep.670150114>.

678 [41]S. Yuan, X. Dai, Facile Synthesis of Sewage Sludge-Derived Mesoporous Material as an
679 Efficient and Stable Heterogeneous Catalyst for Photo-Fenton Reaction, *Applied*
680 *Catalysis B: Environmental*, 154-155 (2014) 252-258.
681 <https://doi.org/10.1016/j.apcatb.2014.02.031>.

682 [42]M. Roman, W. T. Winter, Effect of Sulfate Groups From Sulfuric Acid Hydrolysis On the
683 Thermal Degradation Behavior of Bacterial Cellulose, *Biomacromolecules* (2004)
684 1671-1677. <https://doi.org/10.1021/bm034519>.

685 [43]G. Li, B. Shen, F. Lu, The Mechanism of Sulfur Component in Pyrolyzed Char From
686 Waste Tire On the Elemental Mercury Removal, *Chem. Eng. J.*, 273 (2015) 446-454.
687 <https://doi.org/10.1016/j.cej.2015.03.040>.

688 [44]M. A. Abdoul, M. S. Islam, Y. Chen, L. Weng, J. Li, J. Ma, Y. Li, Enhanced Adsorption
689 of Polystyrene Nanoplastics (Psnps) Onto Oxidized Corn Cob Biochar with High
690 Pyrolysis Temperature, *Sci. Total Environ.*, 784 (2021) 147115.
691 <https://doi.org/10.1016/j.scitotenv.2021.147115>.

- 692 [45]E. Neyens, J. Baeyens, A Review of Classic Fenton'S Peroxidation as an Advanced
693 Oxidation Technique, 2003. [https://doi.org/10.1016/S0304-3894\(02\)00282-0](https://doi.org/10.1016/S0304-3894(02)00282-0).
- 694 [46]C. Gong, J. Jiang, D. Li, Ultrasound Coupled with Fenton Oxidation Pre-Treatment of
695 Sludge to Release Organic Carbon, Nitrogen and Phosphorus, *Sci. Total Environ.*, 532
696 (2015) 495-500. <https://doi.org/10.1016/j.scitotenv.2015.05.131>.
- 697 [47]S. Xie, G. Yu, C. Li, J. Li, G. Wang, S. Dai, Y. Wang, Treatment of High-Ash Industrial
698 Sludge for Producing Improved Char with Low Heavy Metal Toxicity, *J. Anal. Appl.*
699 *Pyrolysis*, 150 (2020) 104866. <https://doi.org/10.1016/j.jaap.2020.104866>.
- 700 [48]J. Shi, H. Cui, J. Xu, N. Yan, C. Zhang, S. You, Fabrication of Nitrogen Doped and
701 Hierarchically Porous Carbon Flowers for Co₂ Adsorption, *J. CO₂ Util.*, 51 (2021)
702 101617. <https://doi.org/10.1016/j.jcou.2021.101617>.
- 703 [49]X. Xu, R. Tu, Y. Sun, Y. Wu, E. Jiang, J. Zhen, The Influence of Combined Pretreatment
704 with Surfactant/Ultrasonic and Hydrothermal Carbonization On Fuel Properties,
705 *Pyrolysis and Combustion Behavior of Corn Stalk*, *Bioresour. Technol.*, 271 (2019)
706 427-438. <https://doi.org/10.1016/j.biortech.2018.09.066>.

707