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Climate-driven variation in biotic interactions provides a narrow and variable window of opportunity for an insect herbivore at its ecological margin

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Title: Climate-driven variation in biotic interactions provides a narrow and variable window of opportunity for an insect herbivore at its ecological margin

Running Head: Host availability shaped by seasonal variation

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Abstract

Climate-driven geographic range shifts have been associated with transitions between dietary specialism and generalism at range margins. The mechanisms underpinning these often transient niche breadth modifications are poorly known, but utilisation of novel resources likely depends on phenological synchrony between the consumer and resource. We use a climate-driven range and host shift by the butterfly *Aricia agestis* to test how climate-driven changes in host phenology and condition affect phenological synchrony, and consider implications for host use.

Our data suggest that the perennial plant which was the primary host before range expansion is a more reliable resource than the annual Geraniaceae upon which the butterfly has become specialised in newly colonised parts of its range. In particular, climate-driven phenological variation in the novel host *Geranium dissectum* generates a narrow and variable ‘window of opportunity’ for larval productivity in summer. Therefore, although climatic change may allow species to shift hosts and colonise novel environments, specialisation on phenologically-limited hosts may not persist at ecological margins as climate change continues. We highlight the potential role for phenological (a)synchrony in determining lability of consumer-resource associations at range margins, and the importance of considering causes of synchrony in biotic interactions when predicting range shifts.

Keywords: brown argus, Lepidoptera, host shift, specialisation, asynchrony

Introduction

Climate change is causing widespread shifts in species’ geographic range limits [1–3]. The extent of such shifts depends on species’ life histories, potential for plastic responses and the quality of available habitat at the expanding margin [4–6]. Habitat availability may itself be determined by the process of range expansion: range shifts have recently been identified as a cause, rather than consequence, of increased dietary generalism at poleward range margins in herbivorous insects [7,8]. However, dietary generalism and incorporation of novel host plants can be transient in such systems, and the mechanisms underlying gain or loss of hosts from insect diets is poorly known [7–9]. Here,

we highlight phenological (a)synchrony between insects and unpredictable host resources as a potential mechanism for lability in insect-host associations at range margins.

The phenology of many consumers is closely synchronised with the development and availability of their resources, and their interactions occur within an often narrow 'window of opportunity' for the consumer defined by the phenology of both partners [10–14]. For herbivorous insects, the length of the phenological window of opportunity will be partly determined by the specificity of its interaction with the host, the host's growth form, and the broader environmental context. For example, the window of opportunity for herbivory is typically longer on perennials than on short-lived annual plants, the availability of which may be defined by environmental drivers of their germination and senescence [15,16]. The window is also typically longer for polyphagous than obligately monophagous species (which can exploit fewer distinct phenological windows), and for populations inhabiting topographically variable landscapes in which heterogeneous microclimates provide diverse phenological windows [16,17]. Differences in synchrony among nearby microclimates may also cause local variation in host condition, quality and profitability for the herbivore, thereby influencing local patterns of host selection and opportunities for dietary change [7,8,18–20].

Robust evidence is therefore required on the drivers and vulnerability of host-herbivore phenological synchrony (which may scale up to emergent patterns of herbivore range and host shifts [10,19,21]), including the role of microclimate in consumer persistence by potentially buffering asynchrony in biotic interactions [18,20]. Here we address this knowledge gap, using as an exemplar the brown argus butterfly (*Aricia agestis*; Lepidoptera: Lycaenidae) at its range margin in the UK. In doing so, we highlight how phenological (a)synchrony could provide an underlying mechanism for apparently high rates of host shifting near range margins [7,22]. Brown argus butterflies have two generations per year; larval offspring of the second generation emerge from mid-August and feed on leaves of the host plant before overwintering [23]. The brown argus' UK range was historically largely restricted to calcareous grassland where its host, the perennial *Helianthemum nummularium* (Malvales: Cistaceae; hereafter *Helianthemum*), grows. However, since the 1990s, the brown argus has

undergone a climate-driven range expansion associated with rapid evolution of biotic interactions, including specialisation on annual Geraniaceae species (*Erodium cicutarium*, *Geranium dissectum* and *G. molle*; Geraniales: Geraniaceae) mainly in regions beyond the former range limit [21,24–27].

In this study, we test expectations of (a) greater temporal variation in the condition of the annual versus perennial host plants, and (b) more opportunity for asynchrony between the consumer and its annual hosts than its perennial host, as a consequence of (a). We expect asynchrony with the less-predictable hosts to be more pronounced under warm, dry summer conditions, and we test for such effects on the condition and phenology of the annual Geraniaceae hosts that have enabled the range expansion. We conduct these tests (c) across sites and years, and (d) across microclimates within a site. We then consider the implications of asynchrony for host use, shifting host associations at range margins, and the range dynamics of host-limited herbivores.

Methods

Study system

Brown argus butterflies prefer to lay their eggs on Geraniaceae, on which larvae grow 10% larger and faster, than on *Helianthemum* (the ancestral host at the range margin in Britain), and prefer to lay on better condition leaves regardless of host species [19,21,24–26,28,29]. However, the annual, more ephemeral growth form of the Geraniaceae host plants may make them less reliable as a food source than the evergreen perennial *Helianthemum*, especially under more variable climatic conditions at the range margin [16]. To investigate this, we surveyed ten sites fortnightly–monthly between July 2016 and October 2017, to monitor phenology and condition of three host plant species. *Helianthemum* was the dominant host at five calcareous grassland sites, while Geraniaceae (*G. dissectum* and *E. cicutarium*; hereafter *Geranium* and *Erodium*) were dominant at five grassland/ dune sites (Figure 1). The number of sites and quadrats was chosen to maximise spatial coverage and replication within logistical constraints. See SI-1.1 and SI-1.2 for survey dates and site profiles.

116 **Quadrat surveys and analyses**

117 Using an average of 33.7 (0.25 m²) quadrats per survey (SI-1.3 for quadrat
 118 placement), we recorded percentage cover (SI-1.3.2), phenophase
 119 (phenological stage) and condition of each host species. Host phenophase and
 120 condition underpin brown argus egg-laying site choice in the field [19,30].
 121 Phenophase was estimated on a four-point scale (in leaf L, bud B, flower F, or
 122 had set seed S). Condition was visually assessed on a scale of 0–3 (poor–high
 123 quality for egg-laying, following [19,26]; see SI-1.3.3 for details and justification).
 124 We also measured mean sward height and percentage cover of bare ground
 125 (SI-1.3) which can alter local microclimates [31,32].

126 *Quadrat-level* phenophase and condition were estimated based on the average
 127 for each host species in each quadrat. This approach can mask fine-scale
 128 changes in phenophase and host condition, particularly for the annual hosts, so
 129 we also recorded *plant-level* phenophase and condition of the smallest, earliest
 130 phenophase Geraniaceae plant in each quadrat. This approach allows
 131 inference as to whether new germination has occurred, and of the age/condition
 132 of the plant material likely to be available to overwintering larvae. By
 133 September, Geraniaceae plants in later phenophases are typically senescent, in
 134 poor condition and expected to die before the autumn, making them a poor
 135 resource for larvae [30].

136 **Quantifying variation in host plant condition and phenophase**

137 To test expectation (a), for greater temporal variation in the condition of the
 138 annual than perennial hosts, we used a Kruskal-Wallis (KW) test for each
 139 sampling period to compare quadrat-level condition between host species.
 140 Bonferroni-corrected Dunn's tests were then used to identify which hosts
 141 differed significantly from one another in condition score within each sampling
 142 period. We then conducted interannual comparisons of Geraniaceae host
 143 condition (Mann-Whitney U tests, MWU) and phenophase (X^2 tests): these
 144 compared 2016 data with 2017 data, separately for each month between July to
 145 October. These months are the most relevant for host choice and larval feeding
 146 by second generation brown argus and their offspring.

147 **Assessing plant-herbivore (a)synchrony**

148 To contextualise host condition with reference to herbivore phenology, and
 149 address question (b), we overlaid plots of site-specific host condition indices
 150 (calculated using survey data, SI-1.5.1) from all July–October surveys with
 151 emergence phenology curves of adult second generation brown argus and their
 152 larval offspring. Adult phenology was described as site-specific Gaussian
 153 curves for 2016 and 2017, based on output of phenomenological models
 154 following [30,33] (below and SI-1.6). The larval phenology curves track the adult
 155 curves, with an estimated 11-day lag to account for mating and egg-laying (four
 156 days post-emergence) and larval emergence (one week) [24]. Therefore, the
 157 larval emergence curves are presented as indicators rather than precise
 158 evaluations of appearance or abundance at each site.

159 The overlap between plotted condition indices and brown argus phenology
 160 curves was used to generate an area under the curve (AUC) metric of site- and
 161 year-specific synchrony between brown argus and the host plants (SI-1.7).
 162 These AUC metrics were then modelled in a beta regression (logit link; SI-1.7)
 163 to test for effects of site latitude, host plant, year and a host-year interaction on
 164 synchrony.

165 The phenomenological models used to generate phenology estimates account
 166 for variation in phenology between sites and years based on differences in
 167 latitude and temperature (SI-1.6). In summary, brown argus second brood
 168 phenology in Britain varies with latitude (earlier further north) and between-
 169 brood temperature (earlier under warmer conditions between the first and
 170 second brood), and is related to the (latitude-dependent) phenology of the first
 171 brood [30,33]. See SI-1.6 for more information.

172 **Testing climatic predictors of Geraniaceae recruitment and condition**

173 To address our third question (c), using data on the youngest host within each
 174 quadrat, we tested for climatic drivers of recruitment and condition of each
 175 Geraniaceae species in early September, when most larval offspring were
 176 expected to have emerged to feed. We defined recruitment as the presence of
 177 at least one young, leaf-stage host plant in the focal quadrat. Plants in condition
 178 categories 1 and 2 were rarely observed during September surveys. We

179 therefore reclassified condition 0/1 plants as poor condition (0) and condition
180 2/3 plants as good condition (1).

181 To test predictors of recruitment and condition, we used logistic regression with
182 the following putative predictors: year, site, northing, easting, vegetation height,
183 bare ground cover, and (linear and quadratic terms for) local weather estimates
184 based on the UK Meteorological Office's 5 km gridded weather data [34]. Using
185 daily weather data [34], we calculated the minimum, mean and maximum
186 temperature and mean rainfall for three periods in each year (justified in SI-1.7):
187 July, July–September and August–September (only including weather data up
188 to the day of quadrat sampling at each site in early September). We also used
189 daily weather data to calculate the Gaussen Aridity Index (GAI;
190 precipitation/(2 × temperature)) for each period (e.g. [35]). Higher GAI values
191 indicate cooler, wetter conditions. Day of year was tested both as a putative
192 fixed effect predictor and as an offset term to account for day of sampling.

193 All continuous predictors were standardised and site ($n \leq 5$ per host plant) was
194 included as a fixed effect (following [36]). We constructed candidate models by
195 considering all plausible parameter combinations (including temperature-rainfall
196 interactions), estimated parameters using maximum likelihood, and used AIC-
197 based model selection to establish the most parsimonious model(s) (see
198 SI-1.9–1.10 for details of model selection, validation and diagnostics).

199 **Testing microclimate effects on Geraniaceae phenology and condition**

200 In September 2017, we calculated condition and phenophase indices (SI-1.5.2)
201 for each of 31 quadrats at site G1, using the phenophase (L, B, F and S) and
202 condition (0–3) of all *Geranium* plants in each quadrat. The indices range
203 between 0 (quadrats contain only plants at condition 0/seed set stage) and 1
204 (plants at condition 3/leaf stage). To test expectation (d), we modelled the
205 indices (logistic regression) as a function of putative quadrat-specific predictors:
206 microclimate (mean, maximum and minimum temperature and soil moisture;
207 SI-1.11), percentage cover of bare ground and mean sward height. We also
208 considered plausible temperature-moisture interactions.

209 **Results**

210 **Variation in host plant condition and phenophase**

211 There was a substantial decline in quadrat-level condition of the annual host
 212 *Geranium* over summer 2016 that was not observed in the perennial
 213 *Helianthemum* or to the same extent in *Erodium* (Figure 2a; KW tests, Table
 214 S4). However, by early November 2016, senesced *Geranium* had mostly been
 215 replaced by recently-germinated, better condition recruits, and the quadrat-level
 216 condition was at least as high as that of *Helianthemum* (Figure 2a-c; KW tests,
 217 Table S4).

218 Overall, in summer 2016 both Geraniaceae species showed substantial
 219 evidence of leaf senescence, including wilting and abscission, and little
 220 germination or seedling establishment until early October (*pers. obs.*)
 221 (*Geranium*, Figure 2b-c; *Erodium*, Figure S3). By contrast, in 2017, there were
 222 many more seedlings and good condition plants throughout the summer months
 223 and into the main sampling period in September (*pers. obs.*) (*Geranium*, Figure
 224 2b-c; *Erodium*, Figure S3).

225 The condition of the youngest *Geranium* in each quadrat was significantly
 226 higher in all 2017 survey periods compared to 2016 (Figure 2b; MWU tests,
 227 Table S5). *Geranium* phenophases differed significantly between years (X^2
 228 tests, Table S5): in 2016, the youngest plant in each *Geranium* quadrat was
 229 typically an older plant in the seed set stage (Figure 2c) and new plants in the
 230 leaf stage did not dominate until October; however, this younger form was
 231 dominant throughout summer 2017 (Figure 2c).

232 *Erodium* showed similar patterns, though condition was approximately
 233 equivalent between years for the August and September surveys (Figure S3;
 234 Mann-Whitney U tests, Table S5). *Erodium* quadrats were also dominated by
 235 young, leaf-stage plants in 2017: significantly more so than in 2016 during July
 236 and August (Figure S3; X^2 tests, Table S5). Similar patterns of Geraniaceae
 237 phenophase and condition were observed at quadrat-level as these plant-level
 238 assessments (Figures S4 and S5).

239 **Plant-herbivore (a)synchrony**

240 We assessed potential for asynchrony between the butterfly and its hosts
 241 (question (b)) by overlaying plots of site-specific host condition indices with
 242 curves representing brown argus phenology (Figure 3), and performing beta
 243 regressions on derived synchrony estimates. The beta regressions of AUC
 244 synchrony estimates demonstrate that synchrony was lowest for brown argus
 245 on *Geranium* in 2016 (low AUC overlap: AUC range 0.51–0.73), but very high
 246 for other hosts and for *Geranium* in 2017 (high AUC overlap: AUC range 0.92–
 247 1.00) (Figure 3, Tables 1 and S6). Synchrony was lower for larvae than adults,
 248 especially on *Geranium* in 2016 (Table S6). There was no detectable effect of
 249 site latitude on adult or larval AUC overlap (SI-1.12.4).

250 In 2016, *Geranium* condition declined over the peak of adult brown argus
 251 emergence, and was lowest during the period in which most larvae would be
 252 beginning to feed (Figure 3a–c). By contrast, at their respective sites, good-
 253 condition *Erodium* and *Helianthemum* hosts were available throughout egg-
 254 laying and early larval feeding periods of the brown argus butterfly, with little
 255 variation between two climatically different years (2016 and 2017), or among
 256 sites at each sampling period (Figure 3d–j). Among-site variation in condition is
 257 more pronounced for *Geranium*, but this does not mask the temporal variation
 258 within and between years (Figure 3; Table S4). Among-site variation likely
 259 results from local variation in temperature and water relations linked to factors
 260 including weather, topography and geology.

261 **Climatic predictors of Geraniaceae condition and recruitment**

262 We integrated climate data with in-field host plant surveys to investigate
 263 potential drivers of condition and phenology across the *Geranium* sites,
 264 addressing question (c). The probability of the youngest *Geranium* plants being
 265 in good condition increased with summer rainfall (MC_{final} ; Figure 4a, Table 2a).
 266 There was limited evidence that moister, cooler conditions in areas with shorter
 267 vegetation were associated with better condition (Table S8). Forcing models of
 268 *Geranium* condition to include an effect of year resulted in higher AIC values,
 269 and inflated parameter estimates and standard errors by several orders of
 270 magnitude, so these are reported only in the Supplementary Information for
 271 context (SI-1.12.5, Table S8), and there are no effects of year in the final model

set. The evidence suggests that probability of new *Geranium* recruitment was higher in 2017 ($MR_{best,i}$; SI-1.12.5, Table S9), and lower following higher mean daily temperatures during August–September ($MR_{final,i}$; Figure 4b, Table 2b). There were no detectable effects of site northing or easting, or day of sampling on condition or recruitment.

We had low statistical power to detect relationships between weather and the condition and recruitment of *Erodium*, models for which are outlined in SI-1.12.5.

Microclimate effects on Geraniaceae phenology and condition

To address question (d), we assessed *Geranium* phenology and condition across a range of microclimates at site G1 in September 2017. At this site, areas of moister soil (where plants are less likely to dry out) were associated with *Geranium* plants in better condition and earlier phenophases (Figure 4c,d; Table 3). Candidate models suggested such plants were also more prevalent in areas with warmer, moister microclimates (SI-1.12.6).

Discussion

Here, we describe temporal variation in condition and phenology of three plant species and highlight the implications for their use as larval host plants by a butterfly species that has recently expanded its geographic range. Our data, from multiple sites across two years, suggest that the annual host *Geranium dissectum* varies more in condition and availability than both *Erodium cicutarium* and *Helianthemum nummularium*, the species that have been used as long-standing hosts at the range margin, and does so in a way that differs between years and with (micro-)climatic conditions. Such variation in condition and availability likely generates narrow and unpredictable phenological ‘windows of opportunity’ for exploitation of ephemeral annual species that vary among sites and years under the conditions of variable population sizes or phenology observed near the limits of species’ geographic ranges [37]. Though we recognise differences in the hosts other than their perennation strategy, in this case the perennial plant that was used as the main pre-expansion host appears to be a more reliable resource, where present, than the more widespread annual *Geranium* species that has acted as a primary host during

the climate-associated range expansion. *Erodium*, is typically found on sandy soils, is relatively drought-tolerant compared to other Geraniaceae hosts, and appears to respond to a wider range of phenological cues than *Geranium*, which may improve its relative condition and availability as a host under the conditions we observed [30]. Our results suggest that climatic effects on the phenological synchrony of biotic interactions could act as a mechanism generating transient patterns of host associations at range margins, and consequently of habitat availability in the landscape and patterns of range shifting.

(Micro-)climatic variation and host plant phenology

Our data indicate that the phenology of *Geranium*, a widespread annual host plant used by the brown argus butterfly, is sensitive to weather variability. Greenhouse experiments and field observations suggest that summer temperatures and moisture thresholds are crucial to dormancy breaking and germination in *Geranium* [38,39]. A complementary interpretation of these data is that *Geranium* plants may germinate early following a cool spell and early summer rain, but will suffer high seedling mortality where the summer is subsequently hot and dry (e.g. [40,41]). For example, our data show that hot, dry conditions in 2016 were associated with early and pronounced senescence of plants in July, as well as delayed germination and/or early seedling mortality. By contrast, our study sites received relatively high rainfall throughout summer 2017 [30], which is likely to have overcome moisture-dependency in dormancy breaking and/or promoted seedling survival. *Geranium* condition was also higher following wetter (2017) summer conditions, which supports evidence that drought and thermal stress cause premature senescence and declines in the quality (for consumers) of herbaceous plants [42,43]. A higher proportion of younger and better-condition *Erodium* were available in 2017 than in 2016, although we lacked statistical power to associate this with climatic variables (SI-1.12.5). However, our data support previous observations that reduced soil water availability advances the reproductive stage and the end of the growing season in *Erodium*, and reduce its investment in leaf biomass [44,45].

335 **Trophic interactions at range margins**

336 Host condition and phenological synchrony in biotic interactions appear to be
 337 mediated by local (micro-)climatic variation, and are major determinants of
 338 spatiotemporal variation in fecundity and population size of host-specialist
 339 herbivores such as the brown argus [19,22,26,46,47]. By reducing temporal
 340 overlap between suitable resources and key herbivore life stages (e.g. adult
 341 egg-laying and early larval stages), adverse (micro-)climatic conditions may limit
 342 egg-laying and feeding opportunities and reduce larval survival, particularly
 343 where plants with limited temporal availability (such as the annual *Geraniaceae*
 344 studied here) are the main hosts [18,48]. Climatic conditions that are set to
 345 become more common (i.e. variable rainfall and longer, hotter summers
 346 [49,50]), may therefore narrow or close the phenological window of opportunity
 347 for this host-specialist herbivore to exploit these ephemeral annual resources in
 348 late summer [46,47].

349 Our analysis was unable to detect a clear geographic gradient in (a)synchrony,
 350 which may limit the predictability of the window of opportunity for brown argus to
 351 interact with *Geraniaceae* across its range. The widespread annual
 352 *Geraniaceae* are the hosts primarily used by brown argus at the expanding front
 353 of its recent range in Britain, whereas *Helianthemum* represents the apparent
 354 ancestral host at most sites where the species has been present for the past
 355 century or longer [26,30]. Therefore, while there is variation among sites,
 356 variation in butterfly-host synchrony (and the success of this relatively novel
 357 interaction) may be especially pronounced and unpredictable near the range
 358 limits, particularly if abiotic conditions are marginal and population sizes small
 359 [22]. In this respect, our results suggest that the range limit may be set not via
 360 broad environmental gradients in synchrony, but via shifting availability of sites
 361 where herbivore-host synchrony is sufficient. Through a process of ecological
 362 fitting at the novel range margin, populations interact with the resources that
 363 they happen to be synchronous with [51].

364 The relationships between weather, plant condition and phenology, which we
 365 identify here for *Geranium*, are therefore crucial in mediating how climate
 366 change, variability and unpredictability will affect synchrony in biotic
 367 interactions. Given that climate change is causing widespread changes in

phenological synchrony, both existing and novel host interactions may be vulnerable under climate change. However, recent evidence suggests that at least as many biotic interactions are becoming more synchronous as are becoming less synchronous [22]. This evidence therefore also highlights the potential for novel biotic interactions to emerge, and further supports a role for spatiotemporal variation in synchrony underlying transient host interactions.

Spatiotemporal variation in the synchrony of biotic interactions is likely to generate transient mosaics of selection pressures for different diets, and thereby influence patterns of dietary specialisation/generalisation that recent research has shown to be an emergent and surprisingly common property of range dynamics [7,8,21,24]. In range expansions, many new consumer-resource interactions form and some may be lost [7,8]. For example, diet breadths in populations of Edith's checkerspot butterfly (*Euphydryas editha*) increased after colonisation events as individual host preferences diversified, but populations subsequently reverted to monophagy [8].

Spatiotemporal variation in phenological synchrony may prove to be an underlying mechanism not only for lability in insect-host associations at range margins, but also for range limit stability under scenarios of temporal environmental variability. In particular, existing phenological plasticity can increase fitness costs where environments become more unpredictable [52], and range limits are more stable (expansions less likely) where environmental variance is too large for adaptation and colonisation [53]. Therefore, increasing environmental variance may preclude colonisation events that depend on predictably synchronous biotic interactions.

Evolution during range expansion

Following the recent range expansion and incorporation of Geraniaceae into the diet of the brown argus, Geraniaceae-feeding populations that were able to persist became specialised on the novel hosts, losing the adaptive capacity to use *Helianthemum* [21,24]. Though specialisation on Geraniaceae appears to have become more reliable on average [29], our data suggest that interannual variation in phenology could still alter the success of this interaction, which may prove to be locally transient in the face of phenological asynchrony. In a comparable example from Scandinavia, narrow oviposition preferences of the

Glanville fritillary (*Melitaea cinxia*) for phenologically-limited hosts risks high larval mortality under severe drought conditions in some years [20,54].

Conversely, phenological asynchrony at recently colonised Geraniaceae sites could lead to variable population dynamics, and selection favouring continued dispersal from the natal site in search of suitable egg-laying locations [55,56]. During the range expansion, selection has apparently favoured more dispersive phenotypes which have increased flight capacity and more readily accept the geographically widespread Geraniaceae hosts [24,27,28]. Consequently, under certain (micro-)climatic conditions, the dispersive, Geraniaceae-favouring phenotype may represent an alternative life history strategy that drives expansion at range margins and in-filling of the core range. Subsequent migrants that colonise *Helianthemum* sites may need to regain the ability to use *Helianthemum* (as shown in [21]; cf. [8]) in order to benefit from stability of (and phenological synchrony with) the host resource.

Conservation and management implications

Understanding constraints and opportunities for species' distributions is central to successful conservation practices. Our results highlight the importance of considering drivers of synchrony and the outcomes of biotic interactions when examining climate-driven range shifts, and recognising the crucial roles of microclimate and individual behaviour in mediating these interactions [19]. Conservation strategies could seek to maximise habitat and microclimatic heterogeneity to promote diversity in local phenologies across trophic levels [57,58]. In some cases, microclimatic variation may generate sufficient fine-scale spatial heterogeneity in relative phenology and host condition to buffer local herbivore populations against phenological asynchrony [10,18,59,60]. However, to improve our predictions of ecological responses to climate change, and of the critical levels of environmental change likely to cause rapid loss of ecosystem outputs, more empirical data are needed on shifts in biotic interactions across populations, climates and species ranges, their effects on demography, and their rates and patterns of evolution. In the present case, we have highlighted what might be typical variation in phenological synchrony across two years, which emphasises the potentially large indirect impacts of climate change on herbivore success. However, it would be beneficial to

expand sampling across both time and space to better understand general patterns of interannual variation and gain a more holistic understanding of the phenomena discussed.

Conclusions

Our results place a novel emphasis on the interactions between phenology, resource use and climate change in a range-expanding herbivore whose sensitivity to small changes in temperature might otherwise predict a positive range expansion response under future climates. Instead, small changes in temperature and moisture regime have the potential to disrupt the range expansion, mediated by phenological and physiological changes in the herbivore's larval resource that essentially fragment the herbivore's potential range. This suggests that novel host interactions may only remain as transient resources at shifting range margins. The mechanisms underlying insect-plant interactions, and their responses to climate change, are likely to be more complex than they appear, and we need more detailed knowledge of these mechanisms to understand and predict species' interactions and responses to climate change.

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Table 1. Parameter estimates (with errors) for Bayesian beta regression of adult and larval AUC synchrony metrics, also showing the lower (L95) and upper (U95) credible intervals for each estimate. Predictors include year, host and host-year interaction.

	Adults			Larvae		
	Estimate	L95	U95	Estimate	L95	U95
Intercept	2.89 (0.44)	2.03	3.77	2.8 (0.4)	1.99	3.55
Year₂₀₁₇	1.61 (0.65)	0.38	2.93	1.69 (0.65)	0.46	3.01
Host_{Geranium}	-2.01 (0.48)	-2.94	-1.07	-2.23 (0.45)	-3.07	-1.32
Host_{Helianthemum}	0.02 (0.48)	-0.94	0.97	0.01 (0.37)	-0.7	0.73
Year₂₀₁₇:Host_{Geranium}	2.87 (0.73)	1.43	4.29	3.06 (0.74)	1.58	4.5

Estimates for *Erodium* and Year₂₀₁₆ are not shown here as they are the base level for each predictor and are therefore included in the intercept term.

Table 2. Parameter estimates (with standard errors) for the final ($_{final}$; selected, most parsimonious) and null ($_N$) logistic regression models investigating drivers of (a) the condition of the youngest *Geranium* (MC), and (b) the probability of new recruitment of *Geranium* (MR) in each quadrat during September surveys (2016–2017). Also presented for comparison are the restricted models (MX_R) used to check for class bias in MX_{final}. All models bar the null contain a subset of fixed effects from: site (S, using sum contrasts: see SI-1.9), daily mean rainfall (R) or daily mean temperature (T) in the period between August 1st–September sampling date. Other terms and time periods were tested and detected in the final candidate set of models (Table S8 and S9). β_0 is the intercept, which accounts for the mean of site effects in all but the null models, k is the number of parameters, LL is the log-likelihood of the model and ΔAIC is the ΔAIC relative to the model with the lowest AIC in each case (Table S8 and S9).

Model	Model parameters							
	β_0	S_1	S_2	R	T	k	LL	ΔAIC
(a) Condition of youngest <i>G. dissectum</i>								
MC_{final}	2.053 (0.535)	5.749 (0.985)	-1.590 (0.551)	5.871 (0.891)	–	4	-61.17	0.13
MC_N	0.545 (0.141)	–	–	–	–	1	-143.30	168.38
MC_R	1.565 (0.622)	5.071 (1.116)	-1.305 (0.630)	5.611 (1.128)	–	4	*	*
(b) Probability of <i>G. dissectum</i> recruitment								
MR_{final}	0.310 (0.354)	4.366 (0.794)	-3.558 (0.718)	–	-3.872 (0.555)	4	-49.07	1.43
MR_N	0.352 (0.138)	–	–	–	–	1	-147.78	192.84
MR_R	-0.044 (0.495)	4.770 (1.126)	-4.120 (1.017)	–	-4.096 (0.792)	4	*	*

Table 3. Parameter estimates (with standard errors) for the final ($_{final}$; selected, most parsimonious) and null ($_N$) logistic regression models investigating microclimatic drivers of (a) quadrat-specific condition indices (MCI), and (b) quadrat-specific phenophase indices (MPI) of *Geranium* at site G1 in September 2017. Also presented for comparison are the associated null models (MX_N). All models bar MX_N contain a subset of fixed effects from: bare ground cover (G), soil moisture (M), mean (T) and minimum temperature (t). Other terms were tested and detected in the final candidate sets of models (Tables S9–10). β_0 is the intercept, k is the number of parameters, LL is the log-likelihood of the model and ΔAIC is the ΔAIC relative to the model with the lowest AIC in each case (Tables S9–10).

Model	Model parameters					k	LL	ΔAIC
	β_0	G	M	T	t			
(a) Quadrat-specific condition indices								
MCI_{final}	0.496 (0.413)	—	0.810 (0.448)	—	—	2	-15.59	1.88
MCI_N	0.428 (0.380)	—	—	—	—	1	-19.69	8.07
(b) Quadrat-specific phenophase indices								
MPI_{final}	0.366 (0.412)	—	0.889 (0.453)	—	—	2	-14.84	1.57
MPI_N	0.307 (0.376)	—	—	—	—	1	-19.98	9.85

Figure Captions

Figure 1. Map of study sites in England. + denotes *Helianthemum* sites (H1–H5, Table S1), × denotes Geraniaceae sites (*Geranium*, G1–G3; *Erodium*, E1–E2; Table S1).

Figure 2. (a) Mean (\pm SD) of site-averaged quadrat-level host plant condition categorisations (0–3). Starred brackets indicate significant differences in condition between hosts (Table S4). Condition (0–3; b) and phenophase (c) of the youngest *G. dissectum* plant in each quadrat across all sites visited during late July–early October in both 2016 and 2017. Phenophase L: leaf; B: in bud; F: in flower; S: set seed. Equivalents of (b) and (c) for *Erodium* are available as Figure S3.

Figure 3. Plant condition index in 2016 and 2017 for *Geranium* at sites G1–G3 (a–c), *Erodium* at sites E1–E2 (d–e), and *Helianthemum* at sites H1–H5 (f–j), indicating timing of host condition changes relative to the year- and site-specific emergence of second generation brown argus adults and their larvae. Butterfly phenology curves typically overlap at each site. (k) summarises larval AUC synchrony metrics for each host-year combination, summarised from site-specific metrics each calculated as the full area of the phenology curve minus that which lies above the corresponding host condition line. The equivalent plot for adults is shown in SI-1.12.4.

Figure 4. (a,b) Predicted probabilities with 95% confidence intervals (dashed lines) of (a) a state of good condition in the youngest *Geranium* in each quadrat, and (b) recruitment of *Geranium* in each quadrat. The probabilities (a) increase as a function of late summer rainfall and (b) decrease as a function of late summer temperature (MR_{final} , Table 1). Some points offset in x and y planes to show the raw data; mean daily rainfall ranges between -1.76 and 2.18 (1.29–2.87 mm), and temperature ranges between -1.16 and 1.85 (15.61–18.11 °C). The condition (c) and phenophase (d) indices of *Geranium* at site G1 increase as a function of soil moisture. Quadrats with higher soil moisture are more likely to contain high proportions of good condition new recruits, and low proportions of poor-condition, reproductive *Geranium*. Soil moisture ranges between -1.58 and 1.83 (14.6–34.4 %).

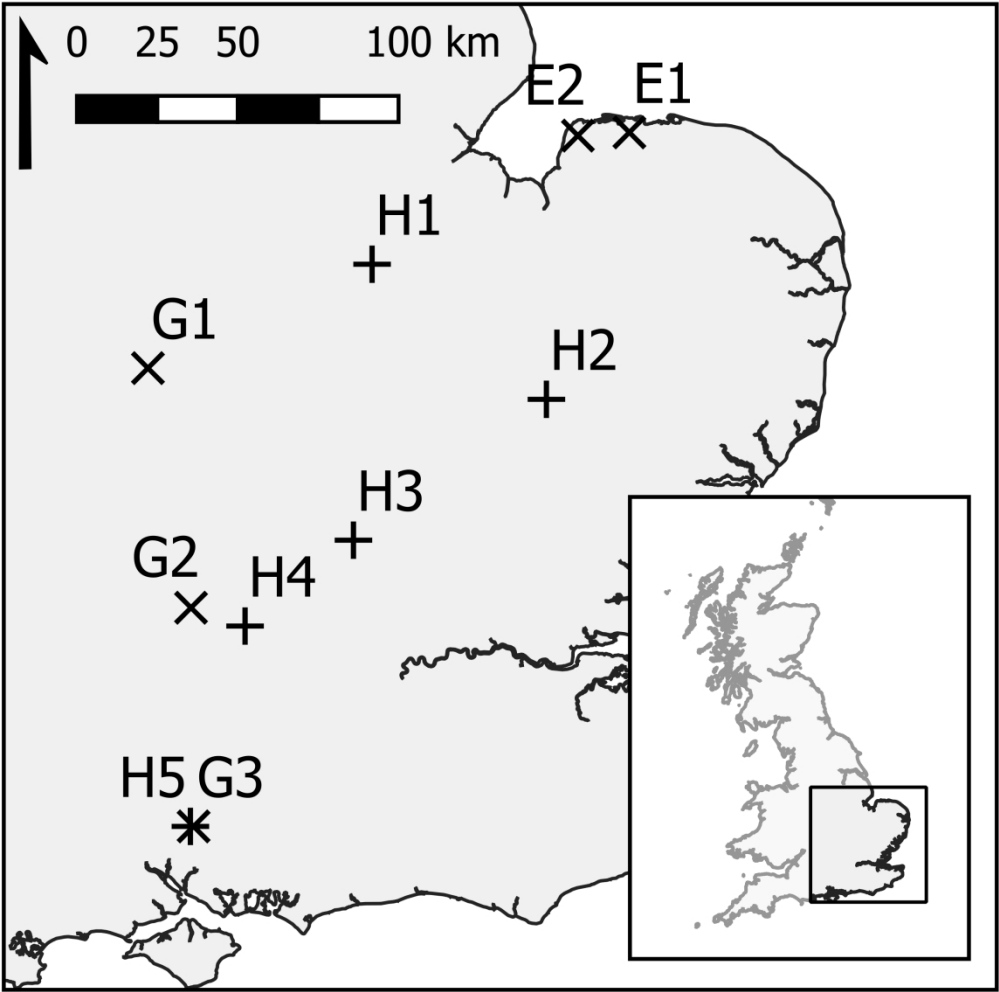


Figure 1. Map of study sites in England. + denotes *Helianthemum* sites (H1–H5, Table S1), x denotes Geraniaceae sites (*Geranium*, G1–G3; *Erodium*, E1–E2; Table S1).

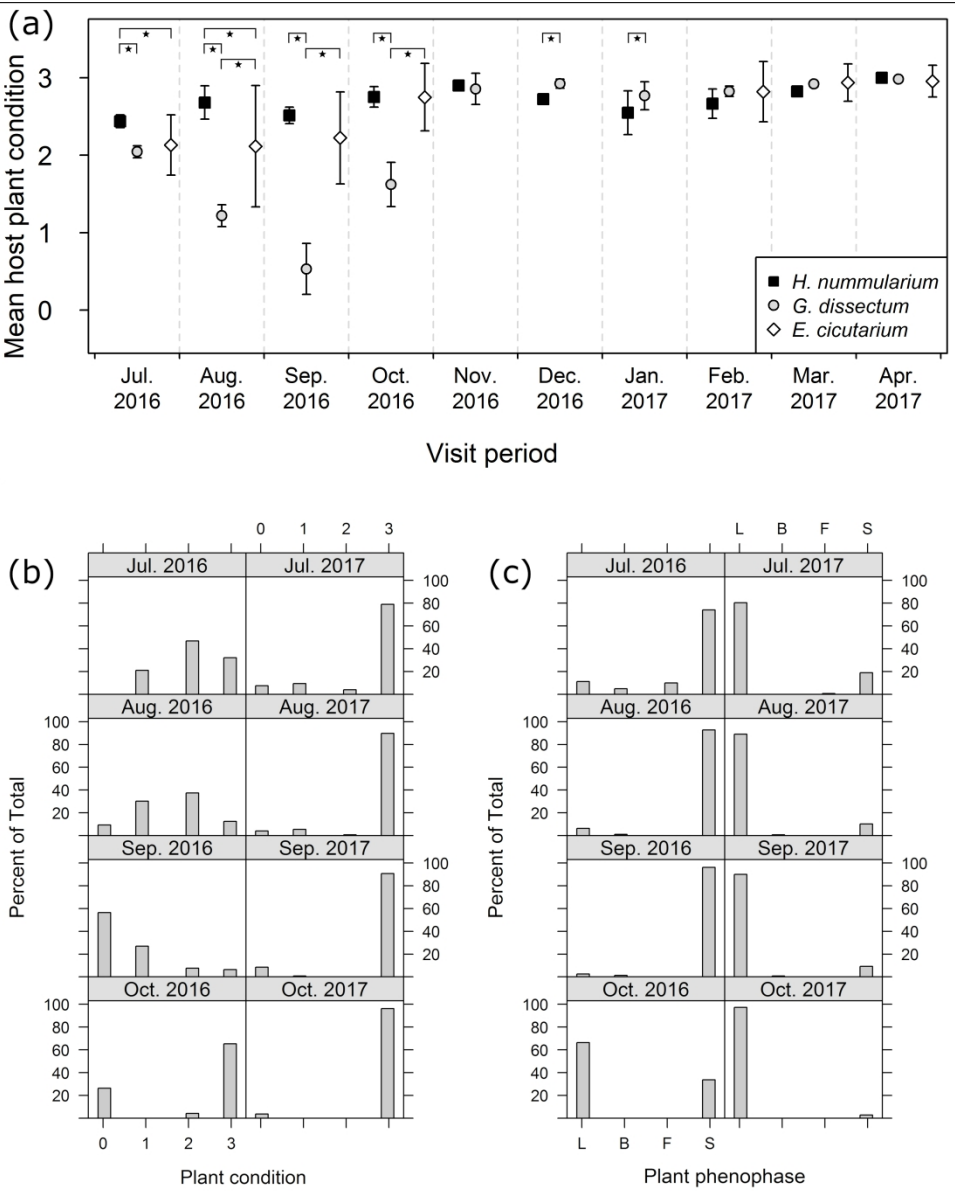


Figure 2. (a) Mean (\pm SD) of site-averaged quadrat-level host plant condition categorisations (0–3). Starred brackets indicate significant differences in condition between hosts (Table S43.1, Appendix 3). Condition (0–3; b) and phenophase (c) of the youngest *G. dissectum* plant in each quadrat across all sites visited during late July–early October in both 2016 and 2017. Phenophase L: leaf; B: in bud; F: in flower; S: set seed. Equivalents of (b) and (c) for *Erodium* are available as Figure S3.

152x190mm (600 x 600 DPI)

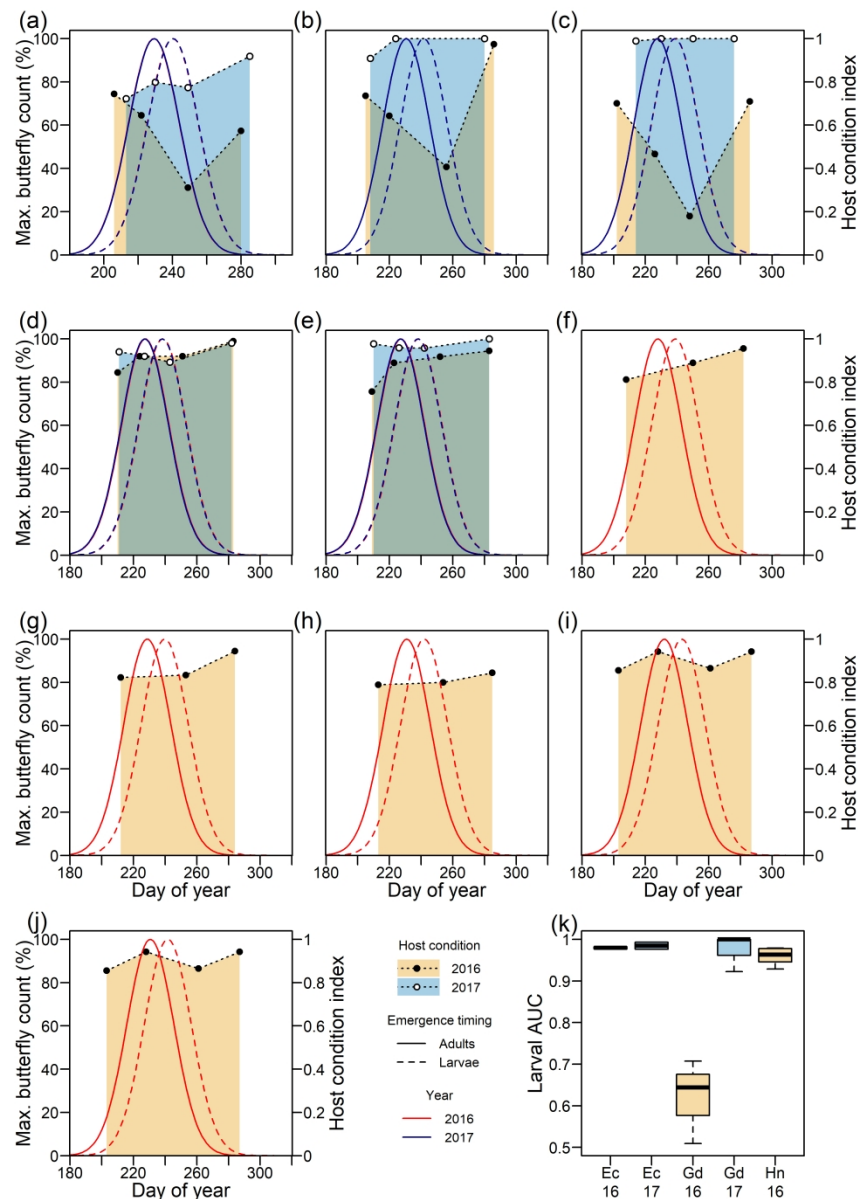


Figure 3. Plant condition index in 2016 and 2017 for *Geranium* at sites G1–G3 (a–c), *Erodium* at sites E1–E2 (d–e), and *Helianthemum* at sites H1–H5 (f–j), indicating timing of host condition changes relative to the year- and site-specific emergence of second generation brown argus adults and their larvae. Butterfly phenology curves typically overlap at each site. (k) summarises larval AUC synchrony metrics for each host-year combination, summarised from site-specific metrics each calculated as the full area of the phenology curve minus that which lies above the corresponding host condition line. The equivalent plot for adults is shown in SI-1.12.4.

152x215mm (600 x 600 DPI)

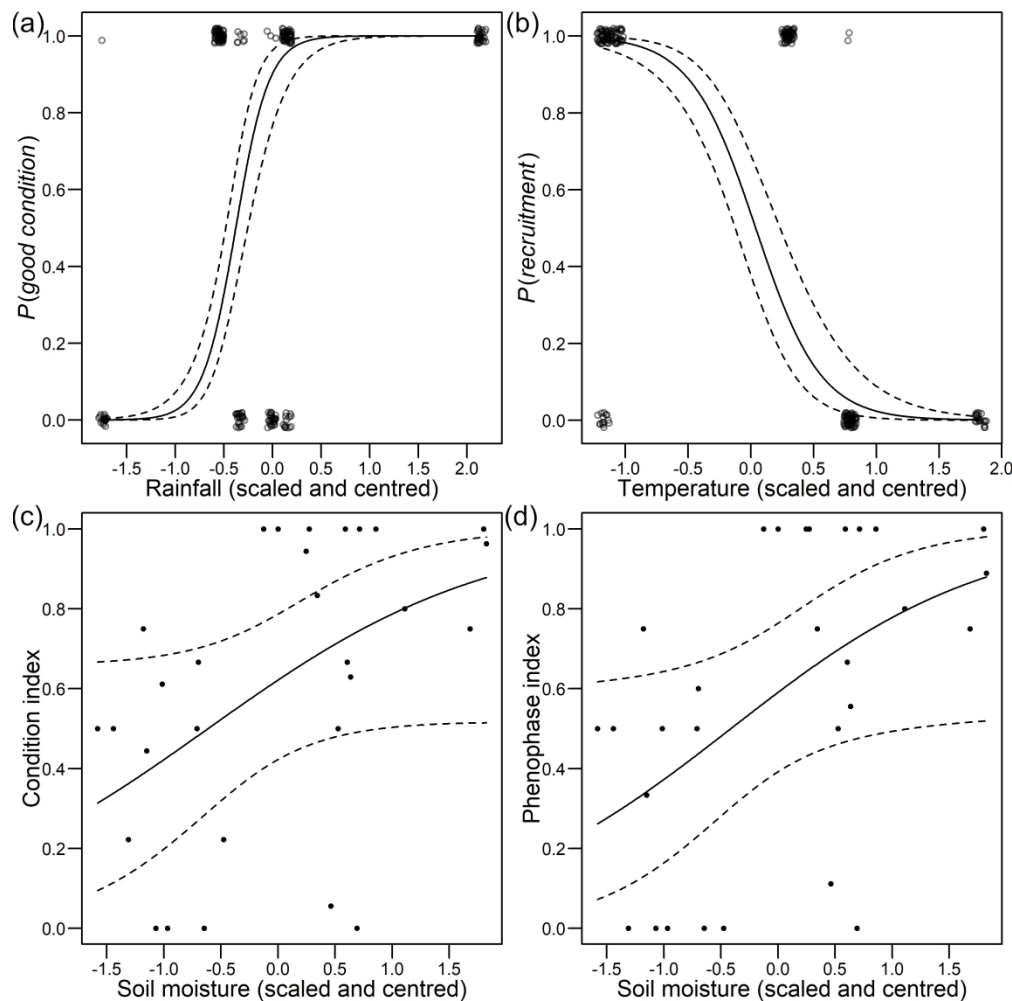


Figure 4. (a,b) Predicted probabilities with 95% confidence intervals (dashed lines) of (a) a state of good condition ($q = 2$ or 3) in the youngest *Geranium* in each quadrat, and (b) recruitment of *Geranium* in each quadrat. The probabilities (a) increase as a function of late summer rainfall (MR_{final} ; Table 1) and (b) decrease as a function of late summer temperature (Table 1). Some points have been offset in the x and y planes to show the raw data; mean daily rainfall ranges between -1.76 and 2.18 (1.29 mm– 2.87 mm), and temperature ranges between -1.16 and 1.85 (15.61 °C– 18.11 °C). The condition (c) and phenophase (d) indices of *Geranium* at site G1 increase as a function of soil moisture. Quadrats with higher soil moisture are more likely to contain high proportions of good condition new recruits, and low proportions of poor-condition, reproductive *Geranium*. Soil moisture ranges between -1.58 and 1.83 (14.6 – 34.4 %). Soil moisture was recorded at a different time of year to plant condition, so is used as a relative measure between quadrats.

152x152mm (600 x 600 DPI)