1	Behaviour of Sr, Ca, and Mg isotopes under variable hydrological
2	conditions in high-relief large river systems
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To assess how chemical weathering processes in large high-relief river systems 28 respond to climatic variability, we studied seasonal changes in radiogenic strontium 29 $(^{87}\text{Sr}/^{86}\text{Sr})$ and stable calcium ($\delta^{44/40}$ Ca) and magnesium (δ^{26} Mg) isotopes in the Jinsha 30 and Yalong rivers, which drain the southeastern Tibetan Plateau. During the low-runoff 31 season, with discharge (Q) $<2000 \text{ m}^3/\text{s}$, the river waters reflect the Sr, Ca, and Mg 32 isotope signatures of recharge meltwaters, with additional isotope fractionation signals 33 34 for Ca and Mg related to secondary mineral precipitation, which might imply that meltwater flushes soil solutions from the soil. During medium-runoff intervals 35 (2000m³/s<O<4000m³/s), the Sr, Ca, and Mg isotope signatures in the Jinsha river 36 37 waters are similar to those of the headwaters, which are influenced by evaporite dissolution, while the Yalong is affected by greater carbonate weathering relative to 38 silicate weathering. In both rivers, bedrock dissolution governs the chemical 39 composition of the river waters. During the high-runoff season ($Q>4000m^3/s$), storms 40 generate rapid overland flow, which transfers large volumes of soil into the rivers, such 41 that soil weathering plays an important role in regulating riverine chemical 42 compositions. At these times, the riverine Ca and Sr isotope evolution is influenced by 43 secondary mineral dissolution and sediment-water cation exchange. Overall, this study 44 highlights the potential of combining multiple isotope systems (Sr, Ca, Mg) to trace the 45 dynamics of water-rock interaction under variable hydrological conditions. 46

47 Keywords: Chemical weathering; Hydrological condition; Water-rock interaction; Ca

48 isotopes; Mg isotopes; Tibetan Plateau.

49 **1 Introduction**

50 During silicate rock weathering, primary minerals are dissolved by carbonic acid, 51 releasing ions and alkalinity into river waters. Following their transport to the ocean, 52 calcium (Ca) and magnesium (Mg) are ultimately precipitated from seawater as 53 carbonates, which impacts long-term global climate (Berner and Berner, 2012). 54 However, the sensitivity of chemical weathering in response to climatic factors (e.g. 55 runoff, temperature) remains controversial (Deng et al., 2022; Maher and Chamberlain, 56 2014; Riebe et al., 2004; West et al., 2005).

Large river systems homogenize chemical characteristics from different processes, 57 which provides an opportunity to address the major controls on chemical weathering at 58 a continental scale (Stallard and Edmond, 1983; Tipper et al., 2012c). Furthermore, 59 evaluating river water chemistry in the context of temporal variations in hydrological 60 processes offers a unique chance to determine the response of chemical weathering to 61 62 climatic changes. For example, seasonal variations in water-rock interaction time (Rai and Singh, 2007; Tipper et al., 2006c; Tripathy et al., 2010), changes in the dominant 63 fluid flow path (Calmels et al., 2011; Torres et al., 2015), solute supply from different 64 lithological units (Bickle et al., 2003; Boral et al., 2021; Chao et al., 2015), or input 65 from additional sources (e.g. plant litter, dust, mine water) (Hindshaw et al., 2019; Pett-66 67 Ridge et al., 2009; Zieliński et al., 2018) may all contribute to temporal variations in river water chemistry within a single catchment. In addition to these hydrologic 68

processes impacting the inputs, secondary processes such as cation exchange (Tipper et 69 al., 2021), adsorption by clay minerals (Michalopoulos and Aller, 1995), secondary 70 71 mineral neoformation (carbonates, silicates, and oxides) (Jacobson et al., 2002), plant growth (Berner, 1997) can also vary in response to changing hydrological conditions 72 73 (Clow and Mast, 2010; Maher, 2011), but they can be hard to identify based on only the dissolved concentrations of major and trace elements (Pett-Ridge et al., 2009). As such, 74 the impact of secondary processes on the chemical composition of river water under 75 different hydrological conditions is currently unclear, and the lack of relevant studies 76 77 hinders our understanding of the chemical weathering process.

Fortunately, the development of metal stable isotope measurements has provided 78 a new perspective for understanding chemical weathering processes (Sullivan et al., 79 80 2016). In particular, stable Ca and Mg isotopes have attracted attention because Ca and Mg are major elements in many Earth surface reservoirs (e.g. rocks, waters, plants), 81 represent important nutrients in the ecosystem (Griffith et al., 2020; Tipper, 2022), and 82 83 their elemental cycles are directly coupled to the carbon cycle through silicate weathering and carbonate precipitation (Berner and Berner, 2012; Wimpenny et al., 84 85 2011). Stable Ca and Mg isotopes can be fractionated during the secondary processes described above and can therefore be used to trace these processes within weathering 86 environments (Cenki-Tok et al., 2009; Ewing et al., 2008; Fan et al., 2016; Fantle and 87 Tipper, 2014; Fries et al., 2019; Holmden and Bélanger, 2010; Ma et al., 2015; Nelson 88 et al., 2021; Nelson et al., 2022; Opfergelt et al., 2014; Ryu et al., 2016; Saulnier et al., 89 2012; Tipper et al., 2006a; Tipper et al., 2012a; Wimpenny et al., 2014; Wimpenny et 90

al., 2010). Although rock endmembers have a wide range of Ca and Mg isotopic 91 compositions, which represents a challenge for distinguishing the controls (solute 92 93 sources vs. secondary fractionation) on riverine Ca and Mg isotopes (Fantle and Tipper, 2014; Pogge von Strandmann et al., 2019b; Tipper et al., 2006a; Tipper et al., 2008), 94 95 radiogenic strontium (Sr) isotopes can provide complementary evidence to deconvolve 96 fractionation and lithological controls (Tipper et al., 2006c). By conducting a multiisotope study (stable Ca and Mg, and radiogenic Sr), the significance of secondary 97 processes can be constrained, thereby enabling a better assessment of the controls on 98 99 chemical weathering. Previous studies have suggested that seasonal variations in riverine Ca and Mg 100 isotopes are related to catchment size (Griffith et al., 2020; Tipper, 2022). In small 101 102 catchments, temporal variations of Ca and Mg isotopes respond to both vegetation and water-rock interactions (Cenki-Tok et al., 2009; Dessert et al., 2015; Lehn et al., 2017; 103 Tipper et al., 2008; Tipper et al., 2012b), leading to the potential to use these isotope 104 105 systems to explore the inter-relationships between plants and chemical weathering (Bolou-Bi et al., 2012; Cenki-Tok et al., 2009). In contrast, the factors controlling 106 temporal variations in Ca and Mg isotopes in large river systems remain poorly 107 constrained, with only a few studies having been conducted on large rivers over 108 seasonal timescales (Hindshaw et al., 2019; Mavromatis et al., 2020; Négrel et al., 2021; 109 Schmitt et al., 2003; Tipper et al., 2006b; Tipper et al., 2010; Zhu and Macdougall, 110 1988). In several large rivers, small systematic seasonal variations in Ca isotopes and 111 Mg/Ca ratios have been observed, with heavier Ca isotopic compositions and higher 112

Mg/Ca ratios during the dry season compared to the wet season, potentially indicating 113 an important role for secondary mineral formation/dissolution (Tipper et al., 2010). In 114 115 contrast, the reported seasonal variations in dissolved Mg isotopes in large river systems are negligible (Hindshaw et al., 2019; Mavromatis et al., 2020). Interestingly, studies 116 in Arctic catchments suggest that inter-annual variations in both $\delta^{44/40}$ Ca and δ^{26} Mg 117 values have the potential to trace permafrost degradation (Hindshaw et al., 2019; Lehn 118 et al., 2017), although data from other permafrost settings are required to assess the 119 wider applicability of these findings. In the context of projected increases in the 120 121 frequency of extreme precipitation and permafrost thawing due to global warming (Myhre et al., 2019; Wild et al., 2019), additional research into the behaviour of Ca and 122 Mg isotopes in river water under various hydrological conditions is evidently required 123 124 to determine the effect of climatic factors on the chemical weathering process.

The Tibetan Plateau is the source of 11 large rivers (Wu et al., 2009), and its 125 temperature is rising twice as fast as the global average (Wang et al., 2021). This 126 127 scenario makes it an ideal natural laboratory for evaluating the effect of climate on the water chemistry of large river systems. To explore the hydrological controls on 128 chemical weathering, we measured 87 Sr/ 86 Sr, $\delta^{44/40}$ Ca, and δ^{26} Mg values in river water 129 samples from both the Jinsha and Yalong rivers, two large rivers draining the eastern 130 Tibetan Plateau, spanning one hydrological year. In combination with previous studies 131 that reported on the spatial variation of these isotope systems in the same catchments 132 (Chen et al., 2020; Ma et al., 2020; Zhang et al., 2019; Zhao et al., 2019), we attempt 133 to determine the major processes that control the behaviour of Ca, Mg, and Sr isotopes 134

in the river waters. A better understanding of the seasonal response of Ca, Mg, and Sr
isotopes to weathering processes serves to improve their use as weathering tracers, to
advance our understanding of the mechanisms of chemical weathering, and to better
assess the impact of climate change on elemental and isotopic fluxes to the oceans.

139 2 Research area

The Jinsha River, as one of the main tributaries of the Changjiang River, originates 140 from the Jianggudiru Glacier on Mt. Geladandong at an altitude of 6621 m, drains an 141 area of 340,000 km², and has a length of 2316 km (Lu et al., 2016). The catchment area 142 of the Jinsha River above our sampling site is 264,810 km². The Yalong River is the 143 largest tributary of the Jinsha River, originating from Mt. Bayan Kara at an altitude of 144 5000 m, with a length of 1571 km and a drainage area of 128,000 km² (Li et al., 2014). 145 The fixed points used to monitor seasonal variations in river water chemistry were just 146 upstream of the confluence of these two rivers near Panzhihua (Fig. 1a). 147

148 *2.1 Topography*

From their headwaters in the eastern Tibetan Plateau, the climate, slope, soil, and vegetation vary widely along the course of both the Jinsha and Yalong rivers. The Jinsha River can be divided into a source area, middle reaches, and lower reaches by the towns of Benzilan and Shigu, respectively, while the Yalong River is divided similarly by the city of Ganzi and the town of Dahewan (Fig. 1a). The source areas of the Jinsha and Yalong rivers are plateau terrains with very gentle slopes, which transition via sloping mountainous areas in the middle reaches to flat lowlands in the lower reaches (Li et al., 2014; Ma et al., 2020). Soil depth increases with decreasing elevation, and the
vegetation ranges from alpine meadow in the source area, through arid and semi-arid
shrubs in the middle reaches, to a combination of coniferous forest and sub-tropical
forest in the lower reaches (Ren et al., 2016).

160 *2.2 Climate*

The study area is influenced by the Asian monsoon, with the climate varying 161 temporally from cold and dry in winter to warm and wet in summer, and spatially from 162 a cold-dry highland mountainous climate in the source areas and middle reaches to a 163 subtropical monsoon climate in the lower reaches (Noh et al., 2009; Zhong et al., 2021). 164 Both the Jinsha and Yalong catchments have a distinct wet season (June to October) 165 and dry season (November to May) (Wang et al., 2019) (Fig. 1c), which is conducive 166 to exploring changes in weathering processes under variable hydrological conditions. 167 The two fixed sampling sites are in the lower reaches of the Jinsha and Yalong rivers 168 (Fig. 1a), with temperatures during the sampling period varying from 5.9 to 34.6 °C, 169 170 with an average of 22.1 °C, based on the nearby weather station in Panzhihua (Fig. 1c). The daily water discharge during a hydrologic year ranges from 353 to 6540 m³/s for 171 172 the Jinsha River based on data from Panzhihua station, and from 399 to 8740 m³/s for the Yalong River based on data from Tongzilin station (2013-2014; Hydrological 173 Yearbook of the People's Republic of China). The precipitation is mainly concentrated 174 in the lower reaches of the catchment, contributing up to 70% of the total runoff. The 175 Jinsha and Yalong catchments experience distinct seasonal changes in hydrology, with 176 surface water mainly fed by rainfall during the wet monsoon season (June to October), 177

and recharged by groundwater and meltwater during the dry season (November to May)
(Wang et al., 2019; Zhu, 2007).

180 *2.3 Lithology*

The source area of the Jinsha River contains evaporites (Fig. 1a) and is 181 182 characterised by permafrost conditions (Fig. 1b), whereas neither is found in the source area of the Yalong River (Chen et al., 2020; Zhong et al., 2021). The middle sections of 183 the Jinsha and Yalong rivers are dominated by low-grade Triassic metamorphic rocks 184 (predominantly sandstones) that are intruded by granites, while the lower section is 185 dominated by Palaeozoic carbonate rocks and Precambrian metamorphic rocks (Fig. 1a; 186 Zhao et al., 2019). As the Jinsha and Yalong are large rivers and thus suitable for 187 188 exploring weathering at a continental scale, a comparison between them will enable an assessment of the effects of specific inputs and influences (e.g. evaporites, permafrost) 189 on river water chemistry and isotopes. 190

191 **3** Sampling and methods

192 *3.1 Sampling*

193 The Jinsha and Yalong rivers were sampled just upstream of their confluence at 194 Panzhihua (30 km and 20 km from Panzhihua, respectively; Fig.1a). Monthly sampling 195 of river waters was conducted from November 2013 to October 2014, with increased 196 sampling frequency during the monsoon period (Table 1). Water samples were collected 197 from each river in pre-cleaned bottles. In the field, pH values and water temperature 198 were measured. pH values and water temperature are measured by WTW3210 pH meter (pH measurement accuracy is \pm 0.01, temperature accuracy is \pm 0.1), and EC is measured by lightning conductance meter (accuracy is \pm 5%). All water samples were filtered through 0.45µm Millipore nitrocellulose membrane filters, titrated for alkalinity using HCl, packaged, and stored in clean bottles within 24 hours. Samples for major cations and trace elements were acidified with ultrapure HNO₃. The sampling methods for the water samples are described in detail in Zhong et al. (2021).

In addition, surface soil and bedload sediment samples from near the sampling 205 206 sites of the Jinsha and Yalong rivers were collected during the wet season (August). The 207 soil samples were collected from the river banks with a spade at a depth of 0-10 cm, and placed in a clean plastic bucket after removing plant material and stones. We 208 gathered three random soil samples near the fixed river water sampling site for each 209 210 river, and thoroughly mixed the three samples to make one uniform sample. The bedload sediment samples were collected using a scoop at depths ranging from 0 to 5 211 212 cm and transferred to pre-cleaned wide-mouth glass bottles. The sediment samples are 213 primarily composed of river sand, with sediment particle sizes ranging from 0.125 to 0.5 mm, and more than half of them larger than 0.25 mm. 214

215 *3.2* Soil and sediment leaching and digestions

After air drying the soils, ~10g was ground to *ca*. 200 mesh with an agate mortar, and then leached to isolate the carbonate and exchangeable fractions. The bedload sediment samples were treated in the same way. The leaching procedures followed Bagard et al. (2013): 1N acetic acid was used to obtain the carbonate phase, then 1N HCl was used to obtain the exchangeable phase, and finally 1N HNO₃ was used to

obtain the organic matter fraction (note that the leaching procedures cannot separate 221 each phase quantitatively, and we expect that the 1N HCl leach may also attack the clay 222 223 minerals). The leaching experiments were conducted at room temperature by reacting 1 g solid sample with 15 ml acid for 15 minutes with constant shaking. The mixtures 224 225 were centrifuged at 7000 r/min, and the supernatants were filtered and collected in Teflon beakers. After each leaching step, the solid phase was rinsed three times with 226 MQ water for 10 minutes, and the rinse liquids were added to the previously collected 227 acid leachates. The leachates were dried at 100°C and re-dissolved in 2N HCl. In 228 229 addition, bulk soil and bedload sediment powders were fully digested with concentrated HF and HNO₃. 230

231 *3.3 Major and trace elements*

For river water samples and soil/sediment leachates, major cations (Ca²⁺, Mg²⁺, 232 Na⁺, K⁺, Si) and anions (Cl⁻, NO₃⁻, SO₄²⁻) were determined by Inductively Coupled 233 Plasma Optical Emission Spectrometry (ICP-OES, Vista MPX) and Ion 234 Chromatography Dionex 90 (ICS-90), respectively. Concentrations of trace elements 235 (Al, Sr, Fe, Rb, F) were measured by Inductively Coupled Plasma Mass Spectrometry 236 237 (ICP-MS, Agilent 7900). The concentrations of ions were calibrated against a series of synthetic standard solutions, which were prepared gravimetrically from single element 238 standard solutions (such as Ca, Mg, Na, K, Si, Li, Al, Rb, Sr, Fe) made in 3% HNO₃. 239 During analysis, the MIX-3 standard was used as the monitor solution for the major 240 elemental analysis and the MIX-4 standard was used for the minor elemental analysis. 241 Both blank (3% HNO₃) and synthetic standard solutions (MIX-3 and MIX-4) were 242

243	measured every 5 samples to monitor drift. The external error (2SD) of the analysis is
244	better than 5% for major elements, and better than 10% for minor elements.

245 *3.4 Ca and Sr isotopes*

Calcium and Sr were purified through one column using AG50W-X12 cation exchange resin. After eluting the matrix using 19 ml 2N HCl, Ca solutions were collected using 18 ml 2N HCl, and Sr solutions were then collected using 10 ml 3N HCl. After collection, Ca and Sr solutions were dried down and redissolved in concentrated HNO₃ three times, and finally dissolved in 0.3 M HNO₃ for analysis.

The Ca isotopic compositions (⁴⁴Ca/⁴²Ca) were measured by Nu Plasma HR 251 Multicollector-Inductively Coupled Plasma Mass Spectrometer (MC-ICPMS) in the 252 Isotope Geology Laboratory of the Chinese Academy of Geological Sciences. Isotopes 253 of ⁴²Ca, ⁴³Ca, and ⁴⁴Ca were measured at low mass resolution. Analyses of ⁴⁴Ca/⁴²Ca 254 and ⁴³Ca/⁴²Ca for samples were bracketed by measurements of the NIST 915b standard 255 (Belshaw et al., 2000). Samples were diluted in 0.3 M HNO₃, ensuring that the ⁴²Ca 256 intensity for samples and standards differed by less than 10%. In addition, a DSN 100 257 or CETAC Aridus desolvating nebulizer was used to reduce and stabilise the 258 interferences (less than 2 mV for ⁴²Ca, ⁴³Ca, and ⁴⁴Ca). The detailed operating 259 parameters for stable Ca isotope ratio measurement used in this study are listed in Table 260 **S**1. 261

Calcium isotope values are reported in this study using the delta notation relative
to the NIST SRM915a standard, as shown in equations (1) and (2).

264
$$\delta^{44/42} C a_{915b} (\%) = \left[\frac{\left(\frac{^{44}Ca}{^{42}Ca}\right)_{sample} - \left(\frac{^{44}Ca}{^{42}Ca}\right)_{915b}}{\left(\frac{^{44}Ca}{^{42}Ca}\right)_{915b}} \right] \times 1000$$
(1)

265
$$\delta^{44/42} C a_{915a} (\%_0) = \delta^{44/42} C a_{915b} + 0.35$$
(2)

All samples were analysed 3 to 5 times per session, and the reported values are the 266 mean of those measurements. Repeated analyses of NIST 915b gave a mean value of 267 $\delta^{44/42}$ Ca_{915a} = 0.35 ± 0.05% (2sd, n = 21), in good agreement with literature values 268 (Heuser and Eisenhauer, 2008; Hindshaw et al., 2011), and indicating an external 269 analytical uncertainty of 0.05% (2sd) over the analysis period. Seawater (IAPSO) 270 measured by this method yields $\delta^{44/42}$ Ca_{915a} values of 0.89 ± 0.05% (2sd, n = 5), in line 271 with measurements from other studies that reported a long-term external analytical 272 uncertainty of $\pm 0.05\%$ (2sd) on $\delta^{44/42}$ Ca values (Hippler et al., 2003; Pogge von 273 274 Strandmann et al., 2019a). Other Ca isotope standards measured in this study were COQ-1, BHVO-2, BCR-2, GSR-3, AGV-2, and GSP-2. Their mean values and the 275 corresponding long-term external two standard deviation (2SD), as well as 276 277 measurements of the NIST 915b and IAPSO seawater standards, are summarised in Table S2. All these results are consistent with those reported in previous studies using 278 high-precision double spike TIMS methods (He et al., 2017; Liu et al., 2017; Sun et al., 279 2021). Following the suggestion of Eisenhauer (2004), the Ca isotope results are 280 expressed as $\delta^{44/40}$ Ca relative to the standard NIST 915a, following the relationship in 281 equation (3) (He et al., 2017): 282

$$\delta^{44/40} Ca_{915a} = 2.048 \times \delta^{44/42} Ca_{915a} \tag{3}$$

We apply
$$\pm 0.10\%$$
 (2sd) as the external uncertainty on our sample $\delta^{44/40}$ Ca values.

285	Radiogenic Sr isotopic compositions were measured by Thermo Neptune MC-
286	ICPMS at the School of Earth System Science, Tianjin University. ⁸³ Kr and ⁸⁵ Rb were
287	measured in order to correct for the isobaric interferences of ⁸⁶ Kr and ⁸⁷ Rb on ⁸⁷ Sr
288	(following 86 Kr = 1.50566 *83 Kr and 87 Rb = 0.3857 *85 Rb). Based on the exponential
289	mass bias law, the ⁸⁷ Sr/ ⁸⁶ Sr values were corrected using an ⁸⁶ Sr/ ⁸⁸ Sr ratio of 0.1194
290	(Weis et al., 2006). Repeated analyses of the NBS 987 standard interspersed with the
291	samples gave an 87 Sr/ 86 Sr ratio of 0.71027 ± 0.00002 (2sd, n = 40), in agreement with
292	previous studies (Bickle et al., 2018; Weis et al., 2006; Wu et al., 2009b). Based on the
293	good agreement of the NBS 987 standard with its true value, there was no need to apply
294	a correction to the Sr isotope data. Other Sr isotope standards measured in this study
295	were IAPSO, GSP-2, BCR-2, and COQ-1. Their mean values and the corresponding
296	long-term external two standard deviation (2SD), as well as measurements of the NIST
297	987 standard, are summarised in Table S3. All these results are consistent with those
298	reported in previous studies (Bellefroid et al., 2018; Ma et al., 2013; Weis et al., 2006).

299 3.5 Mg isotopes

The Mg isotopic compositions were measured in the recently-established LOGIC (London Geochemistry and Isotope Centre) clean laboratory and mass spectrometer, and we have therefore analysed a number of standards to assess accuracy and precision. Samples for Mg isotope analysis were purified through a two-stage column procedure using AG50W X-12 cation exchange resin, adapted from Pogge von Strandmann et al. (2011). The first column used 2 M HNO₃ as an eluent, and the second column used 2 M HCl as an eluent, so as to completely separate both Ca and Fe from Mg. Splits were

collected before and after the Mg cut to test Mg concentrations. The combined splits 307 contained <0.5% of the sample Mg, showing that >99.5% of Mg was collected. 308 309 Samples were analysed using a Nu Plasma 3 MC-ICPMS following published procedures (Pogge von Strandmann et al., 2011; Pogge von Strandmann et al., 2019b), 310 311 with a wet plasma setup to eliminate spectral interferences such as CN⁺. With an uptake 312 rate of 100 ml/min and using the "low mass skimmer" from Nu Instruments, the beam intensity was 100 pA total Mg for a 200 ng/g solution. The background (typically <0.1 313 pA) was subtracted from the sample intensity. The total procedural blank was <0.9 ng, 314 315 which is insignificant compared to the mass of Mg in the samples.

At least four secondary standards were processed through columns with each batch of samples, and analyses were performed by sample-standard bracketing relative to the standard DSM-3 (Galy et al., 2003). Magnesium isotope values are reported in this study using the delta notation relative to the DSM-3 standard, as shown in equation (4).

320
$$\delta^{x} Mg = \left[\frac{\left(\frac{{}^{x} Mg}{{}^{24} Mg}\right)_{sample} - \left(\frac{{}^{x} Mg}{{}^{24} Mg}\right)_{standard}}{\left(\frac{{}^{x} Mg}{{}^{24} Mg}\right)_{standard}} \right] \times 1000$$
(4)

Note that ^xMg is either ²⁵Mg or ²⁶Mg. Seawater (IAPSO) and DSM-3 analysed during the study gave δ^{26} Mg values of -0.82 ± 0.04‰ (2sd, n=11) and 0.02 ± 0.06 (2sd, n=14), respectively, in agreement with previously published values (Pogge von Strandmann et al., 2019a). The Cambridge-1 (CAM-1) standard gave a δ^{26} Mg value of -2.64 ± 0.07‰ (n=5), and the USGS BCR-2 standard gave a δ^{26} Mg value of -0.24 ± 0.08‰ (n=7), also in agreement with other studies (see compilation in Pogge von Strandmann et al., 2011). In a ²⁶Mg vs. ²⁵Mg cross plot (Fig. S1), all samples lie on a line with a slope of 0.523 ± 328 0.002, which is consistent with the theoretical equilibrium slope of 0.521 (Young and 329 Galy, 2004). Therefore, both water and rock analyses appear to be accurate, and the 330 long-term external analytical uncertainty of δ^{26} Mg values using this method and mass 331 spectrometer is taken to be $\pm 0.08\%$ (2sd).

332 3.6 Elemental flux calculations and concentration modelling

The LOADEST software, developed by the United States Geological Survey (USGS), was used to estimate constituent loads in rivers. Using continuous flow and discrete water quality monitoring data, the LOADEST model can automatically build up a statistical regression equation for daily river dissolved loads, which can then be estimated on a daily/monthly/seasonal/annual time scale. The software includes 9 models. In this study, AMLE was used to calculate the daily elemental concentrations and fluxes, and the detailed LOADEST settings used here are reported in Table S4.

340 **4 Results**

341 *4.1 Hydrological conditions*

Both the Jinsha and Yalong rivers have high and variable discharge during the wet season from July to September (Fig. 1 and 2, Table S5). The discharge then falls sharply between October and December due to low rainfall. The discharge of the Jinsha River reaches its minimum during the winter (from January to March) when the source river is frozen, and then increases slightly towards June as temperatures rise (Fig. 2, Table S5). For the Yalong River, there are rapid variations in discharge on monthly timescales during the dry season, which are not reproduced in the runoff of the Jinsha River during

349	the same period (Fig. 2, Table S5). This apparent flow anomaly may be related to
350	seasonal reservoir regulation linked to the Er'tan Hydropower Station, the second-
351	largest hydropower station in Asia, which is located 33 km upstream from our sampling
352	site at the outlet of the Yalong River.

353 4.2 Temporal variability of river water chemistry and isotopic compositions

354 4.2.1 Variations of riverine elemental concentrations with discharge

The concentrations of major anions and cations (Zhong, 2017) in the Jinsha and Yalong river waters show variability during the hydrological year, as do concentrations of trace elements such as Sr, Fe, Al, and F (Fig. 2, Table 1). Moreover, the seasonal trends in concentrations differ between the two rivers.

The major soluble elements in the Jinsha River show broadly similar temporal 359 360 patterns but with some differences in detail (Fig. 2a). For example, Ca, Mg, Na, K, SO₄, and Cl all show concentration peaks during both the pre-monsoon interval in June and 361 during a post-monsoon interval, although the timing of the post-monsoon peak varies 362 between elements (November to February). Similar seasonal trends are seen for minor 363 elements such as Sr, F, and Fe (Fig. 2a, c). For most elements, the lowest concentrations 364 are recorded during the monsoon interval, but for Na and Cl significantly lower values 365 are reached in April to May (Fig. 2a, c). 366

For the Yalong River, major ion concentrations (Ca, Mg, Na, K, SO₄, and Cl) and Sr concentrations show similar patterns to each other, with gradual increases from October to May, followed by a steeper decrease in early July during the monsoon season (Fig. 2b and 2d). Unlike the Jinsha River, the peak concentrations occur later in the year (around April to May) rather than in the post-monsoon interval and there is no distinct
pre-monsoon peak in June (Fig. 2b and 2d). In general, the cation concentrations of the
Yalong River follow a concentration-discharge relationship (Godsey et al., 2009) at low
to moderate flows, indicating some effect of dilution (but offset above a theoretical
dilution line), while the behaviour is closer to chemostatic during the high-water stage
(Fig. 3b-d). Hence, the effect of dilution during the monsoon interval appears to be
limited by enhanced contributions of solutes from other sources.

Additionally, both the Jinsha and Yalong rivers display similar trends in Al concentrations, which in both rivers show an increase during the monsoon season and a positive correlation with discharge (Fig. 2c, d). The concentrations of Si remain relatively constant through time in both the Jinsha and Yalong rivers (Fig. 3a), with such chemostatic behaviour potentially arising from an equilibrium between the river waters and secondary silicates (Clow and Mast, 2010; Torres et al., 2015).

384 4.2.2 Variations of riverine ⁸⁷Sr/⁸⁶Sr values

Both the Jinsha and Yalong river waters have less radiogenic ⁸⁷Sr/⁸⁶Sr values 385 (0.71031-0.71086 and 0.71056-0.71192, respectively; Table 1) than Himalayan rivers 386 (Boral et al., 2021; Tipper et al., 2006c). The ⁸⁷Sr/⁸⁶Sr values in the Jinsha River are 387 relatively homogeneous, with a narrow range of variation during the whole 388 hydrological year (Figs. 3f and 4a, Table 1). The greatest variation in ⁸⁷Sr/⁸⁶Sr values 389 for the Jinsha River is during the low-water stage ($Q < 2000 \text{ m}^3/\text{s}$) and is inversely 390 related to discharge (Fig. 3f). During the high-water stage (Q>4000 m³/s), the ⁸⁷Sr/⁸⁶Sr 391 values are fairly constant, at around 0.71051 for the Jinsha River. The Yalong River also 392

shows a slightly greater range of 87 Sr/ 86 Sr values during the dry season (0.71056-0.71177) than during the wet season (0.71095-0.71192) (Figs. 3f and Fig. 4b), but displays a positive trend between 87 Sr/ 86 Sr values and discharge when Q < 2000 m³/s (Fig. 3f). During the high-water stage (Q>4000 m³/s), the 87 Sr/ 86 Sr values are fairly constant at around 0.71128 for the Yalong River (Fig. 3f).

398 4.2.3 Variations of riverine $\delta^{44/40}$ Ca values

The seasonal variation in $\delta^{44/40}$ Ca values of the Jinsha and Yalong rivers falls 399 within the range of spatial variation observed in the Jinsha and Yalong catchments 400 (Chen et al., 2020; Chen et al., 2022), but the relationship between $\delta^{44/40}$ Ca values and 401 discharge in the two rivers is different (Fig. 3g). For the Jinsha River, the $\delta^{44/40}$ Ca values 402 span a larger range from 0.78% to 1.17% (Fig. 3g, Table 1), displaying a positive trend 403 with discharge when $Q < 2000 \text{ m}^3/\text{s}$, then maintaining high values when $2000 \text{ m}^3/\text{s} < Q$ 404 $< 4000 \text{ m}^3/\text{s}$, followed by a negative trend with discharge when Q > $4000 \text{ m}^3/\text{s}$ (Fig. 3g). 405 Overall, the variation of $\delta^{44/40}$ Ca values in the Jinsha River is larger in the wet season 406 (0.78-1.17%) than in the dry season (0.98-1.13%). For the Yalong River, the $\delta^{44/40}$ Ca 407 values fluctuate in a fairly narrow range between 0.71% and 0.95% (Fig. 3g, Table 1), 408 but show some interesting trends with discharge. Specifically, there is a negative 409 correlation between $\delta^{44/40}$ Ca values and discharge in the dry season (Q < 2000m³/s), 410 and a positive correlation when $2000m^3/s < Q < 4000m^3/s$, before values stabilise at 411 0.94% when Q > 4000m³/s (Fig. 3g). 412

413 4.2.4 Variations of riverine $\delta^{26}Mg$ values

414 Both the Jinsha and Yalong rivers have δ^{26} Mg values that fall within the range of

415	global rivers (from -2.50% to $+0.64\%$, with a mean of -1.09%) (Teng, 2017). The
416	Jinsha River has a range of 0.27% (from -1.30% to -1.03%) (Fig. 3h), which is much
417	smaller than the spatial variation in river waters in the Jinsha catchment (-1.67% to -
418	0.50%) (Xu et al., 2022; Zhao et al., 2019). The Jinsha River δ^{26} Mg values are more
419	variable during the dry season than during the wet season (Fig. 4a), increasing with
420	discharge during the dry season (Q<2000m ³ /s), then maintaining high values under
421	intermediate to high flows (Q>2000m ³ /s) (Fig. 3h). For the Yalong River, the δ^{26} Mg
422	values have a slightly larger range of 0.39% (from -1.39% to -1.00%). The Yalong
423	River δ^{26} Mg values are more variable during the wet season than during the dry season
424	(Fig. 4b), displaying a negative trend with discharge during both dry (Q<2000m ³ /s) and
425	wet seasons ($Q>4000m^3/s$) (Fig. 3h).

426 *4.3 Leaching experiments*

Table 2 reports the elemental ratios and isotopic compositions of the bedload 427 sediment and soil leachates and digests. The carbonate fraction of the two soils (JSJA-428 1 and YLJA-1) have similar $\delta^{44/40}$ Ca values (~0.50%) and slightly different ⁸⁷Sr/⁸⁶Sr 429 values (0.70997 and 0.71091). In contrast, the carbonate fraction of the river sediments 430 (JSJN-1 and YLJN-1) yielded a wide range of $\delta^{44/40}Ca$ values (0.54-0.79‰) and 431 ⁸⁷Sr/⁸⁶Sr values (0.70957-0.71228). Considering the exchangeable fraction, the Sr 432 isotopic composition of the soil on the banks of the two rivers (JSJA-2 and YLJA-2) is 433 similar (~0.71200), but the Ca isotopic compositions are different (0.55 and 0.73%). 434 The exchangeable fraction of the two river sediments (JSJN-2 and YLJN-2) have 435 similar $^{87}\text{Sr}/^{86}\text{Sr}$ values (~0.70980) and slightly different $\delta^{44/40}\text{Ca}$ values (0.72% and 436

437 0.85‰). Furthermore, the ⁸⁷Sr/⁸⁶Sr values of the organic matter fractions in the soil and
438 river sediment samples are both higher than the ⁸⁷Sr/⁸⁶Sr values of the respective
439 carbonate and exchangeable fractions. This observation could indicate that the 1N
440 HNO₃ has also attacked some silicate mineral debris.

441 **5 Discussion**

442 5.1 Hydrological influences on chemical weathering rates

443 Previous studies indicated that the Jinsha and Yalong rivers are dominated by carbonate weathering (Ma et al., 2020; Noh et al., 2009; Wu et al., 2008), and that the 444 mainstream of the Jinsha is also impacted by evaporite dissolution in its source area 445 446 (Chen et al., 2020). A forward model following Noh et al. (2009) was used to calculate the seasonal contributions from rainwater, silicates, carbonates, and evaporites (Table 447 448 S6). We then calculated the weathering rates in the study area, and also calculated ω_{sil} , ω_{carb} , and ω_{eva} as the proportions of silicate, carbonate, and evaporite weathering 449 rates to total chemical weathering rates. The results are shown in Table S7, and details 450 of the calculation can be found in Figure S2. 451

For both the Jinsha and Yalong rivers, chemical weathering rates are higher in the wet season than in the dry season (Fig. S3), while silicate weathering rates (SWR) are lower than carbonate weathering rates (CWR) during the whole hydrologic year (Fig. 5a and 5b). The CWR in the wet season is up to 16 and 14 times higher than in the dry season for the Jinsha and Yalong rivers, respectively (Fig. S3, Table S7). The CWR and SWR of the Jinsha and Yalong rivers show subtle variations during the dry season, then

increase rapidly at the pre-monsoon stage (June), stay at relatively high levels during 458 the monsoon interval (July to September), and then decrease sharply after the monsoon 459 (October) (Table S7). Evaporite dissolution plays an important role in influencing the 460 Jinsha river water chemistry (EWR: 5-212 t/km²/y), whereas the EWR of the Yalong 461 River is insignificant (3-20 t/km²/y) (Fig. 5a and 5b, Fig. S3, Table S7). For the two 462 rivers, although SWR are elevated during the wet monsoon season, ω _sil reaches a 463 fairly stable minimum during this interval (Fig. 5a and 5b), in contrast to elevated ω_{eva} 464 for the Jinsha and ω _carb for the Yalong. 465

466 5.2 Major solute sources to the Jinsha and Yalong

Even when the solute concentration behaviour is largely chemostatic (Fig. 3), 467 changes in the sources of elements to river waters could still be occurring (Clow and 468 Drever, 1996; Clow and Mast, 2010) and can potentially be traced using radiogenic Sr 469 isotopes (Pett-Ridge et al., 2009). Previous studies on small watersheds have shown 470 that longer water residence times during the dry season can lead to greater contributions 471 from silicate versus carbonate weathering, which typically increases the riverine 472 ⁸⁷Sr/⁸⁶Sr values, while increased carbonate dissolution during the wet season can lower 473 ⁸⁷Sr/⁸⁶Sr values (Tipper et al., 2006c). However, this scenario is not entirely applicable 474 to the large river catchments of this study. There is no significant seasonal relationship 475 between riverine ⁸⁷Sr/⁸⁶Sr values and the ratio of SWR to CWR in either the Jinsha or 476 Yalong river basins (Table S7). Therefore, seasonal changes in the relative contributions 477 of silicate and carbonate weathering do not appear to be the main factor responsible for 478 temporal variability in the ⁸⁷Sr/⁸⁶Sr values of these rivers. 479

480	Instead, mixing between different water sources under changing hydrological
481	conditions may be the dominant control on the temporal variations in ⁸⁷ Sr/ ⁸⁶ Sr values
482	(Calmels et al., 2011; Jacobson and Blum, 2000; Jacobson et al., 2002). This hypothesis
483	is tested using mixing diagrams between ⁸⁷ Sr/ ⁸⁶ Sr values and elemental ratios (Fig. 6).
484	Potential water sources include groundwater, meltwaters, and surface waters. Based on
485	chemical composition, the surface waters can be divided into three types: the main
486	tributaries of the two rivers which are dominated by carbonate weathering (Chen et al.,
487	2022), the headwaters of the Jinsha river which are severely affected by evaporite
488	dissolution (Chen et al., 2020), and the overland flow which reacts rapidly with topsoil
489	during the monsoon season (Torres et al., 2015). The compositions of evaporites and
490	soil leachates in the study area (Table 2) are taken to provide evidence on the chemical
491	characteristics of the latter two water sources. Although time series samples of both the
492	Jinsha and Yalong fall in a mixing region between meltwaters, evaporite-draining river
493	waters, and soil leachates (grey area in Fig. 6), the contribution of groundwater cannot
494	be ruled out. During the dry season, groundwater and meltwater are potential water
495	sources due to the drying of the evaporite-dominated headwaters and the lack of
496	overland flow to carry soil into the river system (Xie et al., 2018). As such, we consider
497	that the river water sources differ depending on hydrological conditions, resulting in
498	the seasonal variations in river water composition.

499 5.3 Controls during low-runoff periods: Meltwater and secondary mineral 500 precipitation

501

During low-runoff periods (Q < 2000 m³/s), the rivers are mainly recharged by

groundwater (Fig. 9a and b). Amorphous clay minerals can form within a few weeks in 502 groundwaters that are forced out of equilibrium, at least to the extent that they can be 503 detected by sensitive methods such as Mg isotopes (Oelkers et al., 2019). However, in 504 the case of chemical equilibrium, the precipitation rates of crystalline clays tend to be 505 much slower (Jacobson and Holmden, 2008; Jacobson et al., 2010). We therefore 506 assume that seasonal variations in the chemistry and isotopic compositions of the 507 groundwaters are mainly inherited from their hydrologic sources rather than from 508 variability in in-situ water-rock interactions. In this scenario, the groundwater 509 510 compositions are mainly affected by the infiltration of meltwaters during the dry season, and by river water recharge during the wet season, rather than representing a distinct 511 source themselves. 512

513 5.3.1 Jinsha River basin: Permafrost meltwater and secondary carbonate 514 precipitation

The source area of the Jinsha River is covered by permafrost (Fig. 1b), in which 515 Ca-Mg sulphate salts are commonly formed (Cooper et al., 2002; Lehn et al., 2017; 516 Rutter et al., 2011). From March onwards, as the permafrost begins to thaw, the 517 dissolution of Ca-Mg sulphate salts dominates the chemical composition of the 518 519 permafrost meltwater. Unfortunately, no samples of permafrost meltwater during the low-runoff period are available. Nonetheless, previous research has demonstrated that 520 permafrost meltwater in the Jinsha source area and the Jinsha headwaters share the same 521 522 water source (atmospheric precipitation) and solute source (Ca-Mg sulphate dissolution) (Ye and Chang, 2019). As a result, the ⁸⁷Sr/⁸⁶Sr value of the permafrost meltwater 523

should be comparable to that of the river at the Jinsha headwater (sample JS-1 in Chen
et al., 2022). Due to the river drying, the headwaters of the Jinsha River are unable to
provide solutes to the main stream during the low-runoff periods (Xie et al., 2018). At
this point, the ⁸⁷Sr/⁸⁶Sr values in the Jinsha River decrease with increasing flow (Fig.
3f), which can only be attributed to permafrost meltwater input.

In addition, the Ca-Mg sulphate salts in the Jinsha source area have lower $\delta^{44/40}$ Ca 529 (Chen et al., 2020) and similar δ^{26} Mg values (Zhao et al., 2019) compared to the river 530 water samples collected during low-runoff periods (Table 2). As a result, the increase 531 in the $\delta^{44/40}$ Ca and δ^{26} Mg values with increasing flow in the Jinsha River (Fig. 3g and 532 3h) cannot be explained solely by the dissolution of the Ca-Mg sulphate salts. Instead, 533 the Ca-Mg isotopes in the permafrost meltwater could be fractionated by other 534 535 processes. Previous studies have suggested that secondary carbonate precipitation, clay adsorption, and cation exchange preferentially take up lighter Ca and Mg isotopes 536 (Brazier et al., 2019; Cenki-Tok et al., 2009; Jacobson and Holmden, 2008; Nelson et 537 al., 2021; Teng, 2017; Tipper et al., 2006a). In contrast, plant growth preferentially takes 538 up heavy Mg as well as light Ca isotopes (Black et al., 2008; Bolou-Bi et al., 2010; 539 Bolou-Bi et al., 2012; Nelson et al., 2022; Page et al., 2008; Schmitt et al., 2003; 540 Wiegand, 2005), resulting in liquid phases having light Mg and heavy Ca isotopic 541 compositions, which is inconsistent with the patterns observed for $\delta^{44/40}$ Ca and δ^{26} Mg 542 values in Jinsha River waters (Fig. 8b). Hence, vegetation is not the controlling factor 543 driving Ca-Mg isotope fractionation. Although we cannot rule out effects from clay 544 adsorption or cation exchange, the positive correlations between $\delta^{44/40}$ Ca values and 545

Sr/Ca ratios (Fig. 7c), and the positive correlation between δ^{26} Mg values and the 546 dolomite saturation index (DSI) (Fig. 8d, Table S8) for the Jinsha River in the dry 547 season, suggest that secondary carbonate precipitation plays an important role in 548 fractionating Ca and Mg isotopes (Fig. 9a). Previous studies have also demonstrated 549 that Ca and Mg removal by secondary carbonate precipitation is common in eastern 550 551 Tibetan and Himalayan river catchments (Bickle et al., 2005; Bickle et al., 2018; Chen et al., 2020; Jacobson et al., 2002; Tripathy et al., 2010; Zhao et al., 2019), so our 552 interpretation appears reasonable. 553

554 In permafrost regions, increasing air temperatures deepen the seasonally-thawed zone, leading to the dissolution of Ca-Mg sulphate salts. Consequently, carbonate 555 saturation indices are elevated, facilitating carbonate precipitation, and thereby 556 increasing the dissolved $\delta^{44/40}$ Ca and δ^{26} Mg values in the Jinsha River basin (Fig. 9a). 557 In addition, the δ^{26} Mg values in the Jinsha river waters exhibit a negative correlation 558 with ⁸⁷Sr/⁸⁶Sr (Fig. 8a) and also show a positive correlation with chemical ratios 559 560 indicating evaporite contributions, such as SO₄/Mg (Fig. S4d). Further, although the variations may be negligible when considering analytical error ($\pm 0.10\%$), a general 561 positive correlation between temperature and $\delta^{44/40}$ Ca values in the Jinsha River is 562 observed during the dry season (Fig. 7d), which is also consistent with the operation of 563 secondary carbonate precipitation. 564

565 5.3.2 Yalong River basin: Snow meltwater and secondary silicate formation

566

Unlike the Jinsha River, the Yalong River is not influenced by evaporite

567	dissolution (Fig. 1) (Chen et al., 2022; Ma et al., 2020). The ⁸⁷ Sr/ ⁸⁶ Sr values in the
568	Yalong river waters increase with increasing discharge during the dry period (Fig. 3f),
569	which can be attributed to an increasing input of snow meltwater (Fig. 6). In addition,
570	the δ^{26} Mg values of the Yalong River show a negative correlation with illite saturation
571	indices (ISI) (Fig. 8c) in the low-runoff season, implying that the formation of clay
572	minerals dominates the Mg isotopic composition of the river waters. Even though the
573	direction of Mg isotope fractionation for clay minerals is not uniform (Brewer et al.,
574	2018; Hindshaw et al., 2020; Pogge von Strandmann et al., 2008; Ryu et al., 2016;
575	Wimpenny et al., 2010), experimental data indicate that the formation of Mg-rich clay
576	minerals (e.g. illite and montmorillonite) preferentially incorporates ²⁶ Mg, causing the
577	residual solutions to be enriched in ²⁴ Mg (Wimpenny et al., 2014; Wimpenny et al.,
578	2010). Illite is the most abundant clay mineral in the suspended particulate matter and
579	river sediments of the upper Changjiang River (Ding et al., 2013) and the ISI of the
580	Yalong River shows a positive correlation with discharge during the low-runoff period
581	(Fig. S4 and Table S8). We therefore suggest that meltwaters may wash out "old" waters
582	from the soil, which had developed lower δ^{26} Mg values due to illite precipitation during
583	the dry season (Fig. 9b), such that the soil water partly controls the Mg isotopic
584	composition of the Yalong river waters during the dry season. Alternatively, since
585	meltwater seeps through the soil and into the groundwater that recharges the river
586	waters, we cannot rule out the possibility of secondary illite formation occurring in the
587	groundwater. Future monitoring of the chemical composition of groundwater and soil
588	water may aid in determining the location of secondary mineral formation in such

589 systems.

Compared with the Jinsha River, the Yalong River shows a narrower range of 590 $\delta^{44/40}$ Ca values during the hydrological year and does not show a significant correlation 591 between $\delta^{44/40}$ Ca values and SO₄/Ca ratios, CSI, or discharge during the dry season (Fig. 592 7). This differing pattern could be attributed to reservoir operations, which release dam 593 water to increase the water discharge during the low-runoff periods (Wang et al., 2019). 594 Dam-released water has a lower pH value and temperature than river water, making it 595 undersaturated for carbonates (Négrel et al., 2021; Wang et al., 2022). As a result, when 596 597 dam water enters the river, it can be expected to reduce carbonate saturation and inhibit secondary carbonate precipitation. The carbonate saturation indices of the Yalong River 598 decrease with increasing discharge during the low-runoff period (Fig. S4a and 4b), 599 600 which might indicate the input of water released from the reservoir. For the Yalong river water, carbonates change from supersaturated to undersaturated (Fig. S4a), which 601 would presumably result in the dissolution of some of the previously precipitated 602 603 secondary carbonates. Secondary carbonate precipitation preferentially uptakes the lighter Ca and Mg isotopes (Tipper et al., 2006a), so their dissolution would decrease 604 the Ca and Mg isotopic compositions of the Yalong river water (Fig. 3g and 3h). 605 Because Sr is rarely found in secondary carbonates (Stewart et al., 1998), the ⁸⁷Sr/⁸⁶Sr 606 values of the Yalong river waters are not affected by the secondary carbonate 607 dissolution. In conclusion, the changing trend of the carbonate saturation index, Ca and 608 Mg isotopic compositions during the low-runoff period in the Yalong River may 609 indicate the influence of hydrological regulation due to dam construction. 610

In summary, as the air temperature increases during the dry season, some of the 611 meltwater might infiltrate the soil and wash out the aged soil solution, which in turn 612 613 affects river water chemistry. Meanwhile, some of the meltwater might infiltrate into groundwater, indirectly affecting the chemical composition of the river water by 614 615 regulating the composition of the groundwater. The Jinsha River is mainly recharged 616 by permafrost meltwaters during the dry season (Fig. 9a), whereas the Yalong River is fed by meltwaters from snow and glaciers and dam waters (Fig. 9b). These high-relief 617 large river systems might preserve the Sr, Ca, and Mg isotope signatures of recharge 618 619 meltwaters that interacted with soil porewaters and groundwater, where both Ca and Mg isotopes are fractionated by the precipitation of secondary minerals i.e. carbonates 620 in the case of the Jinsha River, and silicates in the case of the Yalong River (Fig. 9a and 621 622 9b).

623 5.4 Controls during medium-runoff periods: Tributary mixing and rock weathering

When the intensity of precipitation is greater than the soil infiltration capacity, the 624 rainwater that exceeds the infiltration capacity forms surface runoff (Yang et al., 2014). 625 During these times of increasing discharge, we hypothesise that the chemical 626 composition of the Jinsha and Yalong rivers are dominated by the mixing of waters 627 from the tributaries and the unfrozen headwaters, which carry rock weathering 628 information from these sub-catchments (Fig. 9c and 9d). Such source mixing is thought 629 to be the main cause of seasonal variations in the chemical composition of many large 630 river systems, including the Ganga and Brahmaputra (Bickle et al., 2003; Boral et al., 631 2021), the Irrawaddy and Salween (Chapman et al., 2015), and the lower Changjiang 632

633 River (Luo et al., 2014).

In the range of $2000 \text{m}^3/\text{s} < Q < 4000 \text{m}^3/\text{s}$, the weathering proportions of each rock 634 635 endmember in the Jinsha and Yalong rivers are insensitive to discharge (Fig. S3). Compared to the dry season, the proportion of evaporite weathering in the Jinsha river 636 637 basin is higher (average value of samples from July to September: 35%), and the proportion of silicate (26%) and carbonate (39%) weathering is lower (Fig. 5a, Fig. S3, 638 Table S7). During these periods, the 87 Sr/ 86 Sr, $\delta^{44/40}$ Ca, and δ^{26} Mg values in the Jinsha 639 river waters are similar to those of the headwaters, which are influenced by evaporite 640 641 dissolution (Fig. 6, 7, and 8), indicating that the evaporite-dominated source rivers govern the water chemistry and isotopic compositions of the Jinsha River (Fig. 9c). 642 This finding is consistent with previous research on the spatial variation of isotopes 643 644 during summer seasons, which showed that evaporite weathering controls the Li (Ma et al., 2020), Mg (Zhao et al., 2019), Ca, and Sr isotopic compositions (Chen et al., 645 2020; Chen et al., 2022) of the Jinsha mainstream. For the Yalong River, the proportion 646 647 of carbonate weathering (average value of samples from July to September: 70%) is slightly higher during these wetter periods compared to the dry season, and the 648 proportion of silicate (27%) and evaporite weathering (2%) is slightly lower (Fig. 5b, 649 Fig. S3, Table S7), indicating greater carbonate dissolution relative to silicate 650 weathering during this period. Apart from the source area of the Jinsha River, 651 carbonates are widely distributed in both the Jinsha and Yalong river basins (Fig. 1a), 652 and the water chemistry of most of their tributaries is dominated by carbonate 653 weathering (Chen et al., 2020; Ma et al., 2020; Zhang et al., 2019). Carbonates have 654

relatively lower ⁸⁷Sr/⁸⁶Sr, $\delta^{44/40}$ Ca, and δ^{26} Mg values than silicates (Table 2), and therefore the decreases in ⁸⁷Sr/⁸⁶Sr, $\delta^{44/40}$ Ca, and δ^{26} Mg values in the Yalong river waters as discharge increases in the range of 2000m³/s<Q<4000m³/s (Fig. 3f-h) could be attributed to inputs from carbonate-dominated tributaries (Fig. 9d).

In summary, we suggest that the increase in precipitation during the intermediateflow periods increases the catchment area in the whole catchment, which might induce more rock weathering. The chemical composition of the Jinsha River is dominated by inputs from evaporite-dominated surface flow (Fig. 9c), whereas the Yalong River water is governed by inputs from carbonate-dominated surface flow (Fig. 9d).

5.5 Controls during the high-runoff period: Rapid surface flow and soil weathering

665 Based on research in watershed hydrology, overland flow is generated during heavy rainfall periods and flows through gullies into the river network (Yang et al., 666 2014). Such overland flows can also erode surface soils, and thereby carry both soil 667 debris and soil water into the rivers. In previous studies, the weathering of soil minerals 668 has been shown to impact river water chemical compositions during monsoon 669 conditions (Bagard et al., 2011; Gurumurthy et al., 2015; Wei et al., 2013). We therefore 670 suggest that soil weathering may contribute to the dissolved solutes in the Jinsha and 671 Yalong river waters during the heavy rainfall period (Q>4000m³/s). Support for a 672 surface soil input also comes from a recent study on the same river water samples, 673 which showed that dissolved organic carbon (DOC) concentrations increase during the 674 storm periods (Zhong et al., 2021). 675

676 5.5.1 Exchange reactions

It was recently demonstrated that the chemistry of the exchange pool on suspended 677 particulate matter is in exchange equilibrium with river water in a series of global rivers 678 679 (Tipper et al., 2021). In this case, we hypothesize that changes in the chemical composition of the river waters during the high-runoff period could be caused by inputs 680 from the soil exchange pool. During the high-runoff period, the ⁸⁷Sr/⁸⁶Sr values in the 681 Jinsha and Yalong river waters (Table 1, Jinsha: 0.71038-0.71065; Yalong: 0.71114-682 0.71155) are higher than in the exchangeable fraction of the riverbed sediments 683 collected from this period (Table 2, JSJN-2: 0.70978 and YLJN-2: 0.70990). Since the 684 685 relative contributions from different lithologies appears to have been almost constant (Fig. 5), the input of Sr from the soil exchange pool with a high Sr isotopic composition 686 (Table 2, ~ 0.71200) could provide a possible solution. 687

688 In addition, Ca isotope data also support a soil exchange pool input during the high-runoff period. For the Yalong River, the $\delta^{44/40}$ Ca value of the soil exchangeable 689 fraction (0.73±0.10%) is consistent with the $\delta^{44/40}$ Ca value of the river water (Table 2, 690 $\sim 0.90 \pm 0.10\%$) and the exchangeable fraction of the river sediment (0.85 \pm 0.10\%), 691 within analytical uncertainty. Therefore, the input of Ca from the soil exchange 692 reservoir would not cause a significant change in the $\delta^{44/40}$ Ca value of the river, which 693 is consistent with the $\delta^{44/40}$ Ca value of the Yalong River being stable at 0.90% during 694 the high-runoff period (Fig. 3g, Fig. 4b). For the Jinsha River, the $\delta^{44/40}$ Ca value of the 695 soil exchangeable fraction (0.55±0.10%) is lower than that of the river water (0.78-696 1.14%), so with the increase of soil reservoir input, the $\delta^{44/40}$ Ca value of river water 697 would be expected to gradually decrease, as observed (Fig. 3g, Fig. 4a). 698

699 5.5.2 Dissolution of secondary soil minerals

Secondary minerals such as pedogenic carbonates precipitate in soils following 700 701 previous flood events and can then be washed into rivers and re-dissolved during the wet season (Huth et al., 2019; Kajita et al., 2020; Meyer et al., 2014). Unlike the almost 702 constant radiogenic Sr isotopes, the $\delta^{44/40}$ Ca values in the Jinsha River system decrease 703 with increasing discharge during the high-runoff period (Fig. 3g). In addition to the 704 input of the soil exchange pool mentioned above (Section 5.5.1), another possible 705 explanation is that the input of Ca from secondary carbonates in soils (Table 2, JSJA-706 1) could lower the $\delta^{44/40}$ Ca values of river water, while having no significant impact on 707 the riverine ⁸⁷Sr/⁸⁶Sr composition, since secondary calcite has a low Sr concentration 708 (Tesoriero and Pankow, 1996). Compared with the Jinsha River, the dissolution of 709 secondary carbonates in soil has less impact on the Yalong River (Fig. 3g), which may 710 be an expected consequence of the limited production of secondary carbonates in the 711 Yalong River basin (Chen et al., 2022). In addition to secondary carbonates, the constant 712 dissolved Si concentrations throughout the hydrological year in both rivers suggests a 713 buffering by the precipitation and dissolution of secondary silicates (Fig. 3a) (Clow and 714 Mast, 2010; Georg et al., 2006; Torres et al., 2015). Since Ca is the dominant element 715 in carbonates but a trace element in clay minerals, we expect the weathering of 716 secondary silicates to have little effect on the riverine $\delta^{44/40}$ Ca values (Nelson et al., 717 2022). 718

The effects of both secondary mineral precipitation and exchange reactions on riverine Mg isotopes remain controversial. While secondary carbonates preferentially take up light Mg isotopes (Mavromatis et al., 2013; Oelkers et al., 2018; Pearce et al.,

722	2012; Schott et al., 2016), the direction of Mg isotope fractionation during clay mineral
723	formation is controlled by the mineral type, structure, and solution conditions (Chapela
724	Lara et al., 2017; Hindshaw et al., 2020; Pogge von Strandmann et al., 2008).
725	Furthermore, illite preferentially adsorbs light Mg isotopes, resulting in higher δ^{26} Mg
726	values in solution (Wimpenny et al., 2014), whereas gibbsite and Fe-(hydr)oxides
727	preferentially adsorb ^{26}Mg , resulting in lower solution $\delta^{26}Mg$ values (Chapela Lara et
728	al., 2017; Liu et al., 2014). In the Jinsha and Yalong river basins, illite is the most
729	abundant clay mineral, but the increase in Al and Fe concentrations when Q>4000m ³ /s
730	(Table 1) could also be related to the input (and dissolution) of gibbsite and Fe-
731	(hydr)oxides during storms (Bagard et al., 2011). Therefore, the effects of adsorption
732	by different clay minerals and oxides on Mg isotopes in the river waters may
733	approximately cancel out. For both rivers, the Mg isotope data from the monsoon
734	season show a limited range of values (Fig. 4), which suggests that any fractionation is
735	minor, making it hard to resolve the possible effects of cation exchange or secondary
736	mineral dissolution based on Mg isotopes. However, the observed correlations between
737	δ^{26} Mg and 87 Sr/ 86 Sr values throughout the hydrological year (Fig. 8a) seem to indicate
738	that lithology is the primary control on seasonal variations in riverine δ^{26} Mg values.
739	To summarise, rapid overland flow transfers large amounts of soil into the river
740	during the monsoon season (Q>4000m ³ /s), with soil weathering playing an important
741	role in regulating river water chemistry and isotope values during these times (Fig. 9e
742	and 9f). Variability in $\delta^{44/40}$ Ca and 87 Sr/ 86 Sr values in the river waters are attributed to a

743 combination of secondary mineral dissolution of carbonates (and potentially

aluminosilicates) and exchange reactions (Fig. 9e and 9f).

745 6 Conclusions

746	In order to provide a better understanding of the effects of hydrological processes
747	on chemical weathering in large river systems, we studied temporal variations in river
748	water chemistry and 87 Sr/ 86 Sr, $\delta^{44/40}$ Ca, and δ^{26} Mg values in the Jinsha and Yalong river
749	waters. Temporal variations in the river water chemistry and isotopic compositions are
750	attributed to differences in the flow paths, source material being weathered, and
751	accompanying secondary processes under different hydrological conditions.

During the low-runoff period (Q<2000m³/s), meltwater seeps through the soil and into the groundwater that recharges the river. The chemical compositions of both the Jinsha and Yalong river waters are therefore governed by the chemical composition of the soil water and groundwater, which is affected by permafrost meltwater in the Jinsha river basin and by glacial meltwater in the Yalong river basin.

During the medium-runoff period (2000m³/s<Q<4000m³/s), rainwater interacts with bedrock and the contribution of bedrock chemical weathering (relative to soil water and groundwater inputs) to river water chemistry increases. During this period, the mainstream water chemistry and isotopic compositions are dominated by the inputs from the tributaries and headwaters. The evaporite-dominated headwater governs the water chemistry and isotopic compositions in the Jinsha River, whereas the Yalong River is governed by carbonate-dominated tributaries.

764 During the high-runoff period (Q>4000m³/s), heavy rainfall leads to rapid surface

flow, which transports a large volume of soil minerals into the Jinsha and Yalong river 765 waters. Those minerals include both secondary carbonates and clays, which regulate 766 river water chemistry through mineral dissolution and cation exchange. Dissolution of 767 amorphous aluminosilicates during heavy rainfall intervals could contribute to the 768 chemical stability of Si, which may in turn inhibit the weathering of primary silicate 769 minerals. In addition, Na ions can be released from soil-derived clays into river water 770 through cation exchange. Therefore, while silicate weathering fluxes are observed to 771 increase during the wet monsoon season, coeval increases in cation exchange could 772 773 result in an over-estimation of the magnitude of the increase in silicate weathering rate and its CO₂ consumption when conventional approaches are applied. In the context of 774 projected increases in the frequency of extreme precipitation due to global warming, 775 776 our observations suggest that the impact of hydrological events on chemical weathering fluxes and carbon cycling requires further consideration. 777

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793 Appendix A. Supplementary Material

Supplementary data (Figures S1-S4 and Tables S1-S8) to this article can be found 794 795 online. Fig. S1. shows magnesium three-isotope diagram for samples analysed in this study. Fig. S2. shows the forward model used to calculate the contribution of different 796 endmembers to the river waters. Fig. S3. shows relationships between discharge and a 797 range of chemical weathering parameters. Fig. S4. shows relationships of (a) dolomite 798 saturation index (DSI), (b) calcite saturation index (CSI), and (c) illite saturation index 799 (ISI) with discharge in the Jinsha (JS) and Yalong (YL) rivers. (d) Relationship between 800 δ^{26} Mg values and SO₄/Mg ratios. Table S1 shows detailed operating parameters for 801 stable Ca isotope ratio measurement using Nu Plasma HR MC-ICPMS in this study. 802 Table S2 shows stable Ca isotopic composition of geological reference materials from 803 this study. Table S3 shows ⁸⁷Sr/⁸⁶Sr values of geological reference materials from this 804 study. Table S4 shows detailed LOADEST program settings used for concentration and 805 flux calculations. Table S5 shows monthly average discharge and suspended particulate 806 matter content in the Jinsha and Yalong rivers. Table S6 shows contribution of different 807 808 endmembers to river water chemistry. Table S7 shows seasonal variations in chemical weathering rates and proportions for different lithologies. 809

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- 1186 1187

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1189 **Figure captions**

1190 Fig. 1. Setting and characteristics of the upper reaches of the Changjiang River. (a)

- 1191 Geological map with locations of the river water sampling stations marked by red stars
- 1192 (modified from (Chen et al., 2020)). (b) Permafrost distribution (modified from (Zhong
- 1193 et al., 2021)). (c) Air temperature (red line) and daily precipitation (blue bars) for the
- 1194 Yalong and Jinsha rivers from November 2013 to October 2014 (based on data from
- 1195 Panzhihua station).
- 1196

Fig. 2. Seasonal variations of discharge (blue shading) and elemental concentrations (Ca, Mg, Na, K, Sr, Al, Fe, F, Cl, SO₄) (coloured lines) calculated by LOADEST for (a,c) the Jinsha River and (b,d) the Yalong River. The purple bar in (a,c) represents the initial permafrost thaw season for the Jinsha River. The grey bar in (b,d) represents the initial snow melt season for the Yalong River.

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Fig. 3. Relationships between (a-d) selected elemental concentrations (Si, Sr, Ca, and Mg), (e) Si/Ca ratio, (f-h) isotopic compositions (Sr, Ca, Mg), and discharge during one hydrological year. In panels (a-d), the red and blue solid curves correspond to theoretical dilution lines for ion concentrations of the Jinsha and Yalong River, respectively. The vertical grey dashed lines represent discharge values of 2000 m³/s and 4000 m³/s, used to separate the low, medium, and high runoff regimes.

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Fig. 4. Seasonal changes in Si/Ca, 87 Sr/ 86 Sr, $\delta^{44/40}$ Ca, and δ^{26} Mg values, and discharge 1210 in (a) the Jinsha River and (b) the Yalong River. The purple shaded bar in (a) represents 1211 the initial thawing of permafrost. The yellow shaded bar in (b) represents the initial 1212 melting of glaciers and snow. The blue shaded bar in both panels represents the wet 1213 season. The red and grey symbols for the Sr isotope data in panel (a) represent measured 1214 values and values corrected for evaporite input, respectively. Error bars for Ca and Mg 1215 isotope data are shown in panel (a) and represent the external 2sd. For Sr isotope data, 1216 the 2sd on 87 Sr/ 86 Sr values is 0.00002, which is smaller than the symbol sizes. 1217

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Fig. 5. (a, b) Seasonal variations in carbonate (ω_{carb}), silicate (ω_{sil}), and evaporite (ω_{eva}) weathering proportions in (a) the Jinsha River and (b) the Yalong River. Discharge is also shown and the blue shaded bar in both panels represents the wet season.

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Fig. 6. Geochemical cross plots for Sr isotopes in the Jinsha and Yalong rivers. (a) 1224 ⁸⁷Sr/⁸⁶Sr versus Ca/Sr. (c) ⁸⁷Sr/⁸⁶Sr versus Na/Sr. Panels (b) and (d) correspond to panels 1225 1226 (a) and (c), but with samples separated according to the low, medium, and high runoff 1227 periods. The grey shaded area represents the mixing field between the main sources, including meltwater, evaporites, and soil solutions (inferred from leachates) (Table 2). 1228 In the legend, JS and YL represent the time series samples from the Jinsha and Yalong 1229 1230 rivers, respectively, while samples divided as 'High', 'Medium', and 'Low' are the time series samples collected during the high, medium, and low runoff periods. Data from 1231 the tributaries of the two rivers are also plotted (Chen et al., 2020; Chen et al., 2022). 1232

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Fig. 7. Geochemical cross plots for Ca isotopes in the Jinsha and Yalong rivers. (a) $\delta^{44/40}$ Ca versus 87 Sr/ 86 Sr; (b) $\delta^{44/40}$ Ca versus SO₄/Ca; (c) $\delta^{44/40}$ Ca versus Sr/Ca; (d) $\delta^{44/40}$ Ca versus water temperature (T). In panels (a-c), the grey shaded area represents the mixing field between meltwater, evaporites, and soil solutions (inferred from leachates). The red dashed lines represent best-fit lines for the Jinsha time series samples during the low runoff season. Symbols, abbreviations, and data sources are as in Figure 6.

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Fig. 8. Geochemical cross plots for Mg isotopes in the Jinsha and Yalong rivers. (a) 1242 δ^{26} Mg versus 87 Sr/ 86 Sr; (b) δ^{26} Mg versus $\delta^{44/40}$ Ca; (c) δ^{26} Mg versus illite saturation 1243 index (ISI); (d) δ^{26} Mg versus dolomite saturation index (DSI). Symbols, abbreviations, 1244 and data sources are as in Figure 6. In panel (c), the blue dashed line is the best-fit line 1245 1246 for the low runoff seasons of the Yalong river waters. 1247 Fig. 9. Schematic illustration of water sources and flow paths for the Jinsha and Yalong 1248 river basins during (a,b) the dry season with low runoff (Q<2000m3/s), (c,d) the 1249 intermediate runoff period (2000m3/s<Q<4000m3/s), and (e,f) the wet season with high 1250

1251 runoff (Q>4000m3/s).

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1255 **Tables**

1256 **Table 1.** Water chemistry and Sr, Ca, and Mg isotope data for the Jinsha and Yalong

1257 rivers at fixed sampling sites.

1258 **Table 2.** Chemistry and Sr, Ca, and Mg isotope data from soil and river bedload (bulk

1259 and leachates) and bedrock from the upper reaches of the Changjiang River basin.

1260



Fig. 1.



Fig. 2.



Fig. 3.



Fig. 4.



Fig. 5.



Fig. 6.



Fig. 7.



Fig. 8.



Fig. 9.

Supplementary information for

Behaviour of Sr, Ca, and Mg isotopes under variable hydrological conditions in highrelief large river systems

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Figure S1-S4:

Fig. S1. Magnesium three-isotope diagram for samples analysed in this study. Note that all samples fall on a single mass-dependent fractionation line.

Fig. S2. The forward model used to calculate the contribution of different endmembers to the river waters (modified from (Noh et al., 2009)). The detailed endmember values for all elements can be found in Table 2.

Fig. S3. Relationships between discharge and a range of chemical weathering parameters in the Jinsha and Yalong rivers. CWR, SWR, and EWR represent the weathering rates of carbonates, silicates, and evaporites, respectively. TCWR represents total chemical weathering rates, and E represents erosion rates. The parameters w_carb, w_sil, and w_eva represent the proportions of carbonate, silicate, and evaporite weathering to total chemical weathering.

Fig. S4. Relationships of (a) dolomite saturation index (DSI), (b) calcite saturation index (CSI), and (c) illite saturation index (ISI) with discharge in the Jinsha (JS) and Yalong (YL) rivers. (d) Relationship between δ^{26} Mg values and SO₄/Mg ratios.

Tables S1-S8:

 Table S1. Typical operating parameters for stable Ca isotope ratio measurement using Nu Plasma HR MC-ICPMS.

Table S2. Stable Ca isotopic composition of geological reference materials from this study.

Table S3. ⁸⁷Sr/⁸⁶Sr values of geological reference materials from this study.

Table S4. LOADEST program settings used for concentration and flux calculations.

Table S5. Monthly average discharge and suspended particulate matter content in the Jinsha and Yalong rivers.

Table S6. Contribution of different endmembers to river water chemistry.

Table S7. Seasonal variations in chemical weathering rates and proportions (w) for different lithologies.

Table S8. Mineral saturation indices calculated for Jinsha and Yalong river waters.



Fig. S1



Fig. S2



Fig. S3



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