

25 Because of eutrophication, the East Gotland Basin (EGB; 249 m water depth at its deepest point), located
26 in the central Baltic Sea, has been almost permanently hypoxic below 80 m (i.e. water oxygen
27 concentration $<2 \text{ ml l}^{-1}$) since the 1950s, and even euxinic in its deepest part (Gustafsson et al., 2012;
28 Carstensen et al., 2014). The expansion of oxygen deficient bottom waters in the central Baltic Sea is
29 interrupted for a few months only during Major Baltic Inflows (MBIs), which are intrusions of oxygen-
30 rich waters from the North Sea (Dellwig et al., 2018; Mohrholz, 2018). MBIs are occasional and occur
31 only when long-lasting easterly winds causing a below-normal Baltic sea level are followed by strong
32 westerly winds (Mohrholz et al., 2015). These inflow events are recorded in sediments through the
33 formation of Mn carbonate (or Ca-rich rhodochrosite) and/or through significant increase in specific trace
34 metals such as Co and V (Dellwig et al., 2019; Dellwig et al., 2021). Except for these mm-thick Mn
35 carbonate-rich layers, a hemipelagic-type of sedimentation with a high content of organic carbon (5 –
36 18%) and relatively high sedimentation rates ($0.3 - 0.5 \text{ cm yr}^{-1}$) has prevailed in the EGB since 1950 CE
37 (Struck et al., 2001; Ilus et al., 2007; Kaiser et al., 2020; Lin et al., 2021). The mean sinking velocity of
38 particles (diatoms) has been estimated around $15 \text{ to } 70 \text{ m d}^{-1}$ (Passow, 1991), which means a particle at
39 the sea surface needs 3 to 16 days to reach the seabed of the EGB. Anoxic conditions in bottom waters of
40 the EGB prevent benthic colonization and bioturbation, and enable the formation of mm-scale laminated
41 sediments in an undisturbed stratigraphic sequence after 1950 CE (Ilus et al., 2007). In contrast to
42 annually or seasonally laminated lake sediments, multiple laminations in sediments **from the central**
43 **Baltic Sea** are not annual but controlled by several factors so the composition of single laminae also
44 varies (Sohlenius et al., 1996; Lepland and Stevens, 1998; Brenner, 2001). Before the mid-20th century,
45 the sediment is grey, homogeneous and silty-clayey. Recent sediments ($<180\text{yr}$) from the EGB are thus
46 well-suited to record precisely and continuously the natural and anthropogenic development of the Baltic
47 Sea and its catchment area. Here, we show that the Baltic Sea represents a high quality candidate to host
48 the onset of the Anthropocene via selection of a Global boundary Stratotype Section and Point (GSSP) for

49 the base of the Anthropocene, that would comprise both a geochronological unit (an Anthropocene epoch)
50 and a chronostratigraphic unit (an Anthropocene series) (Head et al., 2021). The preparatory activities of
51 the Anthropocene Working Group, including events leading to the submission of GSSP proposals and the
52 binding decision that the base of the Anthropocene should align with stratigraphic signals dating to the
53 mid-20th century, are detailed in the introductory article to this special issue (Waters et al., in
54 submission).

55 **Materials and methods**

56 *Geographic settings*

57 The Baltic Sea, the world second largest brackish sea, is a shallow, marginal sea with a mean depth of 55
58 m, and is characterized by an estuarine-like circulation (Elken and Matthäus, 2008). The upper water
59 layer (0 – 60 m) has low salinities (7-8 g kg⁻¹) caused by high freshwater inputs by rivers in the eastern
60 and northern parts of the drainage basin. As a counterpart, relatively high saline water from the North Sea
61 flows in through the Danish Straits. This dense water spreads into deep layers and causes a strong vertical
62 salinity gradient in the central Baltic Sea, with a pronounced halocline around 60–80 m depth. The inflow
63 of saline waters is balanced by upwelling, diapycnal mixing and the outflow of brackish water in surface
64 layers. The catchment area of the Baltic Sea covers 1.7 million km², that is more than four times the area
65 of the Baltic Sea itself (369,000 km²), and is drained by > 200 rivers (Håkanson et al., 2003; Snoeijs-
66 Leijonmalm and Andrén, 2017). About 85 million people live within the Baltic Sea catchment area, and
67 18% of them within 10 km of the coast (HELCOM, 2010). Agriculture land areas are concentrated mainly
68 in the southern part of the drainage basin, where they occupy >65% of the land area (Snoeijs-Leijonmalm
69 and Andrén, 2017).

70 *Field collection, core scanning, and sampling*

71 Sediment core EMB201/7-4 (57°16.980'N, 020°07.228'E; 241 m water depth; Figure 1) was recovered
72 from the EGB on December 6, 2018, during expedition EMB201 on board the research vessel Elisabeth

73 Mann Borgese. The coring site was located in the exclusive economic zone of Latvia. A multicorer,
74 equipped with 60 cm-long polyvinyl chloride tubes, was used as this coring device keeps the water-
75 sediment interface undisturbed. The 45 cm-long core was opened lengthwise on board and kept
76 refrigerated (4°C) in the core repository of the Leibniz Institute for Baltic Sea Research (IOW, Germany).
77 Both core halves were scanned at the IOW with an ITRAX X-ray fluorescence (XRF) scanner (Cox
78 Analytical Systems) equipped with a Cr tube operated at 30 kV and 55mA (exposure time of 5 s) with a
79 200 µm resolution (Table S1). X-ray radiography was recorded at 60 kV and 30 mA, with an exposure
80 time of 225 milliseconds, and at a 200 µm resolution, to identify variations in density along the sediment
81 core. In June 2020, one half of the sediment core was sampled at the IOW in an enclosed laminar flow
82 workstation to prevent airborne contamination. Due to the high water content of recent sediments in the
83 upper part of the core, and to guarantee sufficient amount of sediments for all analyses, the sampling
84 strategy included slicing the core every 1 cm from 0 to 27 cm depth, and every 0.5 cm from 27 to 45 cm
85 depth. This resulted in a total of 63 samples, which were freeze-dried and homogenized before analysis.
86 The second half of the core is kept in the IOW repository as an archive.

87 *Chronological control*

88 An event stratigraphy approach has been developed to accurately date Baltic Sea recent sediments
89 (Dellwig et al., 2018; Kaiser et al., 2020; Dellwig et al., 2021; Lin et al., 2021). This method was applied
90 with modifications in the present study to establish the age model of core EMB201/7-4. Artificial
91 radionuclides ($^{239+240}\text{Pu}$, ^{241}Am , ^{137}Cs) here were not used as time markers, but as anthropocene proxies,
92 except ^{137}Cs for the Chernobyl accident in 1986 (see below).

93 *Anthropocene proxies*

94 Relevant proxies used to define the onset of the Anthropocene epoch follow recommendations from
95 Zalasiewicz et al. (2019).

96 ²⁴¹Am, ¹³⁷Cs, and ²¹⁰Pb. Artificial radionuclides ²¹⁰Am, ¹³⁷Cs, and natural ²¹⁰Pb were analysed at the IOW,
97 Germany, by gamma spectrometry with a Ge-well detector (GCW4021-7500SL-RDC-6-ULB). Raw data
98 were then processed with the GENIE 2000 software (Canberra Industries Inc., USA). Counting statistics
99 were better than 5 – 10% for ²¹⁰Pb and ¹³⁷Cs and better than 20% for ²⁴¹Am activities, respectively. The
100 radionuclide activities were calculated using the following standard reference materials (decay corrected):
101 IAEA-447 (¹³⁷Cs, ²¹⁰Pb, ²²⁶Ra), IAEA-384 (²⁴¹Am) and IAEA-385 (²⁴¹Am, ¹³⁷Cs, ²²⁶Ra). The energies
102 used for quantification of isotopes were 46.5 keV for ²¹⁰Pb, 295 keV and 351 keV for ²²⁶Ra, 59.5 keV for
103 ²⁴¹Am, and 661 keV for ¹³⁷Cs (Moros et al., 2017). A total of 49 samples was analysed, that is all samples
104 between 0 and 31.5 cm and every second sample between 32 and 44.5 cm (Table S2).

105 ²³⁸Pu, ²³⁹Pu, and ²⁴⁰Pu. Pu isotopes were analysed at the GAU-Radioanalytical Laboratories of the
106 University of Southampton, UK. Briefly, sediment samples (n = 20; Table S2) were transferred to glass
107 beakers and ignited at 450°C to remove organic matter. The remaining inorganic fraction was spiked with
108 ²⁴²Pu chemical recovery tracer, mixed with lithium metaborate flux and fused at 1100°C to achieve a
109 homogeneous melt. The obtained liquid glass was then quenched in deionised water and the resulting
110 solidified glass was dissolved in 8M HNO₃ at 90°C (overnight). The resulting solution was cooled to
111 room temperature and the precipitated boric acid, resulting from the added flux, was filtered off using a
112 GF/A filter paper. Pu was pre-concentrated from the sample solution by means of co-precipitation with
113 Fe(OH)₃ at pH 5-6. The precipitate was separated by centrifuging and re-dissolved in 9M HCl. These
114 solutions were transferred to anion exchange columns, previously pre-conditioned with 9M HCl. Pu was
115 retained on the columns and was purified by sequential washes with 9M HCl, 8M HNO₃ and again with
116 9M HCl. Finally, the Pu fraction was eluted to pre-cleaned beakers using 0.1M NH₄I/9M HCl solution.
117 The solution was evaporated to dryness with HNO₃ added to remove excess iodine present and thin alpha
118 spectrometric sources were prepared by electrodeposition. Each source was then counted using Octete
119 (Ortec/Amtek) alpha spectrometers equipped with passivated implanted planar silicon (PIPS) detectors.

120 The resulting spectra were analysed using Maestro spectral analysis software. Following counting,
121 $^{240}\text{Pu}/^{239}\text{Pu}$ ratios were determined using a Thermo Scientific Neptune Plus MC-ICP-MS after
122 radiochemical re-purification of the Pu alpha spectrometry discs (Łokas et al., 2022).
123 *Radiocarbon (^{14}C)*. ^{14}C was analysed by accelerator mass spectrometry (AMS) at the ETH Zürich,
124 Switzerland. Sediment samples (n = 20; Table S2) were selected between 1 and 31 cm depth, where the
125 highest ^{14}C activity was expected. Each sample (0.1 – 0.3 g) was subject to acid-base-acid treatment
126 (Hajdas, 2008; Hajdas et al., 2021). Briefly, in the first acid step bulk sediment was decarbonated with
127 0.5M HCl to remove inorganic carbon. After a washing to neutral with MilliQ water the decarbonated
128 sediment was treated with 0.1M NaOH. The liquid was extracted and a few ml of 0.5 M HCl were added
129 to precipitate the base-extracted organic fraction (or humic acid fraction), and freeze dried. The residue,
130 i.e., the base-insoluble organic fraction (or humin fraction), was washed to neutral, treated with 0.5M
131 HCl, washed again to neutral, and freeze-dried. Both fractions were weighed into Al cups for combustion
132 in an elemental analyser (Vario Micro Cube; Elementar) and subsequent graphitization (Nemec et al.,
133 2010). The graphite powder was pressed into the Al cathodes to measure $^{14}\text{C}/^{12}\text{C}$ and $^{13}\text{C}/^{12}\text{C}$ ratios using
134 the AMS system of MICADAS (Synal et al., 2000; Wacker et al., 2010). Results are reported as fraction
135 modern F^{14}C (Reimer et al., 2004) and radiocarbon ages (Before Present, BP, where 0 BP = 1950 CE;
136 (Stuiver and Polach, 1977). For an overview of the method and data reporting see Hajdas et al. (2021).
137 *Fly-ash*. Spheroidal carbonaceous fly-ash particles (SCPs) were analysed at University College London,
138 UK (Rose, 1994). Briefly, sediments (0.2 g; n = 52; Table S2) were subjected to sequential chemical
139 attack by mineral acids to remove unwanted fractions leaving a suspension of mainly carbonaceous
140 material in water. A known fraction of this final suspension was evaporated onto a coverslip and mounted
141 onto a microscope slide. SCPs were then counted using a light microscope at x450 magnification and the
142 sediment content calculated in units of SCPs per gram of sediment dry weight (SCPs g^{-1} dw). The criteria
143 for SCP identification under the light microscope followed (Rose, 2008). The detection limit for the

144 technique is typically ca. 100 SCPs g⁻¹ dw and accuracy is ca. ± 45 SCPs g⁻¹ dw (Rose, 2008) although
145 detection limits were slightly higher in this core (ca. 300 SCPs g⁻¹ dw). SCPs were sub-divided into size
146 classes during counting. Analytical blanks and SCP reference material were included in duplicate and
147 triplicate, respectively. Reference materials agreed closely with expected values (6005 ± 70 SCPs g⁻¹ dw),
148 while no SCPs were observed in blanks. SCPs were not analysed between 2 and 9 cm due to the low
149 amount of available sediment.

150 *Microplastics.* Microplastics (MPs) sampling strategy aimed to maximize MP particles recovery by
151 merging consecutive layers in the core to reach a minimum of ca. 1 g sediment dw, resulting in a total of
152 20 samples (Table S2). Potential MPs (particle size between 11µm and 5mm) were extracted and isolated
153 from sediments at the IOW, Germany, following a well-established protocol (Enders et al., 2020). Briefly,
154 sediments were submitted to a density separation by using sodium polytungstate (1800 kg m³), which
155 allows all plastic particles to float, and digested with 30% H₂O₂ to eliminate organic particles. The
156 method recovery rate has been estimated to 80%. Automated particle-based analysis was performed at the
157 Biological Institute Helgoland (BAH) of the Alfred Wegner Institute, Helmholtz Centre for Polar and
158 Marine Research, Germany, following a well-established protocol (Primpke et al., 2018). Purified
159 samples were submitted to a second density separation and digestion steps (Abel et al., 2021) to
160 disaggregate flocculates and eliminate potential biofilm immediately before micro-Fourier transform
161 infrared spectroscopic analysis (µFTIR; Hyperion 3000m, Bunker Optics GmbH). This technique
162 measured 100% of remaining particles in each sample to ensure all potential MP particles could be
163 associated to a specific polymer type, or excluded from being classified as plastic. Both extraction and
164 spectroscopic analysis followed protocols to prevent contamination during all steps, including processing
165 blanks to allow estimation of false-positive MP particles that were excluded from final results after
166 interpretation. Contamination from the core tube was not observed.

167 *Persistent organic pollutants and biomarkers.* Polycyclic aromatic hydrocarbons (PAHs),
168 dichlorodiphenyltrichloroethane (DDT) and biomarkers (sterols) were analysed at the IOW, Germany
169 (Kanwischer et al., 2020; Kaiser and Lerch, 2022). Briefly, homogenized sediments (0.3 – 0.8 g dw; n =
170 62; Table S2) were extracted with a DCM:MeOH mixture using an Accelerated Solvent Extraction device
171 (Dionex™ASE™ 350; Thermo Fisher Scientific). After the addition of internal standards, the extracts
172 were desulfurized with activated copper and separated into four fractions by microscale flash column
173 chromatography using hexane (F1), hexane:DCM (F2), DCM (F3) and DCM:MeOH (F4) as eluting
174 solvents (Kaiser and Arz, 2016). The F1 and derivatized F4 fractions were analysed by gas
175 chromatography - mass spectrometry (GC-MS; Agilent Technologies 7890 B GC system and 5977 B
176 Mass Selective Detector). PAHs, p,p'-DDD (1,1-dichloro-2,2-bis(p-chlorophenyl)ethane) and p,p'-DDE
177 (1,1-dichloro-2,2-bis(chlorophenyl)ethylene) were identified by comparing retention time with certified
178 external standards (DRE-L20950009CY and DRE-L18000014CY). PAHs and DDT, considered here as
179 the sum of its two degradation products, p,p'-DDD and p,p'-DDE, were quantified using calibration
180 curves based on six calibrator points and expressed as Σ PAH and Σ DDT. Brassicasterol (24-
181 methylcholesta-5,22E-dien-3 β -ol), tetrahymanol (gammaceran-3 β -ol), coprostanol (5 β -cholestan-3 β -ol),
182 epi-coprostanol, 5 β -stigmastanol (24 β -ethyl-5 β -cholestan-3 β -ol), and epi-5 β -stigmastanol were identified
183 by chromatographic and mass spectral comparison with published data. The analytical recovery was
184 estimated to 90% and the limit of quantification to 5 ng g⁻¹ sediment dw. Results were normalized to the
185 total organic carbon content for each sample (μ g gTOC⁻¹).

186 *Total carbon (TC), total organic carbon (TOC), total nitrogen (TN), and C/N ratio.* Bulk sediment TC,
187 TOC and TN were measured at the IOW, Germany (Müller and Voss, 1999; Voss et al., 2000). The
188 samples (30 mg dw; n = 62; Table S2) were analysed with a Carlo Erba Flash EA elemental analyser.
189 TOC and TN were measured by weighing the sediment into Ag boats and treating it with 2N HCL until
190 all carbonate compounds had reacted out. The sample was dried at 60°C, pelletized and measured. The

191 calibration material for C and N analysis was acetanilide (Merck). The analytical precision was better
192 than 1% for TC and TOC, and better than 4% for TN. The C/N ratio was defined as the molar ratio
193 between TOC and TN: $C/N = (TOC/12) / (TN/14)$. Data were not corrected for salt content due to low
194 salinity of the Baltic Sea (Snoeijs-Leijonmalm and Andr en, 2017).
195 *Stable carbon ($\delta^{13}C$) and nitrogen ($\delta^{15}N$) isotopes.* Bulk $\delta^{15}N$ and $\delta^{13}C$ were analysed at the IOW,
196 Germany. Briefly, untreated and treated (acidified) fractions of sediments (n = 62; Table S2) were
197 analysed following M uller and Voss (1999) and Voss et al. (2000). The isotopic compositions were
198 determined using flash combustion in a EA Isolink CN (Thermo Fisher Scientific) elemental analyser at
199 1020 C coupled via a ConFlo IV (Thermo Fisher Scientific) split interface to a Delta+ V advantage
200 (Thermo Fisher Scientific) isotope ratio mass spectrometer. The isotopic values are reported relative to
201 atmospheric N₂ ($\delta^{15}N$) and Vienna PeeDee Belemnite (VPDB; $\delta^{13}C$). The reference materials used for
202 stable isotope analysis were IAEA-N1, IAEA-N2, IAEA-N3, NBS 22, IAEA-CH-3 and IAEA-CH-6. The
203 analytical precision for both stable isotope ratios was <0.2‰. To correct for the Suess effect, that is a
204 change in $\delta^{13}C$ by the admixture of large amounts of fossil-fuel derived CO₂, the following equation
205 (Schelske and Hodell, 1995; Verburg, 2006) was applied:

$$206 \quad \delta^{13}C_{org,corr} = \delta^{13}C_{org} - (4,577.8 - 7.3430 * y + 3.9213 * 10^{-3} * y^2 - 6.9812 * 10^{-7} * y^3)$$

207 where y = year CE.

208 *Inorganic geochemical markers.* Sedimentary major and trace element contents and stable Pb isotopes
209 were analysed at the IOW, Germany. Briefly, the sediments (100 mg dw; n = 62; Table S2) were digested
210 with a HNO₃-HClO₄-HF mixture in closed Teflon vessels at 180 C for 12 h and the residues were fumed-
211 off 3-times with 6 M HCl and finally diluted with 2 vol% HNO₃ (Dellwig et al., 2019). The content of Al
212 was measured by inductively coupled plasma optical emission spectrometry (ICP-OES, iCAP 7400 Duo,
213 Thermo Fisher Scientific) using external calibration and Sc as internal standard. Analytical precision and
214 accuracy were estimated with the international reference material SGR-1b (USGS) and were better than

215 3.8% and -3.9%, respectively. Trace metals Cu, Pb (including ^{206}Pb and ^{207}Pb) and Zn were determined
216 by inductively coupled plasma mass spectrometry (ICP-MS, iCAP Q, Thermo Fisher Scientific) coupled
217 to a prepFAST system (Elemental Scientific) using He as collision gas, external calibration, and Rh and Ir
218 as internal standards. Analytical precision and accuracy of concentration measurements were also
219 checked with SGR-1b and were better than 3.3% and -3.4%, respectively. For stable Pb isotope ratios,
220 NIST SRM-981 was used resulting in a precision and accuracy of 0.31% and -0.04%, respectively
221 (Dellwig et al., 2021). Results were normalized to Al content for each sample to minimize dilution effects
222 caused by organic matter, carbonate and/or opal. To determine Hg contents, sediments (100 mg dw; n =
223 63; Table S2) were analysed with a Milestone DMA-80 EVO Direct Mercury Analyser by thermal
224 decomposition at 750 °C, gold amalgamation and atomic absorption spectrophotometry at 253.65 nm.
225 Data were calibrated against the BCR 142R certified reference material and a Baltic Sea reference
226 sediment. Analytical precision based on reference materials was better than 6%.
227 *Numerical analysis.* To detect positions of abrupt change (changepoints) within the different datasets, a
228 modelling approach (Gallagher et al., 2011) was applied using the software Past v4.10 (Hammer et al.,
229 2001). Before analysis, the datasets were interpolated every 1 cm, missing values were treated by linear
230 interpolation, and mean and standard deviation were automatically normalized in order to equally weigh
231 each dataset. The default value of 100,000 iterations (Markov chain Monte Carlo) was used and the
232 number of changepoint set to n = 1 (Table S3).

233 **Results**

234 *Lithology*

235 The lithology of core EMB201/7-4 was characterized by a pronounced visual transition at around 27 cm
236 sediment depth (Figure 2). Between 45 – 27 cm depth the sediment was grey, relatively homogeneous,
237 silty-clay, with low organic carbon and water contents. Between 27 – 0 cm depth it was brown to dark
238 brown and black, cm-thick layered and sometimes laminated, with high organic carbon and water

239 contents. Note that, after oxidation, sediment colours change to dark grey below 27 cm and to brown to
240 dark brown between 27 – 0 cm. The upper part of the core was also characterized by a few light grey,
241 mm-wide, clayey to silty laminations mainly between 21 – 25 cm depth.

242 *Chronology*

243 The use of unsupported ^{210}Pb ($^{210}\text{Pb}_{\text{unSUPP}}$) to date recent sediments is challenging in the Baltic Sea because
244 the depositional environment and the sedimentation type have varied over the past 100 years, resulting in
245 an irregular decrease in $^{210}\text{Pb}_{\text{unSUPP}}$ (Mattila et al., 2006; Moros et al., 2017; Häusler et al., 2018). Seven
246 parameters were considered to build the stratigraphy ($^{206/207}\text{Pb}$, Hg, ΣPAH , ΣDDT , Mn/Ti, Co/Al and
247 ^{137}Cs ; Figure 2b-h), and twelve stratigraphical events, or time markers, were used to date the sediment
248 (Table 1). The uncertainties of the time markers were estimated following Moros et al. (2017), Kaiser et
249 al. (2018, 2020), and Lin et al. (2021). The early increases in $^{206/207}\text{Pb}$, Hg and ΣPAH reflect the early
250 increase in wood and coal combustion around 1870 CE related to the beginning of the Second Industrial
251 Revolution, or Technological Revolution (Hites Ronald et al., 1977; Elmquist et al., 2007; Díaz-Somoano
252 et al. 2009; Hanke et al., 2019). The abrupt rise in ΣDDT is likely related to increased DDT use for
253 agriculture since 1950 CE (Li and Macdonald, 2005), a well-known pattern in Baltic Sea and European
254 lake sediments (Thevenon et al., 2013; Sabatier et al., 2014). Mn and Co enrichments are attributed to
255 four MBIs (1978, 1994, 2003 and 2014 CE (Dellwig et al., 2021; Lin et al., 2021), and tentatively to other
256 four MBIs (1964, 1967, 1970 and 1972 CE; Figure S1). The ^{137}Cs increase has been attributed to the
257 Chernobyl accident in 1986 CE. The year of the core recovery (2018 CE) was attributed to the core top
258 layer. Linear sedimentation rates were assumed between time markers resulting in a mean sedimentation
259 rate increasing from 0.2 cm yr^{-1} between 45 cm (ascribed the date $1840 \pm 10 \text{ CE}$) and 28.4 cm depth
260 ($1950 \pm 4 \text{ CE}$) to 0.4 cm yr^{-1} from 28.4 cm to the top of the core (Figure 2g).

261 *Radioisotopes (actinides and ^{14}C)*

262 ^{241}Am and $^{239+240}\text{Pu}$ profiles showed similar patterns (Figure 3a-b). Both radionuclides were close to the
263 detection limit below 28.8 cm depth (1947 ± 4 CE). They started to increase at 27.3 cm depth (1953 ± 4
264 CE) and reached a maximum at 24.5 cm depth (1963 ± 4 CE) with activities of 8.4 Bq kg^{-1} for ^{241}Am and
265 0.025 Bq kg^{-1} for $^{239+240}\text{Pu}$. Above this depth, the activities of both radionuclides decreased abruptly. A
266 second, sharp peak occurred at 15.5 cm depth (1986 ± 2 CE) in both radionuclide profiles. Values of
267 $^{240/239}\text{Pu}$ (Figure 3c), a tracer for Pu source, were >0.301 at 28.8 and 28.3 cm depth (1947 ± 4 and $1950 \pm$
268 4 CE, respectively) and between $0.160 - 0.219$ from 27.8 cm depth (1952 ± 4 CE) except for slightly
269 higher values at 15.5 cm (1986 ± 2 CE) and 10.5 cm depth (1998 ± 2 CE). F^{14}C (Figure 3d), the
270 normalized fraction of ^{14}C (Hajdas et al., 2021), was relatively stable (~ 0.8) between 30.8 and 27.3 cm
271 depth ($1933 - 1953$ CE). It started increasing at 26.5 cm depth (1956 ± 4 CE), reached a maximum of 1.2
272 between 23.5 and 20.5 cm ($1966 - 1974$ CE), and decreased to 1.0 at 1.5 cm depth (2014 ± 1 CE).

273 *Spheroidal carbonaceous particles.* A first appearance of SCPs (Figure 4a) occurred at 36.3 cm depth
274 (1897 ± 8 CE), but SCP contents remained low and sporadic until 27.3 cm depth (1953 ± 4 CE). Only
275 SCPs $<25 \mu\text{m}$ (and frequently $<10 \mu\text{m}$) were observed in this lower section. Above 27.3 cm depth, SCP
276 contents increased, initially rapidly, through to a peak of $>7,700 \text{ SCPs g}^{-1} \text{ dw}$ at 16.5 cm depth (1983 ± 3
277 CE). From this peak, SCP contents showed a general decline through to the top of the core. Above 28.3
278 cm, larger SCPs appeared for the first time and SCPs had a wide size range from $<5 \mu\text{m}$ to $>50 \mu\text{m}$ (Table
279 S2).

280 *Microplastics.* A total of 178 MP particles and six different polymers were identified. Most MP particles
281 were smaller than $100 \mu\text{m}$ in size, and fibres were almost absent. MP amounts (Figure 4b) ranged
282 between 248 and $7,135 \text{ MPs kg}^{-1} \text{ dw}$. MPs were identified at all depths between 42.3 and 6.5 cm ($1857 -$
283 2005 CE), with a maximum abundance at 20.5 cm depth (1974 ± 4 CE). In terms of polymer diversity, a
284 near-consistent pattern was observed within all layers with predominance of ethylene-propylene-diene
285 monomer rubbers (Table S2).

286 *Polycyclic aromatic hydrocarbons and dichlorodiphenyltrichloroethane.* The most abundant PAHs were
287 phenanthrene, fluoranthene, pyrene, benzo[a]anthracene, chrysene, benzo[b]fluoranthene,
288 benzo[k]fluoranthene, benzo[a]pyrene, indeno[1,2,3-cd]pyrene, dibenzo[ah]anthracene and
289 benzo[ghi]perylene. Their sum (Σ PAH) increased gradually from 41.3 to 17.5 cm depth (1863 – 1981
290 CE), reaching a maximum around $170 \mu\text{g gTOC}^{-1}$, and decreased gradually to reach a minimum of $20 \mu\text{g}$
291 gTOC^{-1} near the top of the core (Figure 4c). Σ DDT (Figure 4d) was below detection limit between 44.8
292 and 28.8 cm depth (1840 – 1947 CE). It increased rapidly at 28.3 cm depth (1950 ± 4 CE) to reach a
293 maximum of $20 \mu\text{g gTOC}^{-1}$ between 25.5 and 21.5 cm depth (1960 – 1971 CE). After the peak, Σ DDT
294 decreased gradually towards the top of the core to reach contents $<1 \mu\text{g gTOC}^{-1}$.

295 *Total carbon, total organic carbon, total nitrogen, total organic nitrogen, and the C/N ratio.* Both TC and
296 TOC values (Figure 5a-b) were relatively constant around 3-4% between 44.8 and 29.8 cm depth (1840 –
297 1940 CE). Above 29.8 cm depth (1940 ± 5 CE), they increased gradually to reach 19.8 and 17.3% at the
298 top of the core, respectively. TC and TOC showed a significant positive linear correlation ($r^2 = 0.97$). TN
299 values were $<0.5\%$ below 27.8 cm depth (1952 ± 4 CE), followed by a gradual increase to the top of the
300 core, where the values reached 2% (Figure 5c). The C/N ratio (Figure 5d) ranged between 10 and 12
301 below 27.8 cm depth (1952 ± 4 CE), and between 8 – 10 above 23.5 cm depth (1966 ± 4 CE).

302 *Stable carbon and nitrogen isotopes.* The record of bulk organic matter $\delta^{13}\text{C}$ ($\delta^{13}\text{C}_{\text{org}}$) presented a -1.2‰
303 shift from $-26.2 \pm 0.2\text{‰}$ below 27.8 cm depth (1952 ± 4 CE) to $-25.0 \pm 0.3\text{‰}$ above this depth (Figure
304 5e). Both $\delta^{13}\text{C}_{\text{org}}$ and $\delta^{13}\text{C}_{\text{org,corr}}$ vertical profiles (Figure 5e) were relatively similar, except above 11.5 cm
305 depth (1996 ± 2 CE), where $\delta^{13}\text{C}_{\text{org}}$ values decreased gradually compared to $\delta^{13}\text{C}_{\text{org,corr}}$. $\delta^{15}\text{N}$ values
306 presented a relatively large scatter (Figure 5f). A general increasing trend of $+2\text{‰}$ occurred between 44.8
307 – 27.8 cm depth (1840 – 1952 CE). Above 26.5 cm depth (1956 ± 4 CE), $\delta^{15}\text{N}$ values fluctuated around
308 $3.3 \pm 0.4\text{‰}$.

309 *Inorganic geochemical signals.* Hg/Al, Cu/Al, Zn/Al, and Pb/Al values (Figure 6a-d) increased from 28.8
310 cm depth (1947 ± 4 CE). The ratios were highest between 24 – 18 cm depth (1966 – 1978 CE), and
311 decreased until 16.5 cm depth (1983 ± CE). Cu/Al and Zn/Al increased again above 16.5 cm depth (1983
312 ± 3 CE). ^{206/207}Pb values (Figure 6e) decreased gradually from 1.253 to 1.164 between 38.8 – 22.5 cm
313 depth (1873 – 1969 CE) with a slight increase from 33.3 to 31.8 cm depth (1917 – 1927 CE). Above 22.5
314 cm depth (1969 ± 4 CE), ^{206/207}Pb values increased and reached a value of 1.209 at the top of the core.

315 *Biotic markers.* Brassicasterol, contents increased significantly from 1.5 to 8.5 µg gTOC⁻¹ between 26.5
316 cm depth (1956 ± 4 CE) and the top of the core (Figure 7a). Tetrahymanol contents (Figure 7b) showed a
317 similar trend with very low values below 26.5 cm depth (1956 ± 4 CE) followed by an increase until 10.5
318 cm depth (1998 ± 2 CE). Above this depth, the contents were relatively stable until the top of the core.

319 Faecal lipids (the sum of coprostanol and epi-coprostanol) first appeared at 26.5 cm depth (1956 ± 4 CE)
320 and increased until the top of the core (Figure 7c). The R2 ratio (R2 = coprostanol + epicoprostanol / 5β-
321 stigmastanol + epi-5β-stigmastanol + coprostanol + epicoprostanol; Kaiser and Lerch 2022) ranged
322 between 0.4 – 0.8 and increased from 26.5 cm (1956 ± 4 CE) to the top of the core (Figure 7d; Table S2).

323 **Discussion**

324 *Chronology.* The age model of core EMB201/7-4 (Figure 2) is well-constrained by eleven stratigraphic
325 events between 1950 – 2018 CE (28.3 – 0 cm depth). Most of these stratigraphic events are known in the
326 Baltic Sea and have already been used to date recent sediments (Moros et al., 2017; Häusler et al., 2018;
327 Kaiser et al., 2020; Dellwig et al., 2021; Lin et al., 2021). Before 1950 ± 4 CE (28.3 cm depth), however,
328 the age model is based on a single stratigraphic event at 1870 ± 10 CE (40.3 cm depth) and on a linear
329 extrapolation to the bottom of the core dated 1840 ± 10 CE (45 cm depth). Between 1870 – 1950 CE,
330 older-than-expected ¹⁴C ages (1892 ± 80 year BP; n = 7) were obtained (Table S2). This indicates an
331 admixture of relatively older organic matter most likely due to the inwash of pre-aged terrestrial organic
332 matter from the drainage basin (Howarth et al., 2013; Schroeter et al., 2021), or sediment reworking

333 (Moros et al., 2020). The $^{206/207}\text{Pb}$ and ΣPAH profiles from core EMB201/7-4 (Figure 2b,d) support the
334 absence of sediment reworking as they are comparable to existing records from lakes within the Baltic
335 Sea drainage basin (Renberg et al., 2001, 2002). Therefore, the inwash of pre-aged terrestrial organic
336 matter likely explains older-than-expected radiocarbon ages, but the sediments were not affected
337 significantly by reworking before 1950 CE. The event stratigraphy approach followed here represents the
338 most advanced method to date Baltic Sea recent sediments.

339 *Radionuclides as primary markers for the Anthropocene GSSP level in Baltic Sea sediments.* For the
340 Anthropocene to be a formal chronostratigraphic/geochronologic unit of geological time, a synchronous
341 base is needed (Head et al., 2021; Waters et al., in submission). $^{239+240}\text{Pu}$ and ^{241}Am artificial
342 radionuclides produced by atmospheric nuclear bomb tests are the main candidates as primary markers in
343 the context of the Anthropocene because of their longevity (half-lives of 24,110 and 432 years,
344 respectively) and appearance around 1950 CE in different archives worldwide (Zalasiewicz et al., 2017;
345 Waters et al., 2018; Waters et al., 2019). While $^{239+240}\text{Pu}$ was first identified above detection limit in 1947
346 ± 4 CE (28.8 cm depth) in core EMB201/7-4, both ^{241}Am and $^{239+240}\text{Pu}$ (Figure 8a-b) started to increase
347 markedly after 1953 ± 4 CE (27.3 cm depth). They reached maximum values in 1963 ± 4 CE (24.5 cm
348 depth), when the Limited Test-Ban Treaty was established, which was followed by a rapid decline in
349 radionuclide fallout worldwide (Waters et al., 2018; Waters et al., 2019). $^{240/239}\text{Pu}$ values (Figure 3c)
350 above 0.30 between 1947 – 1950 CE (28.8 – 28.3 cm depth) suggested a main radionuclide source from
351 test sites in the Pacific Ocean (Pacific Proving Grounds; (Koide et al., 1985; Buesseler, 1997; Muramatsu
352 et al., 2000; Lachner et al., 2010), while values below 0.19 since 1953 ± 4 CE (27.3 cm depth) are typical
353 of global fallout (Buesseler and Sholkovitz, 1987; Kelley et al., 1999). Slightly increased values in 1986 \pm
354 2 CE (15.5 cm depth) are consistent with the Chernobyl accident (Muramatsu et al., 2000).

355 The ^{137}Cs artificial radionuclide (Figure 2h) activity increased slightly from 1956 ± 4 CE (26.5 cm depth)
356 due to atmospheric nuclear weapon tests, but the most prominent increase occurred after 1981 ± 3 CE

357 (17.5 cm depth) because of the Chernobyl accident in 1986 CE. As documented, the radioactive cloud
358 from this accident first travelled northward and resulted in a high deposition of ^{137}Cs radionuclides in the
359 Baltic Sea region (Mattila et al., 2006; Ilus et al., 2007; Moros et al., 2017; Dellwig et al., 2018). F^{14}C , the
360 atmospheric radiocarbon excess released by nuclear bomb tests, started to increase in 1956 ± 4 CE (26.5
361 cm depth) and reached maximum values in 1966 ± 4 CE (23.5 cm depth), about three years after ^{241}Am
362 and $^{239+240}\text{Pu}$ in both cases (Figure 3d). This delay can be attributed to atmosphere-ocean ^{14}C equilibration
363 time (Waters et al., 2019). Also, post-depositional, vertical migration within sediments, sediment
364 remobilization, as well as sediment inwash from the drainage basin are typical potential issues affecting
365 radionuclide profiles (Davis et al., 1984; He et al., 1996; Owens and Walling, 1996; Mattila et al., 2006;
366 Ilus et al., 2007; Hancock et al., 2014; Waters et al., 2019; Haltia et al., 2021; Schroeter et al., 2021).
367 These issues likely explain the detection of ^{137}Cs since 1917 ± 6 CE (33.3 cm depth) because it is very
368 mobile in sediments compared to ^{241}Am , as well as the detection of radionuclides in the sediments until
369 the present-day.

370 Artificial $^{239+240}\text{Pu}$ and ^{241}Am radionuclides from nuclear weapons testing are therefore considered here as
371 the most relevant primary markers in the context of the Holocene-Anthropocene boundary. Results from a
372 changepoint model (Figure 8c; Table S3) indicated a change starting in 1948 ± 4 CE (28.5 cm depth), but
373 the main change occurred in 1956 ± 4 CE (26.5 cm depth), which would be the proposed GSSP level in
374 Baltic Sea sediments.

375 *Anthropogenic pollutants as secondary markers for the Anthropocene GSSP level in Baltic Sea sediments.*
376 SCP contents (Figure 8d) increased markedly after 1953 ± 4 CE (27.3 cm depth) due to an increase in
377 coal and fuel oil combustion related to a high electricity demand after WWII. Furthermore, the
378 introduction of new SCP sources is suggested by their wider range of sizes since that time (Table S2).
379 Decreasing SCP contents since 1983 ± 3 CE (16.5 cm depth) reflects the reduction in SCPs release to the
380 atmosphere due to the introduction of particle-arrestor technology since the 1970s and the shift away from

381 coal and oil fuels to natural gas, nuclear and renewable energy sources (Rose, 2015; Rose and Gałuszka,
382 2019). The Baltic Sea SCP profile is compatible with other lake sediments reported globally (Rose,
383 2015).

384 MP amounts (Figure 4b) increased between 1960 – 1974 CE (25.5 – 20.5 cm depth) as a result of a rapid
385 and intense increase in both production and use of an array of polymer types (Zalasiewicz et al., 2016;
386 Ivar do Sul and Labrenz, 2020). Ethylene-propylene-diene monomer rubbers, the prevalent category
387 within polymers identified here (Table S2), are widely used for a range of marine-and terrestrial-based
388 activities and very common in the marine environment (Primpke et al., 2018; Lorenz et al., 2019). The
389 MP profile suggests that the MP stock in the Baltic Sea marine environment consists of polymers from
390 various provenances affected by different taphonomic processes, and hence does not reflect a direct or
391 straightforward response to world plastic production (Bancone et al., 2020).

392 Hg, Pb, Cu and Zn are markers of anthropogenic industrial pollution in sediments (Rauch and Pacyna,
393 2009; Birch et al., 2015; Gałuszka and Rose, 2019). The main trends in Hg/Al, Cu/Al, Zn/Al and Pb/Al
394 (Figure 6a-d) reflect an increase of coal combustion and metal production after the mid-20th century, and
395 a decrease from 1970s CE due to a reduction in industrial dust emission and fossil fuel combustion in
396 northern and western Europe (Rühling and Tyler, 2001; Callender, 2014; Gałuszka and Wagreeich, 2019).
397 However, a decrease to background levels in present-day sediments is not observed for most of the
398 metals. This is because trace metals are generally sensitive to changing redox conditions (Bruland et al.,
399 2013), which biases the anthropogenic signal in Baltic Sea sediments. ^{206/207}Pb is not redox-sensitive, and
400 reflects therefore an unbiased signal (Figure 6e). Decreasing ^{206/207}Pb values between the early-1870s CE
401 and the late-1960s CE are most likely due to both increasing coal combustion during the Second
402 Industrial Revolution and the spread of leaded gasoline combustion in Europe since 1950 CE (Renberg et
403 al., 2002; Díaz-Somoano et al. 2009; Boyle et al., 2014). Increasing values since the early-1970s (21.5 cm

404 depth) reflect the decline of lead pollution in Europe due to both the implementation of stricter industrial
405 emission standards and the phase-out of alkyl-lead gasoline (Renberg et al., 2001; Renberg et al., 2002).
406 The appearance of the persistent organic pollutant DDT (Figure 8e) in 1950 ± 4 CE (28.3 cm depth) and
407 its pronounced increase up to 1960 ± 4 CE (25.5 cm depth) is related to the increasing use of DDT in
408 Baltic Sea countries after 1945 CE to control insects on agricultural crops. Indeed, the former Soviet
409 Union had the second highest (after the United States) historical DDT usage in agriculture from 1952
410 to 1971 CE (Li and Macdonald, 2005). The decrease in DDT content after 1971 ± 4 CE (21.5 cm depth)
411 reflects the ban on its use within Baltic Sea countries since the early-1970s (HELCOM, 2010).
412 Coprostanol and epicoprostanol faecal lipids (Figure 7c-d) first appeared in 1956 ± 4 CE (26.5 cm depth)
413 and increased subsequently as a result of increasing sewage pollution in the central Baltic Sea (Kaiser and
414 Lerch, 2022). The fraction of human-derived faecal lipids was also increasing since 1956 ± 4 CE as
415 suggested by higher R2 values. Therefore, sedimentary faecal lipids reflect the increasing human-derived
416 sewage pollution related to population growth in the Baltic Sea drainage basin (Kaiser and Lerch, 2022).
417 For anthropogenic pollutants, results from the changepoint model (Table S3) indicated a change starting
418 in 1956 ± 4 CE (26.5 cm depth). This change is, however, more pronounced in 1960 ± 4 CE (25.5 cm
419 depth), which corresponds closely to the inflection point of many indicators of the Great Acceleration
420 (Head et al., 2021).
421 *Organic proxies as secondary markers for the Anthropocene GSSP level in Baltic Sea sediments.* The
422 pronounced increase in TOC values (Figure 8f) since 1956 ± 4 CE (26.5 cm depth) reflects the growing
423 burial of phytoplankton biomass under hypoxic conditions due to the eutrophication of the central Baltic
424 Sea deep basins since 1950 CE (Emeis et al., 2000; Struck et al., 2000). Indeed, a shift towards heavier
425 $\delta^{13}\text{C}_{\text{org,corr}}$ values (Figure 8g) after 1952 ± 4 CE (27.8 cm depth) indicates a higher proportion of
426 phytoplankton biomass, as organisms are forced to use the heavier C isotope (^{13}C) when primary
427 production is high (Mizutani and Wada, 1982; Struck et al., 2000; Teranes and Bernasconi, 2005). A

428 decrease in C/N values (Figure 5e) after 1960 \pm 4 CE (25.5 cm depth) also indicates an increasing fraction
429 of phytoplankton biomass in Baltic Sea sediments (Emeis et al., 2000; Lamb et al., 2006). The contents of
430 brassicasterol (Figure 7a), or “diatomsterol”, a biomarker for the occurrence of diatoms, increased
431 substantially after 1952 \pm 4 CE (27.8 cm depth), also indicating the growth of phytoplankton biomass
432 mainly due to eutrophication. The similar increase in tetrahymanol contents (Figure 7b) reflects the
433 concomitant development of a pelagic redoxcline in the central Baltic Sea as tetrahymanol is a biomarker
434 for ciliates thriving abundantly near the redoxclines of hypoxic environments (Harvey and Mcmanus,
435 1991; Anderson et al., 2012; Edgcomb and Pachiadaki, 2014; Pachiadaki et al., 2014).

436 An increasing trend in $\delta^{15}\text{N}$ values (Figure 5g) between 1840 – 1950 CE (44.8 – 28.3 cm depth) resulted
437 most likely from the phytoplankton assimilation of land-derived, isotopically heavier nitrate inputs
438 (Struck et al., 2000; Renberg et al., 2001). Indeed, large-scale changes in land use occurred in the Baltic
439 Sea watershed after the beginning of the Second Industrial Revolution (Zillén and Conley, 2010). After
440 1950 CE, however, $\delta^{15}\text{N}$ values started to decrease slightly, possibly due to higher inputs of isotopically
441 lighter nitrogen, and increasing biomass of diazotrophic cyanobacteria, an important contributor to the
442 pelagic N-cycle of the central Baltic Sea since 1950 CE (Voß et al., 1997; Struck et al., 1998; Struck et
443 al., 2000; Kendall et al., 2007; Kaiser et al., 2020). Periods with increasing $\delta^{15}\text{N}$ values after 1950 CE
444 may be related to higher inputs of isotopically heavier, wastewater to the Baltic Sea (Kendall et al., 2007).

445 As shown for anthropogenic pollution markers, the results from the changepoint model (Table S3)
446 suggested that the main change in organic proxies occurred in 1960 \pm 4 CE (25.5 cm depth),
447 synchronously with many indicators of the Great Acceleration (Syvitski et al., 2020; Head et al., 2021).

448 *Final remarks regarding the Baltic Sea as GSSP for the Anthropocene series.* The present study supports
449 the Anthropocene as a stratigraphically substantiated unit, that should be defined by a GSSP and
450 formalized at the rank of epoch with an inception at around 1950 CE (Zalasiewicz et al., 2017). Artificial
451 radioisotopes released during atmospheric nuclear weapons testing, which are the most significant

452 markers in the context the Anthropocene, showed a main change in core EMB201/7-4 in 1956 ± 4 CE as
453 indicated by the changepoint model. This point is therefore proposed as the GSSP level for the
454 Anthropocene series (Figure 8). It corresponds to a marked change in lithology and sediment colour from
455 light grey to dark brown (or dark grey to brown after oxidation), which reflects the beginning of
456 eutrophication in the central Baltic Sea due to an increasing anthropogenic pressure. Despite the most
457 accurate approach used to date core EMB201/7-4, the lack of annual lamination may flag the Baltic Sea
458 as a GSSP candidate for the Anthropocene series.

459 If accepted, the East Gotland Basin GSSP would be in accordance to all 65 GSSPs presently ratified,
460 which occur within strata deposited in marine sedimentary environments with the exception of the three
461 Holocene GSSPs (Walker et al., 2018); Waters et al., 2019). The archived half of core EMB201/7-4 is
462 preserved and accessible at the IOW, Germany. Although marine GSSP sites are not as accessible as
463 terrestrial ones, new sediment cores from exactly the same location can be recovered during expeditions
464 on board research vessels. Finally, if appropriate, the name “Baltician” is suggested as potential stage/age
465 (Luciano, 2022).

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732

733 **Figure and Table captions**

734 **Figure 1.** (a) Study area (black rectangle) in the Baltic Sea; the countries, main lakes and rivers of the
735 catchment area - delimited by a bold line - are shown; (b) bathymetric chart (Baltic Sea Bathymetry
736 Database) of the study area with the location of core EMB201/7-4 in the East Gotland Basin (EGB).

737 **Figure 2.** Pictures (true colour and radiography), depth profiles of $^{210}\text{Pb}_{\text{unsupp}}$ (a), $^{206/207}\text{Pb}$ (b), Hg (c),
738 ΣPAH (d) and ΣDDT (e) contents, Mn/Ti (f), Co/Al (g) and ^{137}Cs (h), and age-depth relationship (i) of
739 core EMB201/7-4. The assigned time markers are shown by horizontal grey lines and listed in Table 1.
740 **Figure 3.** Depth profiles of ^{241}Am (a) and $^{239+240}\text{Pu}$ (b) activities, $^{240/239}\text{Pu}$ (c), and F^{14}C (d) in core
741 EMB201/7-4.
742 **Figure 4.** Depth profiles of SCPs (a), microplastics (b), ΣPAH (c) and ΣDDT (d) contents in core
743 EMB201/7-4.
744 **Figure 5.** Depth profiles of TC (a), TOC (b), and TN (c) contents, C/N (d; reversed plotted), $\delta^{13}\text{C}_{\text{org}}$ and
745 $\delta^{13}\text{C}_{\text{org,corr}}$ (e), and $\delta^{15}\text{N}$ (f) in core EMB201/7-4.
746 **Figure 6.** Depth profiles of Hg/Al (a), Cu/Al (b), Zn/Al (c), Pb/Al (d) and $^{206/207}\text{Pb}$ (e; reversed plotted) in
747 core EMB201/7-4.
748 **Figure 7.** Depth profiles of brassicasterol (a) and tetrahymanol (b) contents, the sum of coprostanol and
749 epicoprostanol contents (c), and the faecal stanol-based R2 ratio (d) in core EMB201/7-4.
750 **Figure 8.** Pictures of core EMB201/7-4 (true colour and radiography), primary markers (a-b), and
751 changepoint modelling results (c) used to define the GSSP level in 1956 ± 4 CE (26.5 cm depth;
752 horizontal line). Secondary markers selected to illustrate changes in anthropogenic pollution (d-e) and
753 eutrophication (f-g) in the Baltic Sea are also shown. A close-up (h) indicating the proposed GSSP level
754 at 26.5 cm in the core is represented by a yellow dotted line.
755 **Table 1.** List of the time markers used to establish the age model of core EMB201/7-4.
756 **Supplementary data**
757 **Table S1.** XRF core scanner data (in count per second) of core EMB201/7-4.
758 **Table S2.** Datasets of core EMB201/7-4. Note that some datasets were not discussed in the text.
759 **Table S3.** Changepoint modelling results.
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777 **Supplemental material**

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