

1 **Global sulfur dioxide emissions and the driving forces**

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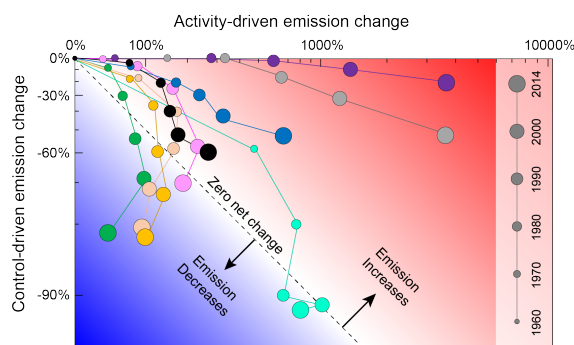
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14 **TOC Art**



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16

17 **Abstract**

18 The presence of sulfur dioxide (SO<sub>2</sub>) in the air is a global concern because of its severe environmental and  
19 public health impacts. Recent evidence from satellite observations shows fast changes in the spatial  
20 distribution of global SO<sub>2</sub> emissions, but such features are generally missing in global emission inventories  
21 that use a bottom-up method due to the lack of up-to-date information, especially in developing countries.  
22 Here, we rely on the latest data available on emission activities, control measures, and emission factors to  
23 estimate global SO<sub>2</sub> emissions for the period 1960–2014 on a 0.1° × 0.1° spatial resolution. We design two  
24 counterfactual scenarios to isolate the contributions of emission activity growth and control measure  
25 deployment on historical SO<sub>2</sub> emission changes. We find that activity growth has been the major factor  
26 driving global SO<sub>2</sub> emission changes overall, but control measure deployment is playing an increasingly  
27 important role. With effective control measures deployed in developed countries, the predominant emission  
28 contributor has shifted from developed countries in the early 1960s (61%) to developing countries at present  
29 (83%). Developing countries show divergency in mitigation strategies and thus in SO<sub>2</sub> emission trends.  
30 Stringent controls in China are driving the recent decline in global emissions. A further reduction in SO<sub>2</sub>  
31 emissions would come from a large number of developing nations that currently lack effective SO<sub>2</sub> emission  
32 controls.

33

## 34 **Introduction**

35 Atmospheric emissions of sulfur dioxide (SO<sub>2</sub>) are of great concern because of their adverse impacts on the  
36 climate and human health<sup>1,2</sup>. Through decades of efforts to reduce SO<sub>2</sub> emissions, developed countries have  
37 made remarkable progress in mitigating SO<sub>2</sub> pollution<sup>3,4</sup>, with considerable benefits for regional to global  
38 environments<sup>4,5</sup>. Most developing countries, however, either have not adopted effective strategies or have  
39 just started to do so in recent years<sup>6</sup>. Only after 2000, for example, did China start to impose strict controls  
40 on SO<sub>2</sub> emissions<sup>7</sup>, although China's pace in emission reduction has accelerated since 2013 following the  
41 issuing of the Air Pollution Prevention and Control Action Plan<sup>8</sup>. India and many other developing countries,  
42 on the other hand, have seen continuing increases in SO<sub>2</sub> emissions over the past decades because of the  
43 growth of electricity demand and an absence of effective regulations on emission controls<sup>9</sup>. Observations  
44 from space imply that India is overtaking China as the largest SO<sub>2</sub> emitter in the world<sup>10</sup>. Such divided trends  
45 in SO<sub>2</sub> emissions by country have led to large-scale rapid changes in the spatial distribution and hotspots of  
46 global SO<sub>2</sub> emissions in recent years<sup>11,12</sup>, which have not been well captured by global emission inventories  
47 due to a lack of detailed information<sup>11</sup>.

48 Several studies and official reports have provided bottom-up estimates of SO<sub>2</sub> emissions, although there have  
49 been discrepancies in their estimated values (**Table S1**). Nevertheless, it is clear that SO<sub>2</sub> emissions mainly  
50 originate from fossil fuel consumption, especially for power generation, and industrial activities, including  
51 petroleum refining and metal smelting<sup>13,14</sup>. Fuel quality, energy mix, industrial structure, and control  
52 measures are determinants of SO<sub>2</sub> emissions in a given country<sup>15</sup>, and these determinants are ultimately  
53 associated with the socioeconomic development status of the country<sup>6</sup>. Studies have found an inverted-U  
54 shaped relationship between SO<sub>2</sub> emissions and socioeconomic development levels, indicating that SO<sub>2</sub>  
55 emissions generally follow the Environmental Kuznets Curve (EKC) as the economy grows<sup>16,17</sup>. Given recent  
56 changes in developing countries, however, differences in the SO<sub>2</sub> emission trajectories between developed  
57 and developing countries are less clear.

58 Here, we provide the up-to-date version of a global SO<sub>2</sub> emission inventory developed by the research group  
59 at Peking University (i.e., PKU-SO<sub>2</sub>, freely available at <http://inventory.pku.edu.cn/>). PKU-SO<sub>2</sub> is compiled  
60 by the bottom-up method, spans the period 1960–2014, and incorporates the latest information available for  
61 developing countries, especially for China, on activities, emission factors, and control measures (see

62 **Methods**). Additional international fuel trade information was also considered to potentially reduce the  
63 emission uncertainties induced by the spatiotemporally varying sulfur content of fossil fuels<sup>18</sup>. Based on  
64 PKU-SO<sub>2</sub>, we present in this paper a comprehensive assessment of global SO<sub>2</sub> emissions regarding the source  
65 profiles, spatial distribution, and temporal trends. We compare our assessment with previous studies and  
66 discuss the potential reasons for the discrepancies between studies. To address the factors driving the changes  
67 in SO<sub>2</sub> emissions over the study period, we introduce two counterfactual scenarios to decompose the  
68 contributions of emission activities and sulfur-control measures on historical emission trends. We then  
69 investigate the relationship between SO<sub>2</sub> emissions and socioeconomic development status by different  
70 groups of countries, using per-capita gross domestic product (*GDPcap*) as an indicator. These analyses based  
71 on the newly compiled emission data should have important policy implications for SO<sub>2</sub> pollution mitigation.

## 72 **Methods**

### 73 **Emission inventory development**

74 A bottom-up method was used to calculate country-level emissions as follows:

$$75 \quad Emis = \sum_i A_i \times EF_i$$

76 where  $A_i$  and  $EF_i$  represent the emission activity and the emission factor ( $EF$ , defined as the mass of pollutants  
77 emitted per unit of emission activity) for source  $i$ , respectively. A total of 75 emission sources were considered  
78 in the emission inventory, which included eight major sectors and six fuel types (see **Table S2** for detailed  
79 source information). The emission activity data were collected from the International Energy Agency (IEA)<sup>19</sup>  
80 for power generation, industry combustion, transportation, and residential sectors except for China, for which  
81 the residential energy consumption was taken from an updated residential energy dataset based on a recent  
82 national questionnaire survey<sup>20,21</sup>. The activity data of non-combustion industries were provided by the U.S.  
83 Geological Survey (USGS)<sup>22</sup> and the U.S. Energy Information Administration (USEIA)<sup>23</sup>. Information on  
84 the dry matter burned as agriculture waste and in wildfires and deforestation was provided by the Global Fire  
85 Emissions Database (GFED) at a monthly resolution<sup>24</sup>.

86 The  $EF$  data were collected and treated differently among emission sources. For the hard coal and oil  
87 consumed in power and industrial sectors, the  $EF$ s were directly derived from a newly compiled  $EF$  dataset

88 for SO<sub>2</sub>, detailed in a previous study<sup>18</sup>. This *EF* dataset considered the effects of both international fuel trade  
89 on country-specific sulfur contents and control measures (e.g., flue gas desulfurization or FGD, sulfur  
90 removal from petroleum refinery)<sup>18</sup>. The *EFs* of lignite were calculated by country based on the country-  
91 specific sulfur and ash contents of lignite collected from the literature (see **Table S3**) together with the FGD  
92 promotion rate according to our recent study<sup>18</sup>. For the transportation sector, linear regression was adopted  
93 to predict the spatiotemporal variation in *EFs* based on a collection of 125 *EF* measurements obtained from  
94 the literature (**Table S4**). For nonferrous metal smelters (i.e., nickel, lead, copper, and zinc) and natural gas  
95 production, the *EFs* in uncontrolled conditions were originally calculated using a mass-balance method with  
96 information obtained from the USGS and USEIA<sup>22,23</sup> and were further calibrated by the accompanying  
97 production of byproducts (e.g., sulfuric acid)<sup>22,25</sup> to obtain the final *EFs*. For other sources, which account  
98 for small fractions of the total SO<sub>2</sub> emissions, constant *EFs* from previous studies were adopted. The country-  
99 level emissions were spatially allocated into 0.1° × 0.1° grid cells by source using high-resolution energy  
100 data<sup>26</sup> as a surrogate. Emissions from the residential sector were then temporally resolved by month using  
101 grid-specific monthly profiles generated in a previous study<sup>27</sup>.

## 102 **Decomposition of driving forces on SO<sub>2</sub> emissions**

103 In addition to the real case, we developed two counterfactual scenarios to quantify the influences of two  
104 major drivers on SO<sub>2</sub> emissions, i.e., emission activity and control measure. The emission activity considered  
105 not only the changes in the magnitude of energy consumption and industrial production but also the changes  
106 in sulfur content induced by international fuel trade. We set 1960 as the starting point, and by holding control  
107 measures constant as in 1960, we quantified the net influences of emission activity on emission changes from  
108 1960 to 2014. Similarly, the influences of sulfur-control measures were quantified by keeping the sulfur-  
109 control rates during the study period constant as those in 1960 when the rates were equal or close to zero.  
110 The effects of these two drivers constituted unique trajectories of historical SO<sub>2</sub> emissions for both individual  
111 countries and the globe, which allowed us to compare the importance of these two drivers in specific regions  
112 and periods.

## 113 **Uncertainty analysis**

114 A 10000-time Monte Carlo simulation was performed to address the uncertainty in the emission estimates.  
115 The overall uncertainty stemmed from uncertainties in the *EFs*, activity data, and technology division

116 (including control measures). The distributions of the *EFs* were subject to log-normal distributions. Activity  
117 data and technology split were assumed to be uniformly distributed. Following previous studies<sup>28,29</sup>, the  
118 standard deviation was set as 50% for technology division, 5% for fuel consumption in the power and  
119 industrial sectors, 15% for transportation, 20% for indoor biomass fuels, 30% for outdoor biomass burning,  
120 and 10% for all the other sources. The uncertainties were presented here as the medians and the interquartile  
121 ranges (i.e., the interval between the 25th and 75th percentiles) of the emission values given by the Monte  
122 Carlo simulation.

## 123 **Results and discussion**

### 124 **Global emissions and source profiles in 2014**

125 With consideration of global fuel trade, the global total SO<sub>2</sub> emission from all sources (excluding volcanic  
126 emissions) was estimated as 105.4 Tg y<sup>-1</sup> (95.8~119.8 Tg y<sup>-1</sup>) in 2014, with a predominant contribution from  
127 anthropogenic sources (98%). Our inventory features not only the incorporation of the international fuel trade  
128 but also a high spatial resolution (0.1° latitude × 0.1° longitude), a long-term temporal coverage (1960-2014),  
129 and detailed source information (76 sources). **Table S5** compares the emission estimates in this study with  
130 several past global SO<sub>2</sub> emission inventories in their latest years. Generally, our estimates are in line with  
131 these previous studies. For example, the total anthropogenic emission was estimated to be 111.7 Tg y<sup>-1</sup> in  
132 2014 by CEDS<sup>13</sup> and 102.4 and 106.9 Tg y<sup>-1</sup> in 2010 by EDGAR and HTAP, respectively<sup>30,31</sup>, compared with  
133 109.2 Tg y<sup>-1</sup> (99.0~124.2 Tg y<sup>-1</sup>) and 105.4 Tg y<sup>-1</sup> (95.9~119.8 Tg y<sup>-1</sup>) in 2010 and 2014, respectively, in our  
134 study. Despite the overall similarity, substantial differences were found in the source profiles, spatial  
135 distributions, and temporal trends. Potential reasons for these differences are discussed in detail in the  
136 following sections.

137 We estimated that in 2014, 43% of the total anthropogenic emissions were from power plants (**Fig. S1**),  
138 followed by industry (35%) and international shipping (16%). Developed and developing nations showed  
139 little difference in the aggregated sectoral profiles but large differences in the fuel profiles in specific sectors.  
140 Both groups of nations showed dominant contributions of the power and industry sectors and contributions  
141 of <10% from all other sectors together (emissions from international shipping were not assigned to specific  
142 countries). Residential emissions accounted for a higher contribution in developing countries, as coal was

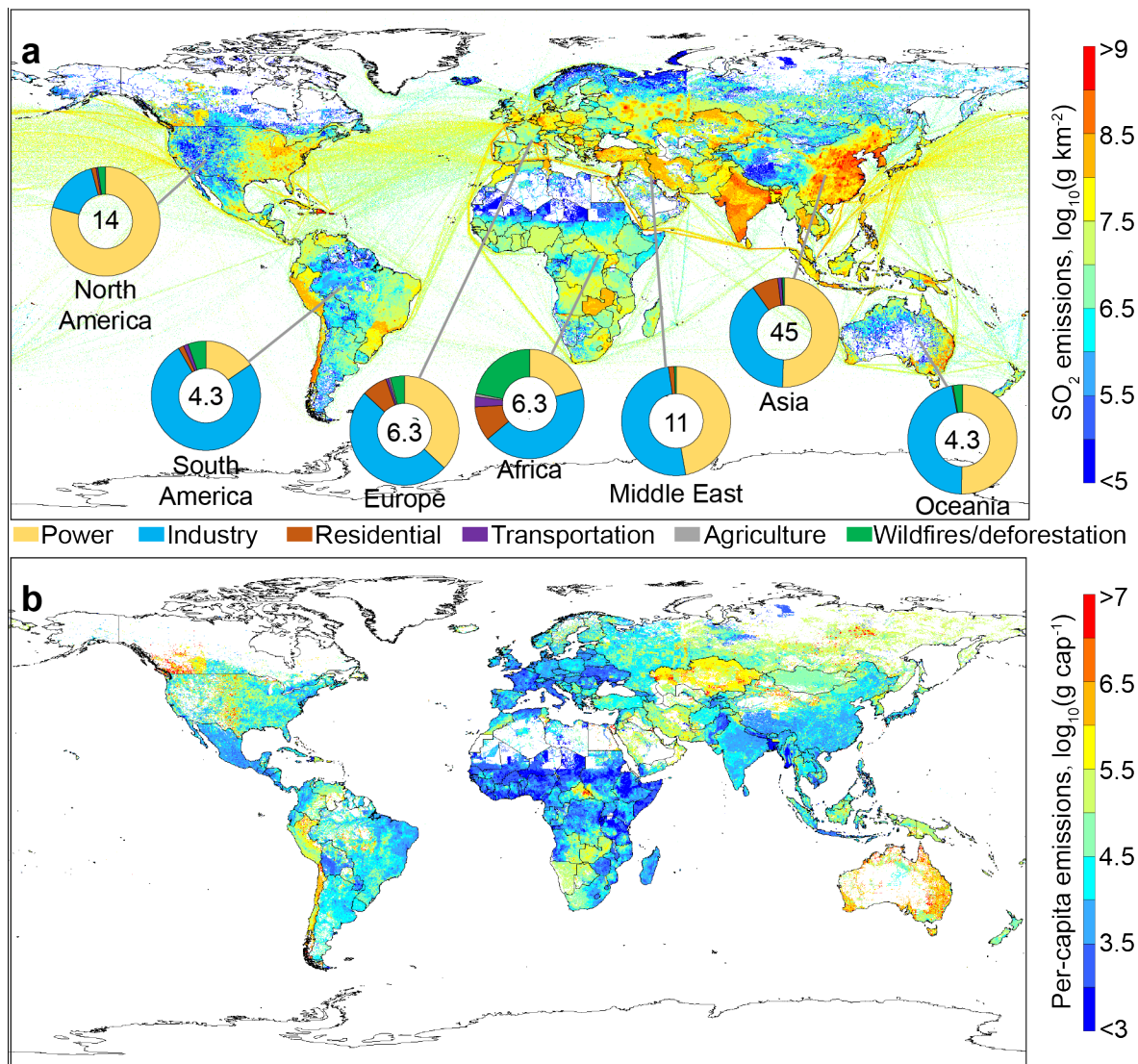
143 still a leading fuel type for household cooking and heating in many developing countries<sup>20</sup>. The combustion  
144 of hard coal and oil accounted for 78% of the global total emission. Note that these important SO<sub>2</sub> sources  
145 were strongly influenced by the international fuel trade, which was detailed in a previous study<sup>18</sup>. Developing  
146 countries showed a larger share (68%) of emissions from coal combustion than developed countries (47%).  
147 The combustion of biomass fuel, a major contributor to many other atmospheric pollutants (e.g., PM<sub>2.5</sub><sup>29</sup> and  
148 PAHs<sup>28</sup>), contributed to less than 2% of the total SO<sub>2</sub> emission. Omitting emissions from shipping, the source  
149 profiles of the SO<sub>2</sub> emissions in this study were compared with the global and regional estimates from CEDS  
150 in 2014 (the latest year of CEDS estimates available) (**Fig. S2**). Our global estimate largely agreed with that  
151 of CEDS but showed a lower contribution from developing nations. Our estimate of China's total  
152 anthropogenic SO<sub>2</sub> emission, for example, was 22.8 Tg y<sup>-1</sup> (20.1~25.8 Tg y<sup>-1</sup>) in 2014, which was close to  
153 the estimate reported by the Multi-resolution Emission Inventory for China (MEIC) (20.4 Tg y<sup>-1</sup>)<sup>8</sup> but was  
154 38% lower than the CEDS estimate (36.5 Tg y<sup>-1</sup>). The major differences with CEDS are in the industrial  
155 sector and are likely attributed to the different FGD promotion rates adopted.<sup>13,18</sup> The FGD penetration in  
156 this study was based on the reported data from China statistical yearbook on environment<sup>32</sup>, which was  
157 generally larger than previous ones<sup>7</sup>. In recent years, China has made considerable efforts to control air  
158 pollution from industrial sector<sup>33</sup>. A recent study has suggested that the industrial sector was the leading  
159 driver of the declined SO<sub>2</sub> emission in China since 2010<sup>8</sup>.

## 160 **Spatial distribution**

161 **Fig. 1** shows the spatial distribution of all-source SO<sub>2</sub> emissions (excluding those from aviation) in 2014 at  
162 a spatial resolution of 0.1°×0.1°. Among all countries, China (especially in the eastern part) and India  
163 exhibited the highest emission densities, mainly due to the intensification of coal consumption and metal  
164 smelting; these two countries accounted for 58% of coal consumption and 44% of metal smelting  
165 worldwide<sup>19</sup>. Emissions from the top 20 coal-fired power plants in China accounted for 6% of China's total  
166 SO<sub>2</sub> emissions, which is equivalent to the annual emission of Brazil. Relatively high emission densities were  
167 also found in Eastern Europe, Middle Europe, and the eastern United States due to large emissions from  
168 power generation and industrial sources. These high-emission regions usually coincided with dense  
169 population, implying that population density could play an important role in determining the spatial variation  
170 of total SO<sub>2</sub> emissions. Per-capita emissions varied substantially both within and across countries (**Fig. 1b**).



171 Densely populated places, such as eastern China, the peninsula of India, Europe, and the eastern United States,  
172 showed lower per-capita emissions ( $< 10 \text{ kg cap}^{-1}$ ) than the global average ( $12.3 \text{ kg cap}^{-1}$ ,  $11.2\text{--}14.0 \text{ kg cap}^{-1}$ )  
173 ).



**Fig. 1** Geographical distributions of (a) annual all-source SO<sub>2</sub> emissions (excluding aviation emissions) and (b) per-capita emissions in 2014. The embedded diagrams in panel (a) show the source profiles for 7 regions. The world shapefiles were obtained from Esri (ArcGIS Hub, *Countries WGS84*, June 21, 2015. <http://www.arcgis.com/home/item.html?id=30e5fe3149c34df1ba922e6f5bbf808f>).

174 According to **Fig. 1a**, the frequency distribution of gridded SO<sub>2</sub> emission densities, as shown in **Fig. S3**, was  
175 more leptokurtic compared to PM<sub>2.5</sub> and other incomplete combustion byproducts (e.g., BC, PAHs, and CO).



176 This difference reflected different geographic features of their dominant sources. The unimodal distribution  
177 of SO<sub>2</sub> emissions indicates a more concentrated spatial distribution compared with other pollutants with  
178 multimodal distribution. SO<sub>2</sub> emissions mainly came from point sources (e.g., power generation and  
179 industrial sources), while one of the major sources of incomplete combustion products was the combustion  
180 of solid fuel in rural households, of which the spatial distribution was more scattered. As a result, SO<sub>2</sub>  
181 emissions were agglomerative in space, with reduced spatial continuities, compared to PM<sub>2.5</sub> (the Moran's I  
182 is 0.50 for PM<sub>2.5</sub> and 0.32 for SO<sub>2</sub>).

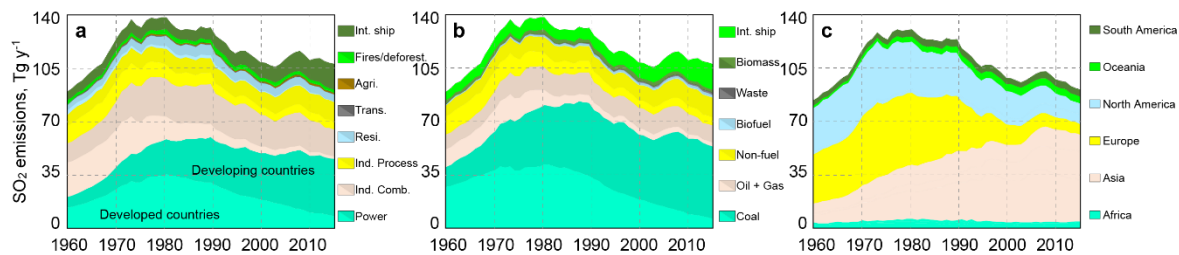
183 The inventory showed differences in both the total emissions and the source profiles among continents (**Fig.**  
184 **1a**). For example, Asia contributed 62% of the global terrestrial SO<sub>2</sub> emission; Europe and Africa only  
185 contributed 8% and 5%, respectively. In Asia, power generation and industry were the predominant  
186 contributors, accounting for 50% and 42% of Asia's SO<sub>2</sub> emission, respectively. In Europe, power generation  
187 contributed 38% of its total emission, which was less than industrial sources (50%), due primarily to a lower  
188 dependency on thermal electricity powered by fossil fuels (42%) compared to many other regions of the  
189 world (e.g., 67% in the United States, 75% in China, and 81% in India). In Africa, wildfires and deforestation  
190 contributed a large fraction of SO<sub>2</sub> emissions (21%).

191 **Fig. S4** shows the geographical distributions of sectoral emissions in 2014. By choosing the sector with the  
192 highest contribution in each grid cell, we derived the sectors most responsible for SO<sub>2</sub> emissions on a 0.1°×0.1°  
193 grid (**Fig. S4**), which provided more detailed spatial information on the source profiles and could be  
194 especially beneficial to emission control management. We found that although emissions from power and  
195 industrial sectors dominated China's total emission, the residential sector was still a major source in some  
196 parts of the northern and western China due to the lower intensity of industrial activities in these regions.<sup>34</sup>  
197 This pointed to the different emission structures between rural and urban areas, as shown in **Fig. S5**. Globally,  
198 rural areas released over 65% of the SO<sub>2</sub> emissions into the atmosphere. Likewise, the per-capita emissions  
199 for rural residents (15.1 kg cap<sup>-1</sup>, 13.7~17.2 kg cap<sup>-1</sup>) were higher than those for urban residents (9.1 kg cap<sup>-1</sup>  
200 <sup>1</sup>, 8.3~10.3 kg cap<sup>-1</sup>), especially for developed countries with less rural populations (31.4 kg cap<sup>-1</sup> on average,  
201 28.6~35.7 kg cap<sup>-1</sup>). Further investigation showed that SO<sub>2</sub> emissions in rural areas were still dominated by  
202 power generation and industrial sources, in contrast with emissions of incomplete combustion products, such  
203 as PAHs, which also showed a higher per-capita emission for rural residents compared to their urban

204 counterparts but was mainly caused by high emissions from the rural residential consumption of coal and  
205 biomass<sup>28</sup>. Relatively small differences were found between developed and developing countries in the per-  
206 capita emissions attributed to power generation. It was estimated that per-capita coal consumed in power  
207 generation was 61% higher in developed countries than in developing countries,<sup>19</sup> which offset the effects of  
208 better controls on emission reduction in developed countries.

### 209 **Temporal trends of SO<sub>2</sub> emissions**

210 **Fig. 2** shows the global trend of SO<sub>2</sub> emissions and the contributions from major sectors, fuel types, and  
211 regions. The global emission increased at an average rate of 2.4 Tg y<sup>-1</sup> (1.2~3.7 Tg y<sup>-1</sup>) during the period  
212 1960–1975, peaked at 136 Tg y<sup>-1</sup> (122~153 Tg y<sup>-1</sup>) in the late 1970s, and started to decrease in the 1980s,  
213 with an inverse trend between 2000 and 2007 as a result of increasing coal consumption in China. This  
214 emission trend was very different from the emission trends of PM<sub>2.5</sub> and many other pollutants for which the  
215 emissions largely continued rising over the last several decades<sup>29,35,36</sup>. Between 1960 and the late 1970s,  
216 global coal consumption increased by 135%<sup>19</sup>, but sulfur controls during this period were only adopted in a  
217 limited number of countries (e.g., Japan and the United Kingdom). The rising energy demand, especially in  
218 developed countries, drove the increase in global SO<sub>2</sub> emissions. Since the late 1970s, FGD technology has  
219 been scaled up in many developed countries as an efficient solution to reducing SO<sub>2</sub> emissions<sup>37</sup>. For example,  
220 the cumulative capacity of FGD units increased from 0.8 GW in 1975 to 21.1 GW in 1990 in the United  
221 States<sup>38</sup>. With policy restrictions, the expanding deployment of sulfur mitigation measures gradually offset  
222 the SO<sub>2</sub> emissions that were otherwise increasing due to the growing energy demand, and the global emission  
223 started to decline. Thanks to the catch-up of control measure deployment in developing countries in the  
224 2000s<sup>8</sup>, the declining trend continued after an 8-year reversion between 2000 and 2007. The declining trend  
225 is likely to continue for years because of the regulations and measures proposed and taken in China and other  
226 developing countries<sup>39</sup>. For example, China banned imports of coal with high ash and sulfur in 2015 and the  
227 sale of high-sulfur diesel used by tractors and ships in late-2017, both of which are expected to further  
228 contribute to the global SO<sub>2</sub> emission reduction in the coming years<sup>40</sup>.



**Fig. 2** Temporal trends of annual SO<sub>2</sub> emissions by sector (a), fuel type (b), and region (c). The light and shaded areas indicate emissions from developed and developing countries, respectively.

Emissions from international shipping, which mainly originated from oil combustion, are illustrated separately in **Fig. 2b** because they cannot be allocated to individual countries.

229 During the last 55 years, SO<sub>2</sub> emissions revealed substantial shifts in terms of emission sectors, sources, and  
 230 regions. As shown in **Fig. 2a**, emissions from industrial combustion, a primary contributor in the 1960s,  
 231 started to decline in the early 1970s due to the transition of major fuel types from coal to electricity and gas<sup>19</sup>.  
 232 On the other hand, emissions from power plants increased rapidly from 20.5 Tg y<sup>-1</sup> (18.5~22.8 Tg y<sup>-1</sup>) in  
 233 1960 to 58.4 Tg y<sup>-1</sup> (52.8~64.9 Tg y<sup>-1</sup>) in 1980 and were driven by the fast growth of electricity demand. The  
 234 contribution of power generation to global total emissions has remained constant at approximately 45% since  
 235 the mid-1980s and has slightly decreased recently (since 2005). The recent decrease was partly caused by the  
 236 growing control efforts in developing countries<sup>41</sup>. The residential sector only accounted for < 8% of the total  
 237 emissions during the entire study period. Notably, the contribution of shipping emissions increased from 6%  
 238 in 1960 to 16% in 2014. Recent studies have showed that controlling sulfur in ship oil can yield considerable  
 239 health and climate benefits<sup>42,43</sup>.

240 In terms of the fuel types (**Fig. 2b**), coal contributed the largest emissions to the global total, with a peaking  
 241 contribution of 63% in the mid-1980s. The decreasing trend since then was a joint result of FGD promotion  
 242 in power plants and the transition from coal-powered to electricity-powered industries<sup>19</sup>. Given economic  
 243 and technological gaps, the emission trends in developed and developing countries differed essentially. The  
 244 total emissions from developed countries peaked at 70.2 Tg y<sup>-1</sup> (63.4~78.4 Tg y<sup>-1</sup>) in 1973 and have  
 245 continuously declined since then, primarily driven by the promotion of control measures in power generation  
 246 (e.g., FGD) and industries (e.g., sulfur recovery technology in petroleum refining and smelting). At the same  
 247 time, emissions from developing countries gradually increased until very recently because of the increase in

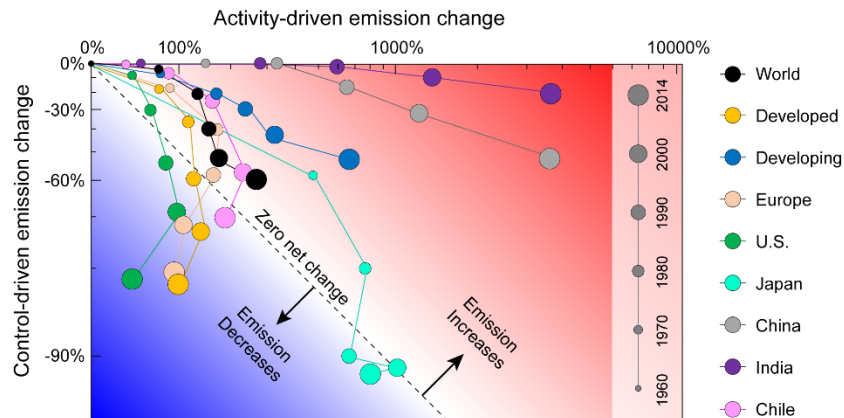
248 energy consumption and the expansion of the metal industry. For example, copper production in China  
249 increased by a factor of 10 between 1990 (560 Gg) and 2014 (6500 Gg)<sup>22</sup>. As a result, the predominant  
250 emission contributor has shifted from developed countries in the early 1960s (61%) to developing countries  
251 at present (83%), which has also led to large-scale spatial changes over time. As displayed in **Fig. 2c**, the  
252 relative contribution from Asia has increased rapidly since 1960, though it has leveled off in recent years,  
253 while the contributions from North America and Europe have decreased over time.

254 The emission trends reported by this study compared with other studies (**Fig. S6**). The comparison showed  
255 general agreement between studies except EDGAR 4.3.2 and Lamarque et al.,<sup>44</sup> of which the reported  
256 emissions were lower than the others (**Fig. S6**). We found that the estimation differences for large developing  
257 countries such as China and India can be substantial between datasets in the recent years, likely due to poor  
258 data availability in these countries. By incorporating more recent information, our estimation suggests that  
259 due to the increasing adoption of strengthened control measures, SO<sub>2</sub> emissions in China is decreasing, which  
260 represents the major reason for the recent decline in global SO<sub>2</sub> emissions. This decreasing trend agreed well  
261 with the up-to-date emissions reported by MEIC<sup>8</sup>.

262 In addition, we compared our emission estimates of China and India with top-down emission estimates  
263 derived from satellite inversion (**Fig. S7**). Large differences by factors of 2 ~ 3 were found among the top-  
264 down estimates. Since all the top-down estimates were constrained by observations of the Ozone Monitoring  
265 Instrument (OMI) aboard the Aura satellite, such large differences could be mostly attributed to the different  
266 models and retrieval algorithms used<sup>45,46</sup>. Despite the large uncertainties, generally decreasing trends were  
267 found in China since 2006, which was in consistent with our results (**Fig. S7a**). While for India, those top-  
268 down estimates were considerably lower than the estimates in this and most other bottom-up inventories,  
269 although all showed continuously increasing trends (**Fig. S7b**). The large disparity could be due to the  
270 uncertainties from various sources in the bottom-up estimation, such as EFs and energy consumption<sup>28,29</sup> but  
271 could also be caused by the quality of the satellite observations, which can be contaminated by a variety of  
272 factors such as cloud cover and land property, and the uncertainties in the retrieval and inversion algorithms  
273 in the top-down estimation could also play a role<sup>46</sup>. Future studies were called for to narrow down the  
274 uncertainties in both emission estimation approaches and reduce the disparity between the two estimates.

275 **Decomposition analysis of global and regional SO<sub>2</sub> emission trends**

276 In addition to the emission estimates using the real-world data (called the “real case”), we designed two  
277 counterfactual scenarios to decompose the historical SO<sub>2</sub> emission trends into the changes in 1) emission  
278 activities (i.e., energy consumption and structure, industrial production, trade-induced changes in sulfur  
279 contents or SCs) and 2) end-of-pipe control measures (i.e., FGD in power and industrial boilers, control  
280 measures in the road transportation, sulfur removal in petroleum refinery, metal smelter, and natural gas  
281 production). Based on the baseline emissions in 1960, the relative emission changes induced by emission  
282 activities were quantified by holding the control measures constant and changing the emission activities over  
283 time as in the real case; the relative emission changes induced by control measure deployment were calculated  
284 by holding the emission activities at their 1960 levels and changing the control measures over time as in the  
285 real case. **Fig. 3** illustrates the SO<sub>2</sub> emission trajectories along the two dimensions (i.e., emission activities  
286 and control measures) on global and regional scales from 1960 to 2014. Emission activity growth and control  
287 measure deployment were competing factors affecting the historical SO<sub>2</sub> emission trend. From a global  
288 perspective, emission activity growth played a more important role (**Fig. 3**), but substantial differences in the  
289 trajectories were found among regions/countries. Before 1970, activity was the leading force for both  
290 developed and developing countries, and the emission trajectories extended mostly along the direction of the  
291 activity’s dimension (the horizontal axis in **Fig. 3**). After 1970, however, the trajectories started to divide:  
292 developed countries redirected their trajectories along the control measure’s dimension towards the low end;  
293 developing countries such as India and China continued to move along the activity’s dimension and did not  
294 show significant impacts from control measure deployment until recently (**Fig. 3**). The redirection in  
295 developed countries was caused by a combination of stringent emission controls, suppressed fossil fuel use,  
296 and the overseas outsourcing of raw industrial materials. For example, coal use (in energy count) and  
297 nonferrous metal production in the United States decreased by 29% and 18% from 2000 to 2014,  
298 respectively<sup>19,22</sup>. For most developing countries, by contrast, activity remained the predominant driver to  
299 decadal SO<sub>2</sub> emission changes with slight differences among countries due to different sulfur control efforts.  
300 For example, China exhibited a decrease of 52% in emissions due to control measure deployment, compared  
301 with 20% in India (**Fig. 3**). Note that both countries showed a similar emission increase induced by activity  
302 changes (**Fig. 3**). Even in some developing countries (e.g., Chile in **Fig. 3**), we found rapid transition in the  
303 emission trajectories from being activity-dominant (the red area in **Fig. 3**) to control-dominant (the blue area),  
304 as in developed countries, leading to a fast decrease in SO<sub>2</sub> emissions.



**Fig. 3** Trajectories of SO<sub>2</sub> emissions driven by emission activity and control measures from 1960 to 2014. Each trajectory shows the emission ratios of a specific year relative to the base year (1960) driven by activity (*x*-axis) and control measures (*y*-axis). The axes are on a logarithmic scale. The dash line indicates equal forces from the two drivers. The red area indicates stronger impacts of activity growth compared to control measure deployment on emissions; the blue area indicates the opposite.

### 305 Environmental Kuznets Curves of SO<sub>2</sub> emissions

306 To better elucidate the varied SO<sub>2</sub> emission trajectories among countries, we derived the country-level per-  
 307 capita SO<sub>2</sub> emissions and investigated the relationship between per-capita emissions and socioeconomic  
 308 development. Although developing countries dominated global SO<sub>2</sub> emissions (84%), the average per-capita  
 309 emissions in developed countries (13.7 kg cap<sup>-1</sup>, 12.4~15.3 kg cap<sup>-1</sup>) were 11% higher than those in  
 310 developing countries (12.1 kg cap<sup>-1</sup>, 10.8~13.8 kg cap<sup>-1</sup>), which was mainly attributed to the much higher  
 311 per-capita energy consumption in developed countries—the average per-capita consumptions of coal and  
 312 electricity in developed countries were 66% and 350% higher than those in developing countries,  
 313 respectively<sup>19</sup>. Stringent controls offset the impact of this high energy consumption on per-capita emissions  
 314 in developed countries<sup>47</sup>. It was estimated that in 2014, the average removal rate of SO<sub>2</sub> emissions from  
 315 developing countries only reached the level in developed countries of more than two decades ago. The global  
 316 SO<sub>2</sub> emissions would be reduced by 41% if emission control measures in developing countries were as  
 317 advanced as those in developed countries.

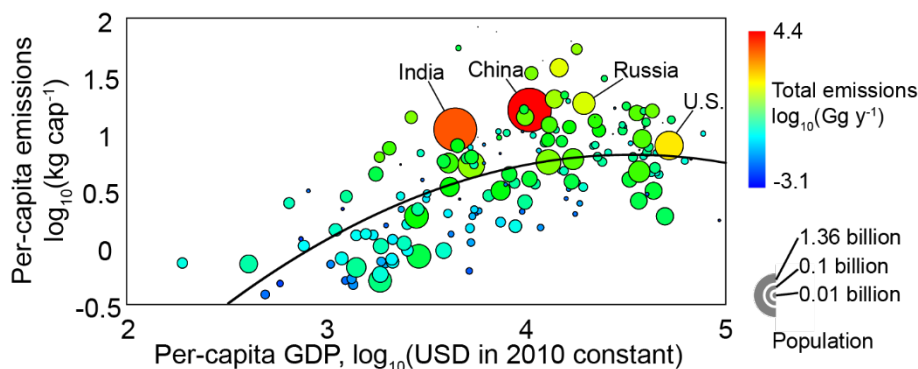
318 Fuel mix and control strength evolve with socioeconomic growth, leading to disparities in SO<sub>2</sub> emissions

319 among countries. **Fig. 4** shows the relationships between log-transformed per-capita SO<sub>2</sub> emissions and  
320 socioeconomic levels as identified by *GDPcap* (in 2010 constant USD) in 2014<sup>48</sup>. Excluding those  
321 countries/regions with populations of less than 0.1 million, we found an inverted U-shaped relationship, i.e.,  
322 an EKC<sup>16</sup>, of SO<sub>2</sub> emissions with *GDPcap*. By using an empirical quadratic regression, the turning point of  
323 SO<sub>2</sub> emissions was identified at *GDPcap* = 30000 USD when the per-capita emission reached its highest  
324 level. It should be noted that the EKC and the turning point for SO<sub>2</sub> emissions were based on a global  
325 perspective, which could vary significantly in individual countries and regions. Directly applying the  
326 relationship to certain countries could induce strong biases. For example, China's *GDPcap* reached 30,000  
327 USD in 2014, but its per-capita SO<sub>2</sub> emission started to decline in 2006, eight years earlier than the turning  
328 point indicated by the global EKC. This result is consistent with previous studies that suggested that the EKC  
329 patterns of SO<sub>2</sub> emissions differ by country or even city<sup>49</sup>. Some studies have further used the EKC  
330 trajectories to develop global emission inventories<sup>50,51</sup>. With a specific focus on SO<sub>2</sub> emissions, we  
331 summarized and compared the estimated EKC patterns of different studies (**Table S6**). The global turning  
332 point of our study (*GDPcap* = 30,000 USD) was close to the estimate by a recent study (24,300 USD)<sup>52</sup>, but  
333 both studies suggested much earlier turning points than past studies that relied on data prior to 2000 (in those  
334 studies, the estimated turning points were all at *GDPcap* > 150,000 USD). Note that before 2000, sulfur  
335 control measures had not been extensively applied in developing countries. This disparity between recent and  
336 past studies indicated a potential shift in the global EKC pattern over time. Further investigation revealed  
337 that different studies showed consistent turning points for OECD countries (**Table S6**), which fell between  
338 11,000–16,500 USD but were divided for non-OECD countries (the OECD/non-OECD classification was  
339 used here to match the classification used in previous studies). For example, our study showed the turning  
340 point of non-OECD countries at *GDPcap* = 79,000 USD, which was approximately 50% lower than a  
341 previous estimate using 1960–1990 emission data<sup>16</sup>; another estimate using 1984–2000 data did not show a  
342 turning point at all<sup>53</sup>, indicating that it may occur at a very high *GDPcap* level. Such a difference in the EKCs  
343 of non-OECD countries was also observed when using our emission data along. By comparing the EKCs  
344 derived from our data in different emission years, we found that the *GDPcap* level (instead of the per-capita  
345 SO<sub>2</sub> emission) associated with the turning points of the non-OECD countries decreased from 490,000 USD  
346 to 79,000 USD between 1970 and 2014. This decrease ultimately drove the global turning point from 150,000  
347 USD to 30,000 USD during the same period. Using the long-term data, we further examined the changes in



348 per-capita emissions with  $GDP_{cap}$  by country. **Fig. S8** illustrated the per-capita trends in eight representative  
349 countries, showing that the turning points for the developing countries appeared much earlier than developed  
350 countries in terms of  $GDP_{cap}$ .

351 The mechanism underlying the temporal shift in the global turning point and the spatial difference across  
352 countries can be likely explained by a “late mover” advantage associated with the transmission of emission  
353 control strategies and technologies between countries/regions<sup>35,54</sup>. For example, decadal experience and  
354 knowledge gained from air pollution mitigation in Europe and North America has helped China to adopt  
355 efficient strategies for sulfur emission controls in recent years<sup>55</sup>, noting that the  $GDP_{cap}$  triggering the turning  
356 point in China (2,000 USD) was much lower than the  $GDP_{cap}$  triggering the turning points in these developed  
357 nations (e.g., 21,700 USD for USA and 18,054 USD for UK) (**Fig. S8**). The turning point in India were found  
358 even earlier (1,550 USD) (**Fig. S8**). Meanwhile, China is outsourcing its experience to less developed  
359 countries, for example, in Southeast Asia.<sup>56</sup> To evaluate the impacts of the “late mover” advantage on global  
360  $SO_2$  emissions, we designed a ‘no late mover’ scenario assuming that developing countries in 2014 strictly  
361 adopted the same sulfur controls as those historically taken by developed countries at the same  $GDP_{cap}$ . It was  
362 found that the total global  $SO_2$  emission would have been 20% higher in the “no late mover” scenario than  
363 was in the real case, highlighting the importance of the “late mover” effect on global  $SO_2$  emission reduction.  
364 Nationally, the ‘late-mover’ advantage in developing countries reduced the per-capita emissions by 42% on  
365 average but with large variation across countries (2% ~ 96%). In China, especially, the ‘late-mover’ advantage  
366 was estimated to reduce  $SO_2$  emissions by 76% in 2014. This analysis may facilitate a better understanding  
367 and provide an easier forecast of future  $SO_2$  emission trends in countries across different socioeconomic  
368 development levels.



**Fig. 4** Dependency of per-capita SO<sub>2</sub> emissions on per-capita GDP for all countries in 2014. The sizes and colors of the circles indicate the country population and total SO<sub>2</sub> emissions, respectively.

369 Based on a newly compiled SO<sub>2</sub> emission inventory, this study elucidated the spatiotemporal variations of  
 370 global SO<sub>2</sub> emissions and the underlying driving forces. According to our estimation, the recent trend of SO<sub>2</sub>  
 371 emissions was subject to a remarkable decrease as a result of both enduring control efforts by developed  
 372 countries and an increasing awareness of sulfur control in some developing countries. Future trends in global  
 373 SO<sub>2</sub> emissions will depend on how developing countries act in response to their rapid economic growth and  
 374 urgent demands for clean air. The world is globalizing, and so is the air pollution issue<sup>57</sup>. Our analysis showed  
 375 that developing countries benefited from globalization through the transmission of emission control strategies  
 376 and technologies from developed countries, suggesting that regional and global cooperation is an important  
 377 way to reduce SO<sub>2</sub> emissions in the genuinely global era.

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### 381 **Supporting Information**

382 Source profiles of SO<sub>2</sub> emissions (Fig. S1), comparison between CEDS and this study (Fig. S2), the statistical  
 383 distribution of gridded SO<sub>2</sub> emissions in comparison with other pollutant emissions (Fig. S3), sector-resolved  
 384 geographical distribution of SO<sub>2</sub> emissions (Fig. S4), rural-urban emission differences (Fig. S5), inter-  
 385 comparison of global and regional SO<sub>2</sub> emission inventories (Fig. S6). Comparisons of emissions in China

386 and India between top-down estimates and this study (Fig. S7). Historical EKC for selected countries (Fig.  
387 S8).

388 Differences of SO<sub>2</sub> emissions reported by multi-sources (Table S1), summary of EFs used in this study (Table  
389 S2), SCs and ACs of lignite (Table S3), regression model of EFs for motor vehicles (Table S4), global SO<sub>2</sub>  
390 emission comparisons (Table S5), summary of turning point of EKC for SO<sub>2</sub> emissions (Table S6).

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